

ASSESSING PREVALENCE AND CHARACTERIZING *SALMONELLA*
POPULATIONS WITHIN ENVIRONMENTAL RESERVOIRS

by

JARED CHRISTOPHER SMITH

(Under the Direction of Nikki W. Shariat)

ABSTRACT

Salmonella enterica is a leading cause of bacterial foodborne illness. Ubiquitous in nature, this pathogen is well characterized in food animals; however, the prevalence, diversity, and dynamics of *Salmonella* in the environment are understudied. This introduces concern of food safety threat spillover from environmental reservoirs such as wild birds and surface water. Recent outbreaks linked to wild birds and surface water demonstrate the need to identify sources of *Salmonella* in our food systems to reduce the incidence and severity of illnesses from environmental contamination. To better understand routes of *Salmonella* transmission at the environment agriculture interface, three studies were completed that aimed to assess prevalence and characterize *Salmonella* populations within environmental reservoirs. First, wild bird feces were collected from produce fields in the southeastern United States, identifying a low prevalence of *Salmonella* and a strong association of viable pathogen with fresh feces. Results of this study suggest, for the first time in this region, a limited food safety risk attributed to wild birds compared to studies on the West Coast finding a more significant risk. This led to identifying more likely routes of produce contamination, including surface water.

Second, a two-year longitudinal surveillance study of four surface creeks determined high *Salmonella* prevalence and serovar complexity, where modeling showed a positive correlation between both prevalence complexity to weather variables such as precipitation and humidity. The incidence of antimicrobial resistance and clinically relevant serovars within complex populations supports the need for consistent monitoring of surface water sources. Finally, a comparative study between two rivers in Pichincha, Ecuador found higher prevalence and complexity in the river within an urban environment compared to a river embedded in animal agriculture. This data highlights potential human health risks associated with contaminated surface water and suggests a need for continued surveillance. Overall, these studies have significantly improved our understanding of *Salmonella* ecology in various environmental reservoirs. Considerable knowledge gaps remain, such as more fully elucidating the complex influence of weather and physiochemical variables on *Salmonella* presence in surface water. Additionally, many wildlife reservoirs remain understudied, limiting our understanding of *Salmonella* ecology and transmission outside humans and food systems.

INDEX WORDS: *Salmonella enterica*, food safety, deep serotyping, environmental surveillance, wild birds, surface water, whole genome sequencing, antimicrobial resistance

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JARED CHRISTOPHER SMITH

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JARED CHRISTOPHER SMITH

Major Professor:	Nikki W. Shariat
Committee:	Xiangyu Deng
	Elizabeth A. Ottesen
	William E. Snyder
	Christine M. Szymanski

Electronic Version Approved:

Ron Walcott
Vice Provost for Graduate Education and Dean of the Graduate School
The University of Georgia
May 2025

DEDICATION

To my mom and dad, Paula and Chris: Your continuous love and support are an immense part of my success. I love you both so much and hope to make you proud every day. Thank you for all that you have done for me!

To my sisters, Lauren and Kadi: I have always seen you both as role models for your commitment to achieving both personal and career goals as well as always pushing me to do the same, no matter how hard things get. I will never be able to thank you enough for the encouragement you both have given me, I love you two!

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CHAPTER 1

LITERATURE REVIEW

1.1 *Salmonella*

Salmonella was first discovered by Dr. Theobald Smith in 1885 during an investigation of pig intestines and subsequently named after his collaborator, Dr. Daniel Elmer Salmon (Salmon & Smith, 1891). This gram-negative, rod-shaped, flagellated, facultative anaerobic bacterium is part of the *Enterobacteriaceae* family. *Salmonella* is the causative agent of salmonellosis, gastroenteritis caused by *Salmonella*, which is generally self-limiting inflammation of the gastrointestinal system (Centers for Disease Control and Prevention, 2023). This can be identified by diarrhea, abdominal cramps, fever, headache, and nausea which can often be cleared naturally within 7-10 days or with the assistance of antibiotics when necessary. Salmonellosis impacts an estimated 1.35 million people in the US each year, with approximately 23,000 hospitalizations and 420 deaths annually (Centers for Disease Control and Prevention, 2023).

***Salmonella* serovars**

Salmonella enterica is highly diverse, including 2,659 serovars (P. A. D. Grimont & Weill, 2007; Issenhuth-Jeanjean et al., 2014). These serovars are designated in the White-Kauffman-Le minor scheme and are phenotypically differentiated by the detection of surface antigens, including the somatic (O), flagellar (H), and capsular (Vi) antigens (Brenner et al., 2000; P. A. D. Grimont & Weill, 2007). Variability within the structure of these surface antigens is the basis for *Salmonella* serotyping. The somatic antigen is the

outermost repeating carbohydrate in the lipopolysaccharide (LPS) layer on the *Salmonella* surface. The LPS layer serves as a barrier to low pH and protects *Salmonella* from the innate immune response, where releasing the lipid A component can trigger a strong immune response (Ernst et al., 2001). Currently, there are 47 variations of the O-antigen that have been identified. All serovars with common O-antigen identity belong to the same serogroup (Grimont and Weill, 2007). *Salmonella* also possesses a flagellum containing an H antigen, of which there are 119 variations. The flagella contains one of two antigens, referred to as H1 or H2, identified by the expression of two genes encoded by the *fliC* and *fliB* genes, respectively (Andrewes, 1922; Macnab, 1996). Because flagella are often the target of the host immune response, *Salmonella* can change the translation of the flagellar antigen between the two proteins (H1/H2) in a process known as phase variation (Silverman et al., 1979). Some serovar variants cannot change phases between the H1 and H2 antigens due to mutations or deletions of either the *fliC* or *fliB* gene. These variants are designated as monophasic serovars such as serovar I 4,5,[12],i:-, a monophasic variant of serovar Typhimurium (Kauffmann, 1964). Additional serovars that lack an H2 antigen include Enteritidis, Typhi, Dublin, and Senftenberg. Notably, serovar Gallinarum does not contain the gene for either flagellar antigen, resulting in nonflagellated and nonmotile cells (P. A. D. Grimont & Weill, 2007). Finally, the Vi antigen, a linear polymer consisting of O-acetylated α -1,4-linked N-acetylgalactosaminuronic acid, is found on the surface of the *Salmonella* capsule (H. Zhang et al., 2006). This antigen is only found in serovars Typhi, Paratyphi C, and Dublin and is associated with enteric disease (Santander et al., 2008).

Salmonella enterica subspecies *enterica* can be divided into two categories that separate typhoidal from non-typhoidal salmonellae (Crump et al., 2004). Serovars Typhi, Sendai, and Paratyphi (A, B, and C), are considered typhoidal serovars and the causative agents of typhoid (enteric) fever and paratyphoid fever (Gal-Mor et al., 2014). Serovars that do not often result in enteric disease are considered non-typhoidal *Salmonella* (NTS). These serovars often induce self-limited gastrointestinal inflammation known as salmonellosis, and in some cases, invasive extra-intestinal bacteremia, that can be more life-threatening (Mandal & Brennand, 1988). At 1582 serovars, these make up the overwhelming majority of serovar diversity (Issenhuth-Jeanjean et al., 2014). Alternatively, nearly 5% of illnesses in sub-Saharan Africa and southeast Asia result in enteric fever-like illness from NTS serovars, such as Enteritidis, Typhimurium, Dublin, and Cholerasuis, where these strains are categorized as invasive NTS (iNTS) (Mandal & Brennand, 1988). The separation of typhoidal and non-typhoidal salmonellae is based on human illness; however, some serovars have unique pathogenicity within specific hosts, such as the presence of typhoid fever-like symptoms in mice colonized with serovar Typhimurium (Santos et al., 2001), serovar Gallinarum causing fowl typhoid in poultry (Pascopella et al., 1995), serovar Abortusovis causing sporadic abortions in sheep (Belloy et al., 2009), and serovar Cholerasuis causing pneumonia and septicemia in swine (Gray et al., 1996). The differences in morbidity between these serovars are an example of the phenotypic and genotypic plasticity between serovars of *Salmonella enterica* subspecies *enterica*.

While the antigenic profile is used to determine the serovar identity, other genomic differences including virulence factors (Andino & Hanning, 2015),

antimicrobial resistances (D. H. Shah et al., 2017), stress response (Gorski & Noriega, 2023), and host restriction adaptations can drastically impact the clinical relevance of a serovar (Cheng et al., 2019). *Salmonella enterica* subspecies *enterica* contains 1,586 serovars, where 80 serovars accounted for 99.9% of infections in the United States between 2018 and 2024 (CDC, 2024a; Hendriksen et al., 2011). Further, more than 75% of those cases were attributed to only 13 serovars (Figure 1.1). These differences are attributed to the requirements for a clinical case to manifest: for a serovar to colonize and cause illness in a human, it must have the virulence factors available to attach, invade, and multiply within human epithelial cells and macrophages (Ernst et al., 1990; Knodler et al., 2010; Monack et al., 1996). Additionally, serovars have different dose responses, meaning that the number of cells required to cause illness is variable between serovars (M. Kim et al., 2024; Teunis et al., 2010). For serovars attributed to non-host contamination, such as produce or water, they must be able to overcome the environmental stressors outside of the host, such as temperature (J. Shah et al., 2013; Sirsat et al., 2011), pH (Spector & Kenyon, 2012), nutrient limitation (Stenström et al., 1989), and radiation (Ormsby et al., 2024). Finally, serovars attributed to food animals must have the adaptations to survive transmission. This includes colonizing both animal and human hosts, overcoming environmental stress between hosts, and facilitating transmission from one host to another (Figure 1.2) (Siceloff et al., 2022).

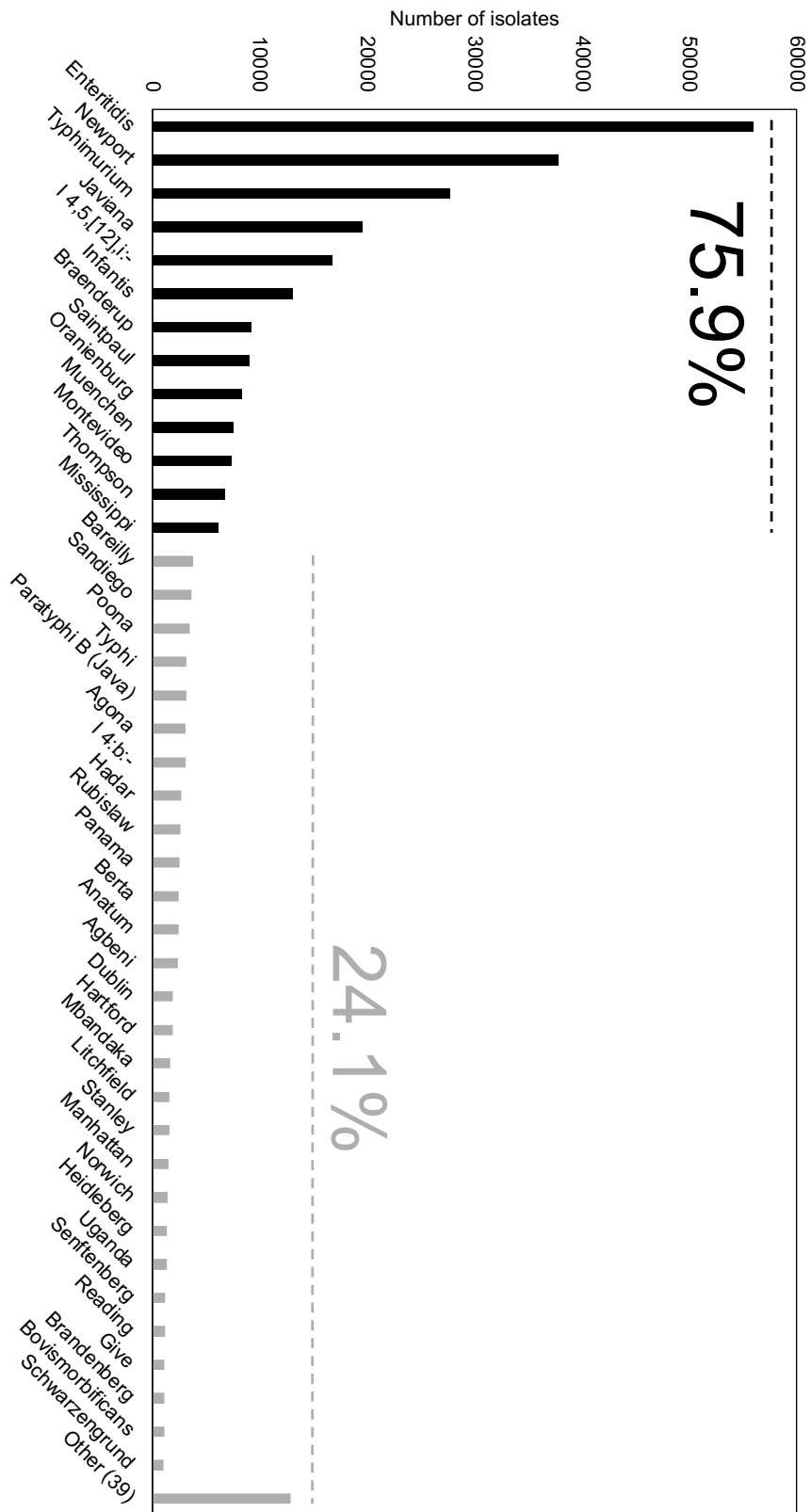


Figure 1.1. Human *Salmonella* isolates from the United States between January 2018 and November 2024. Total isolates from January 2018 through November 2024 are shown next to their respective serovar. Serovars that account for 75.9% of total isolates are shown in black. Serovars with less than 1000 isolates during this time are grouped as “other” where this group contains data from 39 serovars. Data from CDC Beam Dashboard (CDC, 2024a).

Some serovars have adaptations that better promote environmental survival

Surviving cells will not always be introduced to humans

Transmitted cells must survive host conditions and evade immune response

If all previous steps occur, colonization can proceed

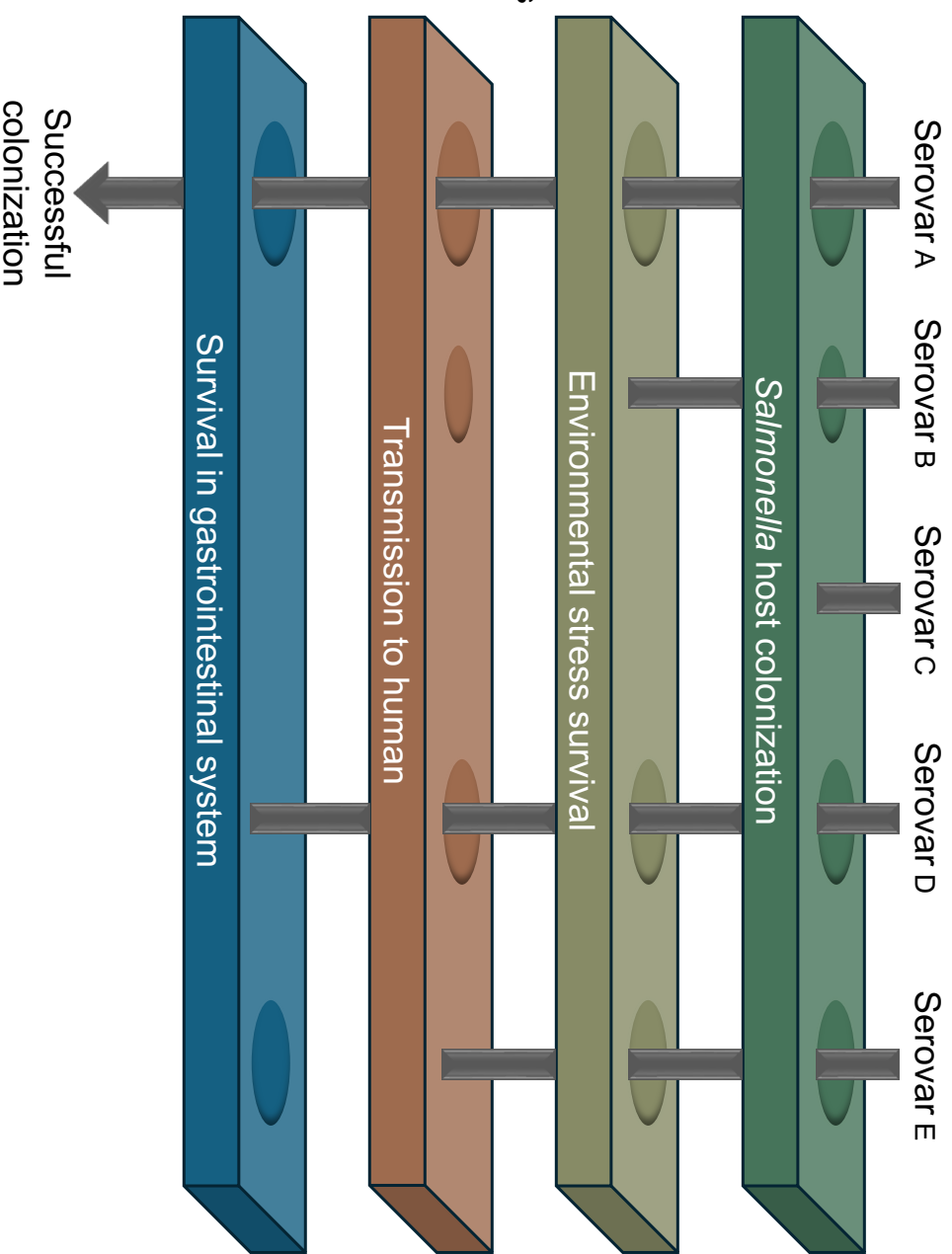


Figure 1.2. Barriers to successful human colonization by *Salmonella*. Rectangles represent obstacles that must be overcome for human colonization by *Salmonella*, where holes in a rectangle represent adaptations or opportunities for a potential serovar to overcome this obstacle. A gray arrow represents the passage of a serovar through the barriers, stopping if adaptations or opportunities are not present to move forward.

***Salmonella* evolution and phylogeny**

The genus *Salmonella* consists of two species, *Salmonella enterica* and *Salmonella bongori* (P. A. D. Grimont & Weill, 2007; Tindall et al., 2005). These species both possess the *Salmonella* pathogenicity island 1 (SPI-1), which differentiates them from other *Enterobacteriaceae* and also promotes host invasion via a type-three secretion system (T3SS) and effector proteins (Lou et al., 2019). *S. enterica* is further differentiated by the addition of SPI-2 containing a secondary T3SS and effector proteins allowing for invasion, survival, and replication within macrophages (Figueira & Holden, 2012). The lack of SPI-2 and the evolution of non-mammalian targeted effector proteins in *Salmonella bongori* restricts hosts to cold-blooded animals, such as reptiles (Bäumler et al., 1998; Hensel, 2000). Within *Salmonella enterica*, there are currently six subspecies, *enterica* (I), *salamae* (II), *arizonae* (IIIa), *diarizonae* (IIIb), *houtenae* (IV), and *indica* (VI) (P. A. D. Grimont & Weill, 2007). While these six subspecies can be differentiated using biochemical tests for phenotypic and metabolic identifiers, they typically occupy separate ecological niches (Ewing & Edwards, 1986; P. A. D. Grimont & Weill, 2007). *Salmonella enterica* subspecies *enterica*, unlike other subspecies, has evolved rapidly to target colonization of warm-blooded animals and makes up more than 99% of human clinical infections (Hadjinicolaou et al., 2009). In rare cases, non-subsp. *enterica* isolates are recovered from humans; however, these are often attributed to young children with weakened immune systems or co-infections with a subsp. *enterica* serovar (Editorial team et al., 2008; Mughini-Gras et al., 2016).

Host specialization

Salmonella serovars can exhibit varying host preferences, ranging from generalist to restricted. Host restriction is the least common, where the serovar can only colonize a single species or a restricted set of organisms, and is unable to colonize most other hosts. *Salmonella* serovars Gallinarum, Typhi, and Abortusovis are examples of this designation (Uzzau et al., 2000). Poultry and aquatic birds are the sole hosts of serovars and Gallinarum and its biovar Pullorum, causing high mortality due to fowl typhoid and Pullorum disease, respectively. These serovars were common in the United States and led to high economic loss in the early 1900s before the National Poultry Improvement Plan (NPIP) put forth mitigation strategies that successfully eradicated these salmonellae from commercial flocks (Andino & Hanning, 2015; Bullis, 1977; Kabir, 2010). Serovar Typhi and the other typhoidal serovars are host-restricted to humans and higher primates (Pascopella et al., 1995). Finally, serovar Abortusovis is restricted to sheep, where it causes sporadic abortions and stillbirths in pregnant ewes (Uzzau et al., 2000).

Host-adapted serovars may specialize in a single host but are capable of colonizing one or more other hosts. Examples of this include serovars Dublin and Cholerasuis, where the primary hosts are cattle, and swine, respectively (Uzzau et al., 2000); however, illness in humans is possible following consumption of or direct contact with infected animals. Serovars Dublin and Cholerasuis are less common in humans, but illness from these serovars can be exacerbated by antimicrobial resistance (AMR) and rare systemic infections (Chiu et al., 2002). Finally, host generalists are ubiquitous in their ability to colonize a wide range of hosts, such as serovars Typhimurium and Infantis. Serovar Typhimurium is one of the leading causes of human salmonellosis in the United States and can be found in almost all warm-blooded animals (Baumler & Fang,

2013). Serovar Infantis has become more prominent over the past two decades, as shown by its prominence in poultry, cattle, swine, and turkey environments and the presence of the pESI megaplasmid containing antimicrobial and heavy metal tolerance (Gal-Mor et al., 2010; McMillan et al., 2020). Infantis now commonly identified as one of the top ten serovars identified in all regularly monitored meat and poultry products in the United States (USDA-FSIS, 2024a).

Impact of *Salmonella* on human health

Foodborne illness is observed an estimated 179 million people in the United States each year. The cases that can be directly linked to the consumption of contaminated food are estimated at 48 million (S. Hoffmann et al., 2012). Of these cases, *Salmonella* is the leading contributor to bacterial foodborne illness, hospitalizations, and deaths (S. Hoffmann et al., 2012). The economic burden of *Salmonella* illness is estimated at over \$4.1 billion, accounting for the highest proportion of the \$17.5 billion burden of all 15 leading foodborne pathogens (USDA-ERS, 2023). These illnesses are often caused by the consumption of contaminated food and water and are commonly identified in raw meat and uncooked produce, but can also be found in eggs, dairy, nuts, flour, and other commodities (Bintsis, 2017). To better predict the risk associated with different food groups, data is shared between federal agencies, including the Centers for Disease Control and Prevention (CDC), the United States Food and Drug Administration (FDA), and the United States Department of Agriculture (USDA) to create the Interagency Food Safety Analytics Collaboration (IFSAC). Each year, a report is released that estimates the source attribution of foodborne illness outbreaks based on data collected between 1998 and two years before the reporting year. These estimates are

developed from a model that relies more heavily on more recent outbreaks (within the previous five years) (Batz et al., 2021). Estimates are given for *Salmonella enterica*, *E. coli* O157, *Listeria monocytogenes*, and *Campylobacter jejuni* (until 2020). Annual data suggests a trend toward increasing foodborne illness attribution to meat and poultry, surpassing produce in total outbreak attribution in 2022 (Figure 1.3) (IFSAC, 2024). Other significant bacterial causes of foodborne illness include *E. coli*, *Listeria monocytogenes*, and *Campylobacter jejuni*. The sources of foodborne illness are more variable in *Salmonella* compared to other bacterial foodborne pathogens such as that of *E. coli* where vegetable row crops (62.2%) and beef (20.9%) alone account for 85.1% of estimated illnesses (IFSAC, 2024). This highlights the complex and widespread nature of *Salmonella* and its ability to cause illnesses from a variety of reservoirs.

Of the estimated 1.35 million salmonellosis cases each year, outbreaks account for only 10% of these (Scallan et al., 2011). Sporadic infections are more common, often due to the consumption of improperly cooked or cleaned food and interactions with contaminated animals, meat, or fecal material (Tack et al., 2019). To limit bias, the Foodborne Disease Active Surveillance Network (FoodNet) includes collected clinical data from ten states (Oregon, Colorado, New Mexico, Minnesota, Tennessee, Georgia, Delaware, Connecticut, and select counties in California and New York), outbreak data from the FDA, and routine surveillance of meat, poultry, and egg products by the USDA-Food Safety Inspection Service (USDA-FSIS). FoodNet was formed in 1995 and currently includes testing for eight foodborne pathogens including *Campylobacter*, *Listeria*, *Salmonella*, STEC, *Shigella*, *Vibrio*, *Yersinia*, and *Cyclospora*. Between 2000 and 2022, 166,103 *Salmonella* infections have been identified on FoodNet. Serovars

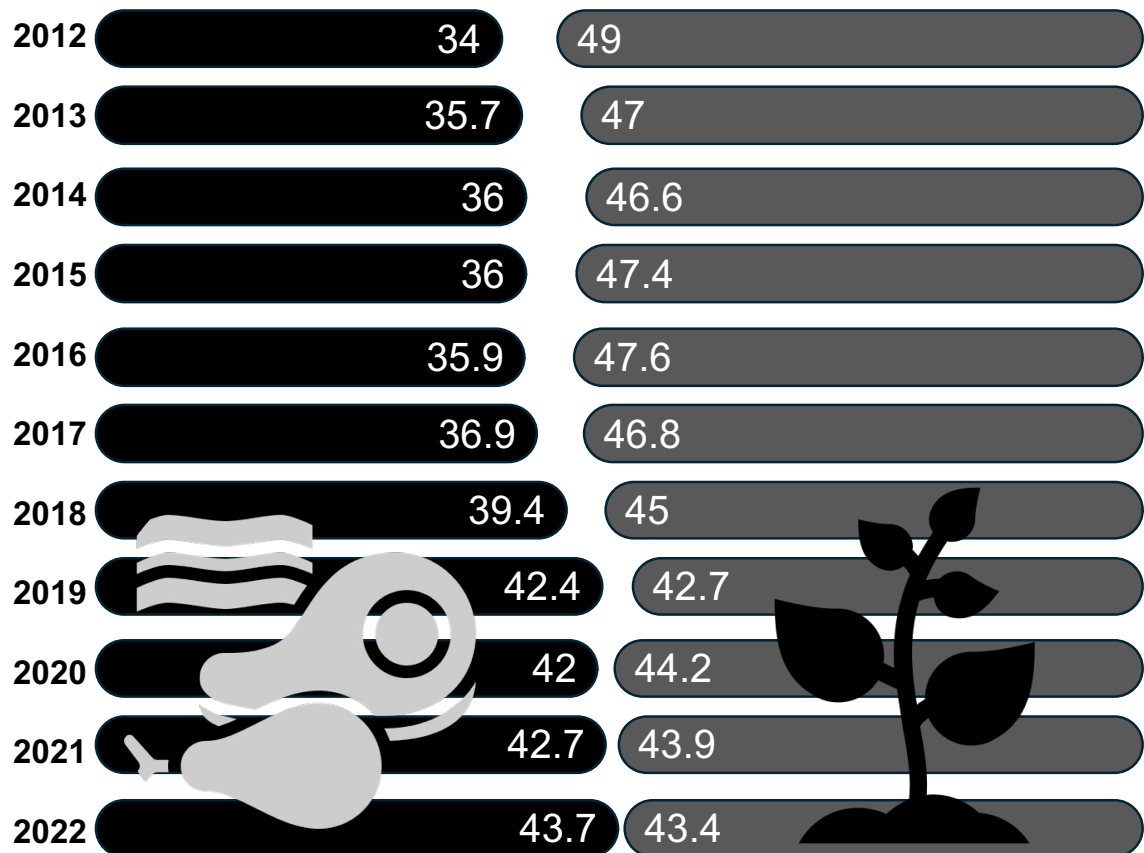


Figure 1.3. Estimated *Salmonella* foodborne illness attribution of meat and poultry vs. produce from 2012-2022. Meat and poultry (black) include chicken, pork, beef, and turkey. Produce (gray) contains fruits, seeded vegetables, vegetable row crops, sprouts, grains and beans, and “other produce” categories from each annual IFSAC report. Percentages are estimated by combining outbreak data from 1998 to each year, weighing the five most recent years more heavily than prior years. Categories not shown include dairy, eggs, seafood, game, oils-sugars, and “other meat/poultry”.

Enteritidis, Typhimurium, and Newport were the top three serovars during this time, accounting for 39% of cases (Centers for Disease Control and Prevention, 2022). While Foodnet only covers approximately 16% of the US population, serovar trends align closely with data from nationwide clinical cases (Figure 1.1). During each year of this period, these three serovars were consistently the top three serovars, except for 2020, where serovar Javiana replaced Typhimurium, although the COVID-19 pandemic likely affected data from this year. Notably, these three serovars are identified most often globally, combining to account for 64.1% of all *Salmonella* cases between 2001 and 2007 (Hendriksen et al., 2011).

Additionally, antimicrobial resistance (AMR) in *Salmonella*, while generally decreasing, continues to be problematic for human clinical cases (Michael & Schwarz, 2016). AMR is a significant contributor to higher financial burden for *Salmonella* illnesses, increasing the length and cost of hospital stays and increasing morbidity of illness (Chiu et al., 2002). This is shown by a persistent multidrug-resistant strain of serovar Newport (REPJJP01) between 2016 and 2022. Beef and cattle products from Mexico associated with this outbreak were linked to high numbers of human illnesses, resulting in one-third of identified patients being hospitalized (Ford et al., 2023). To better understand the extent of AMR in the US, the National Antimicrobial Resistance Monitoring System (NARMS) was established in 1996 in partnership between the CDC, FDA, and USDA to identify incidence and patterns of antimicrobial resistance in humans, food animals, and retail meats. *Salmonella*, *Campylobacter*, *Enterococcus*, and *E. coli* isolates from humans, routine USDA surveillance and FDA retail sampling are tested using a combination of minimum inhibitory concentration (MIC) testing and whole

genome sequencing (WGS) analysis. From 1996 to 2023, 20.5% (10,808/52,471) of human clinical isolates have shown resistance to one or more antimicrobials (Figure 1.4) (CDC, 2025). The percentage of resistant isolates per year has slightly decreased over this time, likely due to improving AMR stewardship in agricultural and medical fields as well as continued efforts both nationally and internationally to combat antibiotic-resistant bacteria (CDC, 2015; FDA, 2015). While regulations such as the implementation of the veterinary feed directive and stewardship programs in hospitals have reduced the impact of AMR, alternative antibacterial approaches and improved antibiotic stewardship education are needed to reduce the burden of *Salmonella* infections.

Although *Salmonella* is an enteric pathogen, human health can be impacted through consumption of non-animal sources as well. Surprisingly, produce is also a major contributor to *Salmonella* outbreaks, accounting for the largest proportion of illnesses from outbreaks since 2012 (Figure 1.3). For these outbreaks to occur, produce must be contaminated by feces from an animal source; however, the sources of contamination are not well characterized. To better prevent these contamination events, there is a need to better understand the sources and routes of contamination on produce.

Wildlife is a largely understudied source of *Salmonella* (Langholz & Jay-Russell, 2013). As wild animals can move freely between natural habitats, agricultural land, and human-populated areas, the transmission of microbes between humans and animals is likely to occur (Dias-Alves et al., 2023). While often limited to sampling of sporadic fecal samples, *Salmonella* surveillance in wildlife has been accomplished, including raccoons (Maurer et al., 2015a; Very et al., 2016), deer (Renter et al., 2006; Salas-Rosas et al., 2020; Topalcengiz et al., 2020),

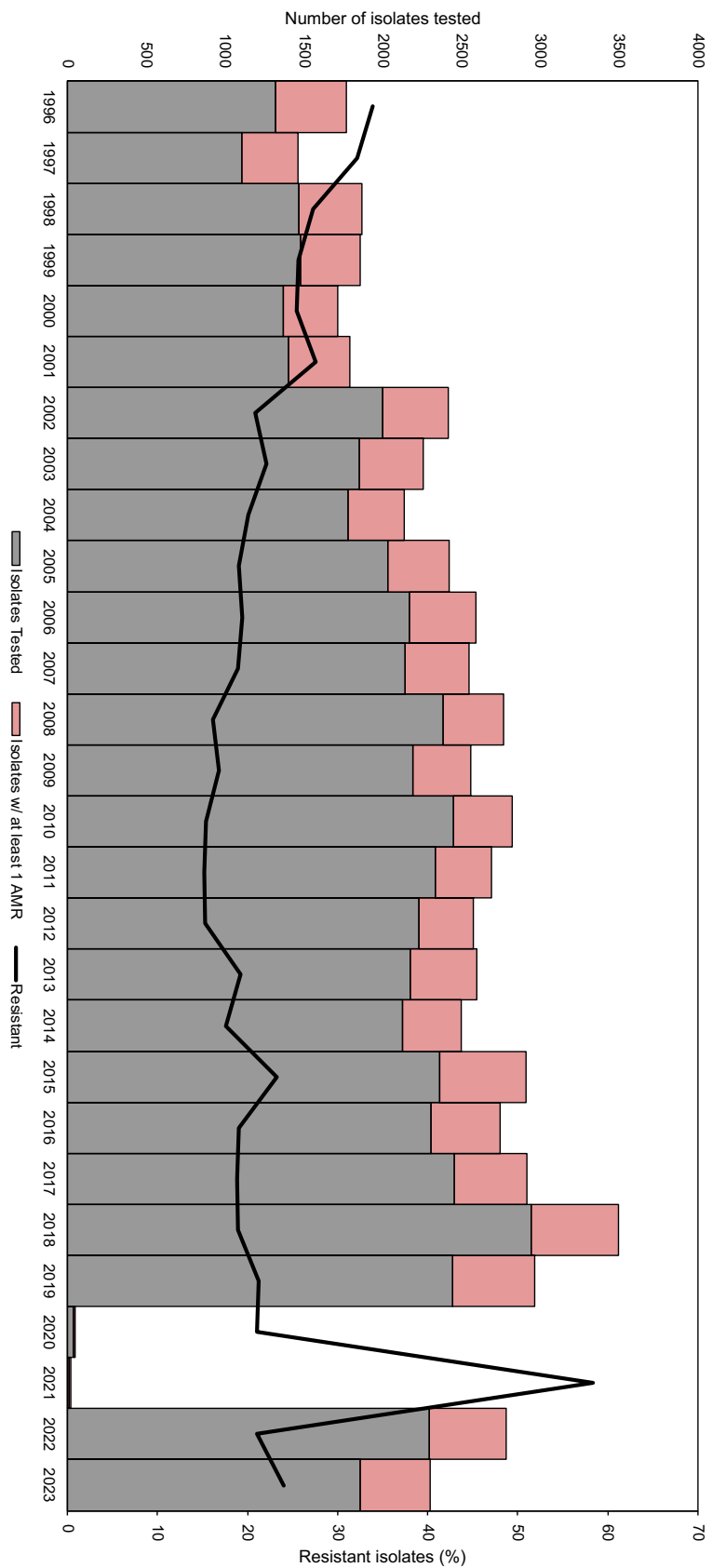


Figure 1.4. Instance of antimicrobial resistance in human *Salmonella* isolates (1996-2023). Human *Salmonella* isolates collected by the CDC with at least one antimicrobial resistance (red) are shown in comparison to all isolates tested (gray). The percentage of resistant isolates (black line) is shown for each year. Years 2020 and 2021 contained significantly fewer samples due to reduced sampling during COVID-19. Data adapted from CDC NORS (National Outbreak Reporting System (NORS), 2022).

reptiles (Doden et al., 2021; Pulford et al., 2019; Saelinger et al., 2006), bats (Adesiyun et al., 2009; McDougall & Power, 2021), and other animals (Millan et al., 2004; Uelze et al., 2021). These studies are necessary, as they can identify serovars not often identified in food animals. Importantly, these serovars are sometimes associated with human illness, suggesting the need for continued surveillance (Pees et al., 2023; Waltenburg et al., 2022). Due to the expertise required for extensive tracking, trapping, and sampling of these sources, longitudinal studies and large-scale sampling efforts have been limited. This knowledge gap restricts the development of mitigation strategies and risk assessment associated with wildlife.

Additionally, *Salmonella* can be found in abiotic sources such as soil and dust. This pathogen is particularly adaptable to conditions outside a host where it has been recovered as many as 332 days following application into soil (You et al., 2006). The ability of *Salmonella* to persist in the environment is exhibited in studies showing detection in dust particles, where viable isolates can be recovered after a significant reduction in moisture (Khouja et al., 2024; Pal et al., 2021). This ability to survive outside a host also allows for the transmission of *Salmonella* from a host to the environment by the spread of fecal material. Fecal samples are an effective tool for studying environmental transmission of *Salmonella*, as feces are indicative of the serovars being shed into the environment. Prevalence and quantification of *Salmonella* within various wildlife feces have been detected for 364 days, suggesting that wildlife pathogens can remain present in the environment for long periods (Topalcengiz et al., 2020). If *Salmonella* is present in the environment, the potential impact on human health can have long-term consequences (Oni et al., 2015). *Salmonella* persistence in these

abiotic sources contributes to food safety threats in produce, as transmission from soil or dust to produce can be difficult to prevent.

1.2 Produce Safety and *Salmonella*

Compared to meat and poultry, produce contaminated with *Salmonella* poses a unique problem, in that there is often no “kill” step to eliminate pathogens (Carstens et al., 2019). Cooking meat and poultry, if done properly, will prevent foodborne illness; however, most produce is eaten raw. After produce harvesting, an antimicrobial wash is often used containing an agent such as peracetic acid (PAA) or a chlorine product; however, this only affects the portion of the surface that encounters the wash (Krishnan et al., 2023). In many cases, the surface of the produce can often contain folds, holes, or other barriers to this wash, where pathogens such as *Salmonella* can “hide” from the antimicrobial agent (Barak et al., 2005). This is particularly concerning in many leafy greens including spinach and lettuce that can trap contaminants between leaves (Grivokostopoulos et al., 2022). Additionally, there is evidence to suggest that some *Salmonella* can invade produce via fruit, stems, leaves, or blossoms of fruiting vegetation, shielding the pathogen from washing (Burris et al., 2020; Hintz et al., 2010). Because of this, mitigation to prevent initial contamination is considered the most effective way to enhance produce food safety. Mitigation strategies include proper management of worker hygiene, irrigation water, soil and soil amendments, and wildlife before contacting the produce.

Between 2009 and 2023, the National Outbreak Reporting System (NORS) identified 638 *Salmonella* outbreaks from produce alone (CDC, 2024a). Significant outbreaks include the sickening of 1,127 and 1,040 people from whole onions in 2020

and 2021, respectively (CDC, 2024d). This outbreak was unique, as produce outbreaks are generally short-lived due to short shelf-life; however, onions have a long shelf-life and therefore contributed to months-long investigations during these outbreaks. A history of serovar Newport outbreaks in cucumbers and tomatoes from the eastern shore of Virginia has most recently resulted in the sickening of 275 patients and a single death in 2014. This persistent strain has been identified in at least 10 outbreaks and likely persists in the environment (Gruszynski et al., 2014a). Finally, a 2022 outbreak of serovar Typhimurium in melons led to the sickening of 87 people (FDA, 2023b). While no contamination source was identified, traceback efforts suggested a link to contaminated soil amendments (FDA, 2023a). As only a portion of these outbreak tracebacks identify the cause of produce contamination, this suggests a need for improved source attribution through rigorous surveillance testing.

Pathogen introduction to produce

The origins of foodborne pathogen contamination in fruits and vegetables are quite diverse (Figure 1.5) (Alegbeleye et al., 2018; Devarajan et al., 2023). These routes of contamination can include the application of contaminated soil or soil amendments that can transfer to growing plants or that can be splashed onto produce following rain or irrigation (Jechalke et al., 2019a; Lee et al., 2019). Biological amendments such as animal manure can have great nutritional benefits for produce plants; however, *Salmonella* can be identified at high loads and for extended periods in manure if left untreated or improperly treated (Hutchison et al., 2004; You et al., 2006). *Salmonella* has the potential to invade germinating seedlings or attach to the leaves and stems of early-stage plants (Barak & Liang, 2008; FDA, 2011). To mitigate these risks, regulations have

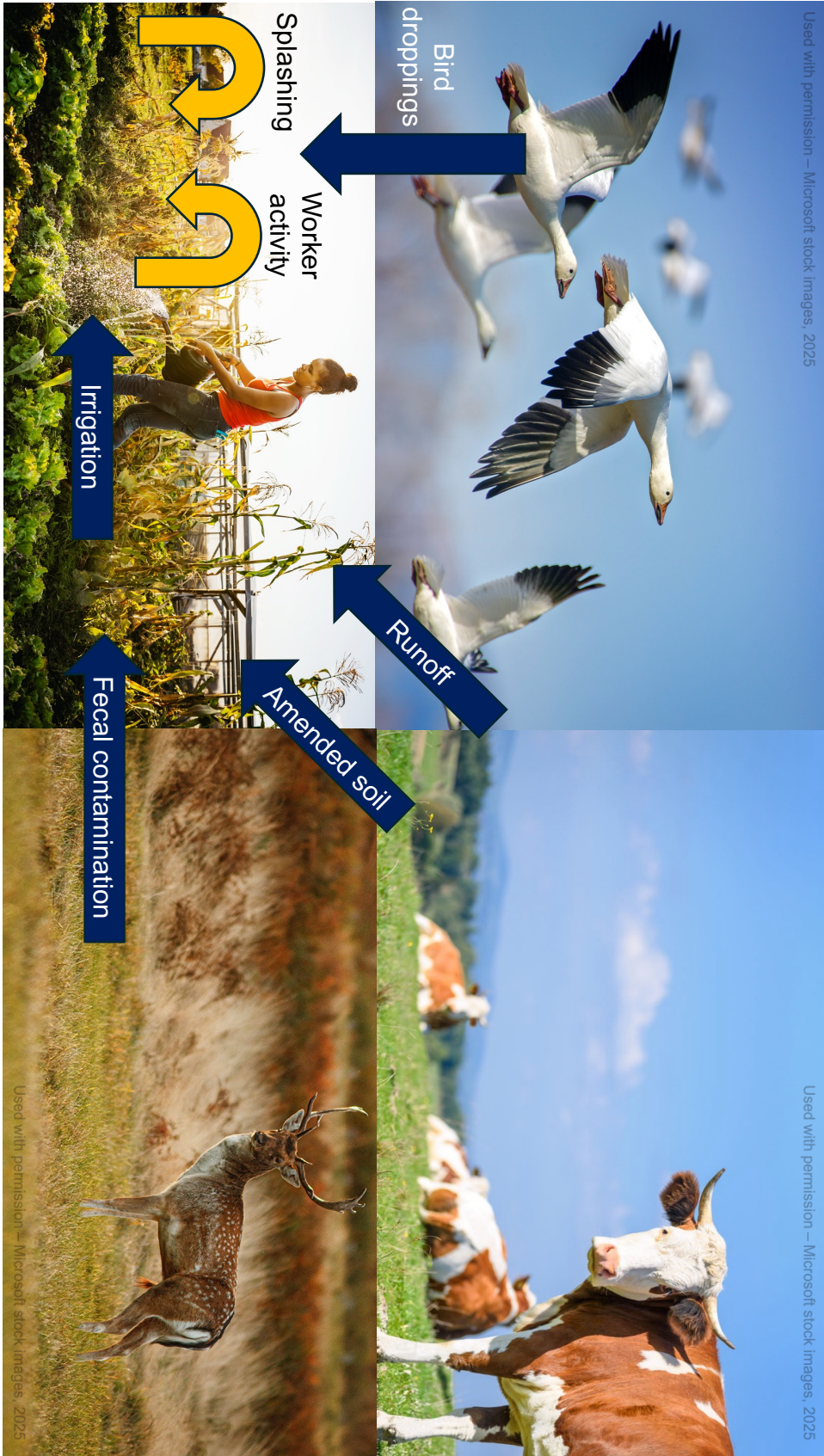


Figure 1.5. Potential contamination introduction routes in a produce environment.

Growing fields face threats of pathogen introduction through wildlife, neighboring agriculture, irrigation water, and soil. Arrows represent examples of potential introduction (blue) and amplification (yellow) mechanisms from these sources to a growing environment.

led to more robust treatment standards for soil amendments and the time required between application and planting (FSMA Final Rule on Produce Safety, 2024).

Another route of *Salmonella* introduction to produce is following the application of contaminated irrigation water. A recent outbreak in cucumbers, which caused 551 illnesses and 155 hospitalizations, was attributed to serovars Braenderup and Africana and linked to the use of contaminated irrigation water from a canal (CDC, 2024c).

Studies have shown that irrigation water can act as a reservoir for *Salmonella* and other foodborne pathogens and can be transmitted onto produce during growth (M. Cooley et al., 2007; Gorski et al., 2011). Click or tap here to enter text. Surface water is the most common source for irrigation, but also introduces the highest food safety risk, as upstream contamination can occur from wildlife, agricultural runoff, or sewage leaks (Johannessen et al., 2015). These risks can be managed using proper treatment methods including combinations of chlorination, peracetic acid, and UV radiation that are effective in reducing bacterial load in irrigation water (Krishnan et al., 2021). Municipal and well water pose a lower risk of contamination intrusion; however, application with these water sources as well as treated surface water, can collect contaminants in the soil and spread them to root vegetables (Islam et al., 2005).

Wildlife intrusion presents yet another source of produce contamination. These events can include direct contamination via defecation during foraging or indirect contamination via deposition into irrigation water or neighboring soil (Gruszynski et al., 2014a; Langholz & Jay-Russell, 2013; Topalcengiz et al., 2020). Generally, mitigating wildlife intrusion can be accomplished using barriers to physically restrict access (e.g., netting, spikes, and fences) or deterrents (e.g., decoys, sonic devices, and reflective

surfaces) (Franklin & VerCauteren, 2016). These management efforts, however, are not always feasible for large farms, as they can be quite expensive and require regular upkeep to be effective (Jung et al., 2014). Controlling wild bird intrusion is particularly challenging, as bird-specific mitigation can be costly and ineffective if not used properly. Strategies such as deterrents lose efficacy quickly and need to be performed in rotation to be successful. Additionally, physical restriction and deterrents may not be effective for all species, requiring a combination approach utilizing multiple mitigation strategies that can become overwhelming for growers (Rivadeneira et al., 2018; Varriano et al., 2025).

Food safety regulation in produce

The Food Safety Modernization Act (FSMA) was passed in 2011 and contained the most sweeping reform of US food safety laws in over 70 years (FDA, 2011). This act included the following seven primary rules to regulate food systems: The Produce Safety Rule, Preventative Controls for Human Food, Preventative Controls for Animal Food, Foreign Supplier Verification Program, Accreditation of Third-Party Auditors/Certification Bodies, Sanitary Transportation of Human and Animal Food, and Prevention of Intentional Contamination/Adulteration. The Produce Safety Rule sought to reduce the risk associated with produce contamination during growing, harvesting, and packing, where *Salmonella* is most likely to come into contact with produce (Produce Safety Rule, 2015). The rule establishes science-based minimum standards for the safe growing, harvesting, packing, and holding of produce (Produce Safety Rule, 2015). The combination of these regulations, when correctly applied, can significantly reduce the risk of *Salmonella* contamination of produce; however, contamination can still occur.

Within the Produce Safety Rule, specific requirements for agricultural water, biologically amended soils, and domestic and wild animal contamination aim to reduce the introduction of *Salmonella* and other foodborne pathogens (Produce Safety Rule, 2015). The final rule on pre-harvest agricultural water seeks to encourage growers to assess points of risk within their field and manage risks accordingly. These points include water quality, crop characteristics, environmental conditions, neighboring land use, and other relevant factors (Figure 1.6). Growers can then utilize appropriate measures such as water and soil testing to reduce contamination risk. The use of biologically amended soils is supported under the produce safety rule; however, standards are set to prevent contamination of produce. These include strict limits on detectable amounts of certain bacteria such as *Salmonella*, *Listeria*, fecal coliforms, and *E. coli*. Additionally, a minimum period of 120 days between the placement of amended soil and harvesting of produce is required to allow for bacterial die-off. If the produce has no contact with the soil, this period is shortened to 90 days. Finally, reasonable measures must be taken to identify produce contaminated from wild animal feces and avoid harvesting; however, no animal exclusion mandates are in effect. Reasonable measures can include marking contamination with a flag, where some certification agencies such as the Good Agricultural Practices (GAPs) require mandatory no-harvest zones around these contamination markers.

These regulations are based on a well-established analysis of hazards within the produce environment. Hazards are defined as any biological, chemical, or physical property that may cause a food to be unsafe for human consumption, such as contaminated water, improperly treated soil amendments, or wildlife intrusion in produce



Figure 1.6. Agricultural water assessment and risk-based outcomes. Growers are responsible for following this risk assessment of water sources, deciding what hazards are unique to their farm; and making the best decision for their situation. Information adapted from the Produce Safety Rule (FDA, 2015).

fields (Pathogen Reduction; Hazard Analysis and Critical Control Point (HACCP) Systems; Final Rule, 1996). Significant work has been completed to characterize these hazards, but more work is needed to translate this into risk analysis. A risk is the likelihood and consequences of hazards contributing to unwanted outcomes. For example, *Salmonella* in contaminated irrigation water is considered a hazard; however, the possibility of irrigation water contamination by agricultural runoff or ineffective treatment of contaminated water would be considered a risk. Risk assessment in a produce environment requires extensive analysis of the literature corresponding to unique farm features that affect a grower such as neighboring land use, sanitation practices, and other relevant influencing factors.

1.3 *Salmonella* in wild birds

While *Salmonella* is best characterized in humans and food animals, wild animals such as wild birds have been understudied (Franklin & VerCauteren, 2016; Hernandez et al., 2012; Navarro-Gonzalez et al., 2020; O. M. Smith, Edworthy, et al., 2020). The term “wild birds” refers to all non-domesticated avian species, excluding domesticated poultry and pet birds. Wild birds can be classified taxonomically, and can also be grouped by their social behaviors, such as songbirds (passerines), waterbirds (ducks, cranes, etc.), and larids (gulls, terns, etc.) (Prum et al., 2015). By grouping species into these categories, social behavior including foraging and nesting can provide insight into their risk to human health concerning *Salmonella*. While many birds are asymptomatic carriers of *Salmonella*, songbirds are the most studied group due to their susceptibility to increased morbidity caused by salmonellosis (Alley et al., 2002; Hall & Saito, 2008; Hernandez et al., 2012; Tizard, 2004). Common songbirds include finches, sparrows, and

woodpeckers, where these birds generally feed on insects, fruits, and seeds. As social species, they will also frequent bird feeders and common gathering spaces. This social nature has contributed to the high transmissibility of salmonellosis within their populations and also to incidences of transmission to humans (M. H. Murray et al., 2021; Patel et al., 2023).

A single bird can be colonized via the consumption of contaminated food or water, then proceed to transmit the bacteria to other individuals in a flock. Waterbirds are particularly high risk for *Salmonella* as a significant portion of their activity is in or near surface water, where *Salmonella* can be readily found (M. H. Murray et al., 2021).

Surface water is defined as any accessible water source above ground, including freshwater sources such as lakes, rivers, and aquifers, but also non-freshwater sources like brackish water and oceans. Ponds and lakes where waterbirds are commonly found tend to be collection points for human and animal waste runoff (Gorski et al., 2022; Murphy et al., 2022; Rodrigues et al., 2019; Weller et al., 2020); therefore, it is unsurprising that waterbirds have been shown to have increased prevalence of *Salmonella* (Hernandez et al., 2016; Krawiec et al., 2015; Tardone et al., 2020). Larid species are commonly found along shorelines, where a natural diet consists of a variety of crustaceans, insects, and fish, but will also include scavenging on animal carcasses (Andersson, 1970). This scavenging behavior has led to significant overlap in habitat with humans, where species such as seagulls will often forage from waste facilities (Olsen & Larsson, 2004; Ramos et al., 2010) leading to increased *Salmonella* prevalence (Antilles et al., 2021).

Other groupings of birds such as birds of prey (falcons, owls, etc.) and Pelecaniformes (pelicans, ibises, etc.) have been studied as well, where predatory species are estimated to have high foodborne pathogen prevalence due to the consumption of contaminated rodents or birds (Tizard, 2004). As urban populations grow, more species are driven from native environments, where interactions with humans have led to increased *Salmonella* prevalence and food safety risks, such as in the case of the white ibis (Hernandez et al., 2016). Alternatively, migratory species such as sandhill cranes have a range of approximately 5,000 miles, where dissemination of clinically relevant bacteria is of higher concern during this migration period (Callaway et al., 2014; Elsohaby et al., 2021). *Salmonella* is most often explored in its ability to spread directly from birds to humans such as in the 2021 outbreak or to cause explosive disease within wild bird populations, as described in the mass mortality of songbirds in the US in 2009 (Hernandez et al., 2012; Patel et al., 2023).

Salmonellosis and adaptation of *Salmonella* in avian hosts

Salmonella does not commonly cause illness in birds but rather is carried as a commensal organism in the intestinal tract (Janecko et al., 2015; Silva et al., 2018). There are, however, significant exceptions. (Cohen et al., 2021; Fukui et al., 2014; Hernandez et al., 2012; Mather et al., 2016; Patel et al., 2023; Refsum et al., 2003). In poultry, *Salmonella* serovars Gallinarum and its biovar Pullorum are associated with fowl typhoid and Pullorum disease, respectively, where both are largely fatal (Bullis, 1977). Genomic alterations including those responsible for the expression of SPI-1 and metabolism have reduced the ability of these pathogens to colonize non-poultry hosts (Langridge et al., 2015) Due to the host-specific nature of these serovars and their ability to cause high

levels of mortality, testing and culling programs have successfully eliminated them from the commercial poultry industry in the United States (Bullis, 1977; Kabir, 2010).

Sporadic cases of *Salmonella* epidemics in wild birds do occur, where serovar Typhimurium is generally the causative agent (Cohen et al., 2021). While no system is in place to regularly monitor *Salmonella* presence and diversity in wild birds, research groups and diagnostic laboratories have identified *Salmonella* in deceased birds (Hall & Saito, 2008; Hernandez et al., 2012). Avian *Salmonellosis* is often expressed through neurological dysfunction, ruffled feathers, and esophageal lesions that make swallowing difficult and lead to weight loss (Friend et al., 2015). Outbreaks are often widespread and occur in winter months due to increased reliance on feeders and heightened transmission during migration (Refsum et al., 2003). A recent outbreak was particularly interesting as it involved avian mortality and human illness (Patel et al., 2023). At least 30 people contracted salmonellosis due to contact with dead birds or contaminated feeders, supporting the need for improved sanitation practices in supplemental bird feeding. By focusing studies on avian salmonellosis, the knowledge of *Salmonella* in wild birds is biased toward only the most invasive strains in birds. This may not be an accurate reflection of strains present across wild bird species including those that could have a more significant impact on human health. For effective risk assessment within wild birds, more information about the prevalence and nature of *Salmonella* in this host is needed.

Avian host adaptations are well documented in serovar Typhimurium, where the definitive phage types (DTs) 40, 56v, and 160 have become widespread within migratory birds and have led to outbreaks of explosive mortality in waterfowl, songbirds, and other species (Alley et al., 2002; Cohen et al., 2021; Fu et al., 2022; Fukui et al., 2014;

Hernandez et al., 2012; Patel et al., 2023; Refsum et al., 2003). These phage types often have distinct genomic lineages that include a loss of the pSLT virulence plasmid (Mather et al., 2016). A second example of this is a newly identified serovar, Tirat-Zvi, which excels in the colonization of passerine species while also showing reduced virulence in mice (Cohen et al., 2024). Additionally, these host-adapted strains are also found to have an accumulation of pseudogenes, or alterations that cause translated proteins to no longer function normally, which often result in the loss of function for virulence factors (Cohen et al., 2021). The loss of these genes that facilitate general colonization and the retention of genes specific to the colonization of avian species has led to significant differences in virulence between wild birds and other hosts (Cohen et al., 2021). This process of transitioning from a host-generalist to a host-specialist by the formation of pseudogenes and gene inactivation is known as genome decay (Baumler & Fang, 2013; Langridge et al., 2015). While specific genes allowing for the increased virulence of these strains have not been identified, the reduction of fitness within non-target hosts is hypothesized to increase fitness within the specific host.

Wild birds provide a unique *Salmonella* risk as they can cover large distances and do not face the same barriers to movement as many other wildlife (Elsohaby et al., 2021; Rivadeneira et al., 2018). This likely facilitates the accumulation and spread of *Salmonella* in birds from a variety of sources and geographic locations, making them a particularly interesting subject of study. To further investigate avian *Salmonella* isolates, Enterobase is a useful database containing genomes from a variety of human, animal, and environmental sources from around the world (Zhou et al., 2020). This database currently contains 2,179 isolates of *Salmonella* that have been associated with wild avian sources.

Of these, 1079 are serovar Typhimurium. This is unsurprising as this generalist serovar is frequently identified in many different hosts, including wild birds (Alley et al., 2002; Cohen et al., 2021; Fukui et al., 2014). Additionally, 154 isolates of serovar Enteritidis and 134 isolates of serovar Infantis are present that are linked to wild birds. Notably, these three serovars that are frequently associated with avian hosts also account for 33% of human clinical *Salmonella* isolates in the United States, suggesting a potential association between these hosts (Figure 1.1).

***Salmonella* prevalence in wild birds**

In the United States, regional testing as part of smaller research studies has shown that *Salmonella* prevalence in birds ranges from 0.5% to 2.5% on the West Coast and 1.9% in the Southwest (Gorski et al., 2011; Jay-Russell & Justice-Allen, 2014; Navarro-Gonzalez et al., 2020; Rivadeneira et al., 2016; O. M. Smith, Edworthy, et al., 2020; O. M. Smith, Olimpi, Navarro-Gonzalez, et al., 2022). Studies outside of the United States, including those in Europe, South America, the Middle East, and Eastern Asia regularly find similar levels of *Salmonella* prevalence in wild birds (Cardoso et al., 2021; Cohen et al., 2021; Lawson et al., 2011; Palmgren et al., 2006). As urbanization and industrialization reduce the natural habitat for wild birds, it is unsurprising that interactions between humans in urban populations or animal agriculture increase. Many of these studies have identified human-associated serovars carried in wild birds, including those more commonly identified in human infections (Hernandez et al., 2016; O. M. Smith, Edworthy, et al., 2020). Wild birds interacting with human populations can become colonized with *Salmonella* via foraging in garbage and refuse (i.e. pigeons, crows, ibises) (Hernandez et al., 2016; Ramos et al., 2010), while those interacting with

animal agriculture (i.e. starlings, blackbirds, and sparrows) can be recipients of transmission from feed, feces, or irrigation water in these settings (Kirk et al., 2002; O. M. Smith, Olimpi, & Karp, 2022).

Public health risks attributed to wild birds

Wild birds can transmit *Salmonella* to humans directly and indirectly. Recently, a 2021 outbreak in humans was attributed to wild birds, where human interactions with bird feeders and deceased birds lead to illness (Patel et al., 2023). In 2008, the CDC reported an outbreak of *Salmonella* Typhimurium linked to peanuts (Wittenberger & Dohlman, 2010). Source tracking later identified that this strain was found in the peanut fields and also in wild birds that were sampled around the fields (Hernandez et al., 2012). Although wild birds were not directly implicated in this outbreak, closely related strains found in birds support the hypothesis that feces from the wild birds contaminated the peanut products. This form of contamination has also been linked to other pathogens, including the 2004 outbreak of *Campylobacter jejuni* that was caused by feces from sandhill cranes on pea plants in Alaska; however, this case involved many unique factors such as the consumption of raw, uncleaned product at a local level and the farm location being near a large wildlife refuge (Gardner et al., 2011). Previous work has investigated factors that affect the survival and transmissibility of *Salmonella* in feces at these foraging sites, finding that recovery can occur for weeks after inoculation (Oni et al., 2015; Topalcengiz et al., 2020). Within animal agriculture environments, wild birds can also indirectly impact human health. Some species such as European starlings, brown-

headed cowbirds, and Brewer's blackbirds are categorized as pest species within concentrated animal feeding operations (CAFOs) where they will forage on the animal feed and fecal material from livestock (Palmer, 1976). Here, *Salmonella* and other foodborne pathogens can be transmitted from birds to livestock, contributing to the pathogen burden in these food production operations (Carlson et al., 2011, 2020).

In the US, wild birds have been identified as a risk for *Salmonella* contamination in regions dominated by monoculture crop production and animal agriculture (O. M. Smith, Edworthy, et al., 2020). Studies around produce fields have identified a range of prevalence, from 0.5% in west coast studies (Gorski et al., 2011; Navarro-Gonzalez et al., 2020; O. M. Smith, Edworthy, et al., 2020) and 1.9% in a southwest study (Rivadeneira et al., 2016). FSMA and the Produce Safety Rule, while urging growers to manage wildlife contamination, do not set guidelines on how to prevent intrusions, which adds to the challenges growers face in trying to control wildlife. The efficacy of wild bird mitigation strategies such as decoys, lasers, air cannons, and falconry have been debated, leading to conflicting measures to reduce bird intrusion (Rivadeneira et al., 2018; Varriano et al., 2025). Conflict is also demonstrated in the debate over the protection or removal of riparian zones (Karp et al., 2015; Strawn et al., 2013). These strips of non-crop vegetation along the perimeter of produce fields are thought to encourage the foraging of wild birds and insects; however, others argue that any encouragement of wildlife near fields increases risk. Considering the relatively low prevalence of *Salmonella* within wild birds from previous studies, the financial burden of mitigation and the potential impact of ecological disruption have recently been questioned (Olimpi et al., 2022, 2024; O. M. Smith, Snyder, et al., 2020) calling for co-management practices

instead. This lack of confidence in risk assessment and mitigation strategies produces a food safety question; what role do wild birds have in pathogen transmission to fresh produce? To answer this question, it is first necessary to identify what is known about the potential for avian *Salmonella* transmission in produce fields.

1.4 Surface water

For produce growers using surface water, a combination of microbial testing and mitigation strategies ensures that only water within a safe range is used (FDA, 2011).

Currently, there are no tests available that can directly and efficiently test for *Salmonella* presence. Most tests target a combination of total coliforms, fecal coliforms, and generic *E. coli* within water. These organisms are associated with fecal contamination and are used because they can serve as indicators for enteric pathogens such as *Salmonella*.

However, previous work has shown an inconsistent relationship between these indicator species and *Salmonella* (Gu et al., 2021; McEgan et al., 2013; Weller et al., 2020).

Although these regulations are in place to prevent the use of unsafe water, *Salmonella* outbreaks in produce continue to implicate contamination in irrigation sources (CDC, 2024c; FDA, 2021). This highlights a critical need to understand *Salmonella* in surface water, which has been well-studied (Gorski et al., 2022; Gu et al., 2019; Hintz et al., 2010; Li et al., 2015; Micallef et al., 2012).

***Salmonella* introduction and survival in surface water**

Salmonella is an enteric pathogen, suggesting that aquatic environments would be inhospitable; however, multiple domestic and international studies have shown high *Salmonella* prevalence in water (Chen et al., 2024; Deaven et al., 2021; Haley et al., 2009; H. Liu et al., 2018; Murphy et al., 2022; Toro et al., 2022). *Salmonella* can be

introduced to surface water via direct or indirect contamination by humans and animals (Casanova et al., 2020; Li et al., 2015; Topalcengiz et al., 2019). When wildlife or livestock have direct contact with water, defecation is likely to occur. Human sewage is also often directed to rivers, providing an influx of new pathogens while also providing fresh nutrients that facilitate pathogen proliferation (Olds et al., 2018). The treatment of this sewage greatly reduces its microbial impact; however, studies have identified *Salmonella* following this treatment as it returns to surface water (Kinde et al., 1997; Odjadjare & Olaniran, 2015). Wastewater conveyance, or the carriage of sewage to the treatment plants which often occurs alongside waterways, can also experience leaks that introduce pathogens (Roehrdanz et al., 2017). Alternatively, indirect contamination often takes the form of agricultural runoff (Gorski et al., 2011). For example, precipitation events can push fecal material from animal agriculture fields downhill to the lowest areas where it will eventually find a surface water source. Surface water can be naturally formed such as creeks or ponds, or can be manmade for fecal runoff of cattle feed yards or dairies, such as lagoons. Presumably, runoff from non-agricultural land can also wash fecal material deposited by wildlife into surface water. Where improperly treated biologically amended soils are used in plant agriculture, they, too, can be washed away following precipitation and irrigation (You et al., 2006). Barriers and treatments can be used to limit this runoff; however, such mitigation is often costly, which reduces the sustainability of these approaches.

In the water, *Salmonella* encounters a variety of stressors that it must overcome to survive. These include osmotic pressure, oxidative stress, pH, temperature, and radiation as well as a significantly diluted supply of essential nutrients (Winfield & Groisman,

2003). Some studies have suggested that *Salmonella* merely persists in the water in a metabolically inactive state or survives in a suboptimal state until re-entering a host (Roszak et al., 1984; Santo Domingo et al., 2000). Alternatively, *Salmonella* is capable of proliferating, either planktonically, or within biofilms or free-living protozoa (Brandl et al., 2005; Byappanahalli et al., 2009; Gaertner et al., 2011; Sha et al., 2013). Previous work has shown that viable *Salmonella* can be found in freshwater after 300 days without any added nutrients (Topalcengiz et al., 2019) and was detected five years after being inoculated into sterile water (Liao & Shollenberger, 2003). Others have observed a significant decrease in viability within days of inoculation into water (Oguadinma et al., 2022). This suggests that *Salmonella* could enter a viable but nonculturable (VBNC) state that allows it to persist for long periods.

Methods for *Salmonella* sampling from surface water

The two overarching methods for *Salmonella* sampling in water are Moore swabs and membrane filtration. Moore swabs are a piece of cheesecloth that is tied in the middle and placed in a sampling environment for 24-48 hours where free-floating *Salmonella* can become caught in the fabric (Moore, 1948; Sikorski & Levine, 2020). The swab is then collected and placed into culture media. As a more efficient method, modified Moore Swabs are collected on-site for 2-30 minutes. Here, a larger piece of cheesecloth is rolled tightly and placed inside a PVC cassette where between 0.1L to 10L of water is pumped through the cheesecloth filter (Bisha et al., 2011). This actively filters any *Salmonella* floating by the filter rather than relying on passive capture. The cheesecloth filter can then be removed and cultured, similar to a conventional Moore swab. Alternatively, membrane filtration traditionally begins by collecting a set volume of

water, generally between 100mL to 1L (Entis, 1990; Lindquist et al., 2007; H. Liu et al., 2018; McEgan et al., 2013). This sample is then filtered using vacuum filtration or adding in perlite “pool” filter into the sample (Meinersmann et al., 2008). Work completed by the CDC has also led to the development of a highly sensitive membrane filtration method known as dead-end ultrafiltration (DEUF) where up to 100L of water are pumped on-site through a small filter on-site before backflushing the filter into 100mL to 500mL of solution that can then be cultured (C. M. Smith & Hill, 2009).

These protocols have advantages and disadvantages that should be considered when choosing research methods. Moore swabs allow the researcher to examine the population of *Salmonella* in the water over time (Bisha et al., 2011; Cho et al., 2020). Additionally, modified Moore swabs can be utilized to analyze varying volumes of water, giving researchers control of the sample size. Prior work compared *Salmonella* recovery in 0.1L, 1L, and 10L using this method and found that 10L samples had 43.5 and 25.5 times better recovery than 1L and 0.1L, respectively (Sharma et al., 2020). Membrane filtration has the benefit of higher sensitivity compared to Moore swabs (Kraft et al., 2023), as samples from this collection method can be quantified due to the concentration of bacteria in a small filter. Additionally, the resulting backflush can be split into multiple growth conditions, allowing for the detection of various organisms from a single sample (Amin et al., 2020; P. Liu et al., 2012; McEgan et al., 2013). Alternatively, Moore swabs are advantageous due to their comparatively low cost and ease of use (P. Liu et al., 2021); however, this method requires researchers to return to the sampling site the following day, making this method less time-efficient than others (McEgan et al., 2013). These

advantages and disadvantages demonstrate the need to take careful consideration of collection methods when sampling water sources.

***Salmonella* prevalence in water**

Salmonella prevalence in water has been well characterized in the United States, with studies taking place in major agricultural regions, including the Southeast, West Coast, Mid-Atlantic, and Northeast (Figure 1.7). The largest proportion of these studies have been completed in the Southeast, where prevalence ranges from 4.8% to 100% (Bell et al., 2015; Gu et al., 2019; Haley et al., 2009; Havelaar et al., 2017; Li et al., 2014; Luo et al., 2015; McEgan et al., 2013, 2014; Murphy et al., 2022, 2024; Rajabi et al., 2011; Vereen et al., 2013). Sampling in the Mid-Atlantic has resulted in a prevalence ranging from 4.2% to 65% (Acheamfour et al., 2024; Callahan et al., 2019; S. Kim et al., 2023; Kraft et al., 2023; Micallef et al., 2012; Sharma et al., 2020). A smaller number of studies in the Northeast, including those in New York and Pennsylvania, have found *Salmonella* prevalence ranging between 4.6 and 49% (Deaven et al., 2021; Strawn et al., 2013; Weller et al., 2020). West Coast studies have been concentrated in agricultural regions of California, where prevalence ranges from 6% to 65% (Benjamin et al., 2013; M. B. Cooley et al., 2014; Gorski et al., 2011, 2022; Partyka et al., 2018). Additionally, international sampling has included a wide range of global regions, where similarly variable *Salmonella* prevalence ranged from

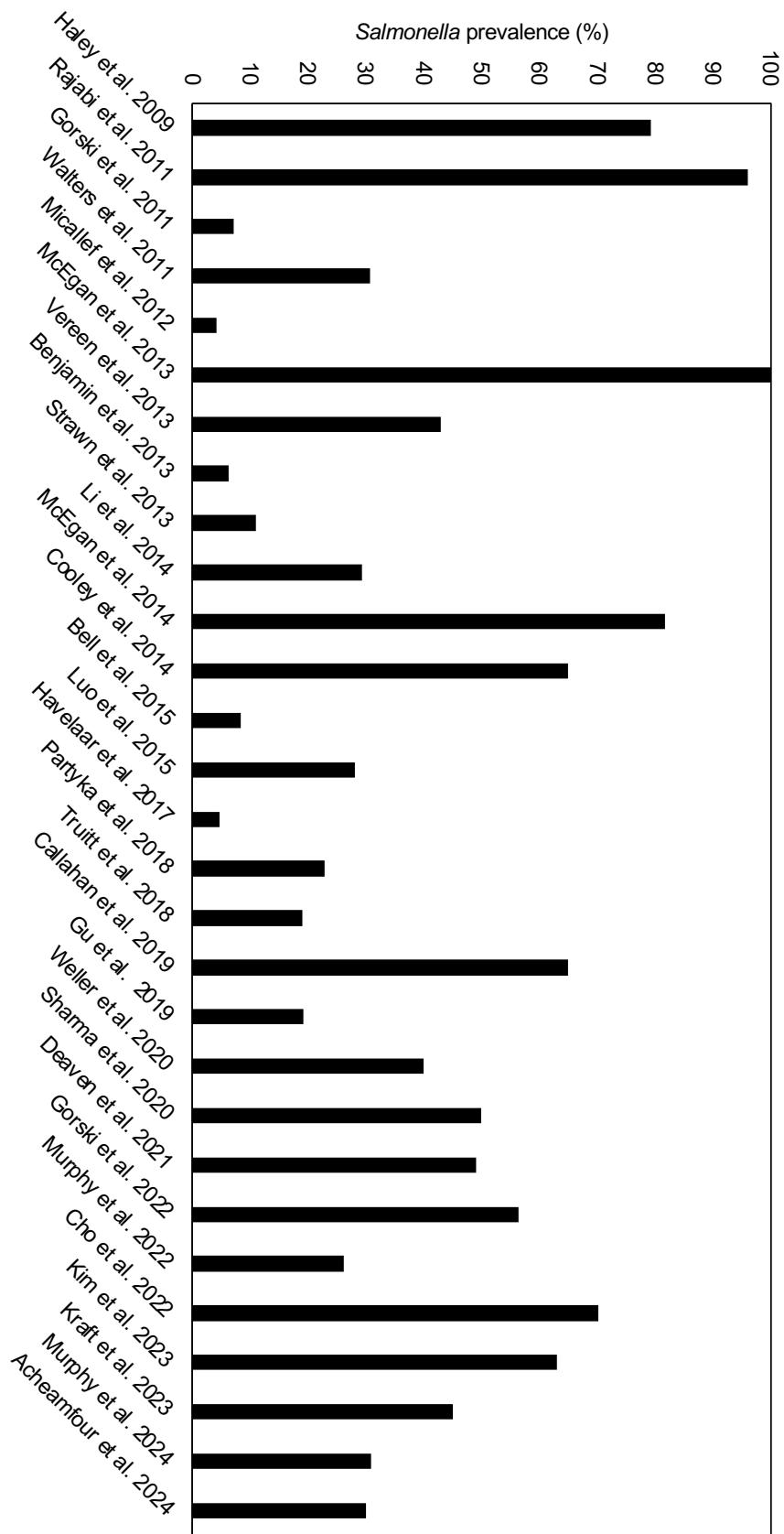


Figure 1.7. *Salmonella* prevalence in surface water studies in the United States.

Chronological display of various studies within the United States that tested for *Salmonella* prevalence in surface water. Prevalence is measured in percentage and labeled with the first author and the year of study.

7.1% to 78.4% (Afema et al., 2016; Arvanitidou et al., 2005; Chen et al., 2024; Díaz-Torres et al., 2020; Ho et al., 2018; Huang et al., 2014; Mahagamage et al., 2020; Polo et al., 1998; Ruiz et al., 1987; Santiago et al., 2018; Setti et al., 2009; Song et al., 2018; Thomas et al., 2013; Toro et al., 2022). These studies show that *Salmonella* can be regularly identified in surface water; however, the prevalence can vary significantly.

In addition to regional differences, the type of water in these studies may influence *Salmonella* prevalence. Generally, water collected from moving sources such as rivers, creeks, and streams yielded *Salmonella* in more than 50% of samples for many studies (Cho et al., 2022; Haley et al., 2009; McEgan et al., 2013, 2014; Rajabi et al., 2011). Alternatively, studies that used stagnant water such as ponds or agricultural water reservoirs often contained *Salmonella* in less than 25% of samples (Gu et al., 2019; Havelaar et al., 2017; Micallef et al., 2012; Partyka et al., 2018; Strawn et al., 2013; Truitt et al., 2018). This suggests *Salmonella* may benefit from the circulation of nutrients or mixing of microbial communities in larger creeks and rivers.

Environmental factors influencing *Salmonella* potential in water

The presence of *Salmonella* in water can be influenced by meteorological factors, such as temperature, wind, and solar radiation, and also by physiochemical factors such as pH, turbidity, dissolved oxygen, and total dissolved solids (Weller et al., 2020). Prior work has identified positive correlations between *Salmonella* prevalence and higher temperatures, including both air and water temperatures (Deaven et al., 2021; Gorski et al., 2011; Haley et al., 2009; H. Liu et al., 2018; Murphy et al., 2022; Thomas et al., 2013; Weller et al., 2020). Additionally, the incidence of rainfall before sampling was often associated with increased *Salmonella* recovery (Deaven et al., 2021; Setti et al.,

2009; Strawn et al., 2013; Thomas et al., 2013; Weller et al., 2020). The association of *Salmonella* prevalence with the presence of indicator species such as *E. coli* was mixed, with some finding positive associations (Weller et al., 2020) and others finding negative correlations (McEgan et al., 2013). Alternatively, many studies found a lack of correlation between these same variables, such as rain (Goyal et al., 1977; McEgan et al., 2013; Santiago-Rodriguez et al., 2012), temperature (Strawn et al., 2013), and indicator species (McEgan et al., 2013; Murphy et al., 2022). Rarely, negative correlations were identified; however, these were not consistently found across studies, for example with conductivity (McEgan et al., 2013). Inconsistencies between these variables highlight the complexities associated with understanding which factors most greatly influence *Salmonella* in water. A better understanding of how these factors affect *Salmonella* introduction and viability in surface water is necessary for building risk assessment models. If these predictive tools could be reliably implemented, growers would be able to monitor water and weather conditions and make more informed decisions about safe water usage.

Global surface water research and improvement efforts

Salmonella cases worldwide are estimated as high as 176 million illnesses annually, where enteric fever from typhoidal serovars contributes to 26 million cases (CDC, 2024b). Notably, infections from typhoidal-serovars almost exclusively occur in Latin America, Sub-Saharan Africa, and Southeast Asia (IHME, 2021b, 2021a). Alternatively, the United States accounted for less than 500 culture-confirmed cases between 2016-2018, a majority of which were attributed to international travel (CDC,

2024b). Typhoidal infections often occur via contact with contaminated water, making surface water studies in these regions particularly important.

Antibiotics are regularly used for human and animal health around the world; however, their misuse has contributed to an increasing concern of AMR, becoming a threat to human health in the United States and internationally. AMR directly caused 1.27 million deaths in 2019 and contributed to 4.95 million deaths, making it a top concern of the World Health Organization (C. J. L. Murray et al., 2022; WHO, 2023).

Antimicrobial-resistant *Salmonella* are regularly identified in surface water, with prevalence ranging from 2% to 100% (Berge et al., 2006; Casanova et al., 2020; Chen et al., 2024; Cho et al., 2022; Dolejská et al., 2009; Gorski et al., 2011; Jokinen et al., 2015; Li et al., 2014; Luo et al., 2015; McEgan et al., 2014; Meinersmann et al., 2008; Micallef et al., 2012; Nguyen et al., 2021; Patchanee et al., 2010; Somda et al., 2021; Suhartono et al., 2021). Varying levels of antibiotic stewardship exist across different geographic regions (Maron et al., 2013). Without proper use and regulation, antimicrobial residues can influence bacterial populations and the presence of AMR genes in the environment (Beattie et al., 2018; Carroll et al., 2015; Dolejská et al., 2009; Laborda et al., 2022; Ramey & Ahlstrom, 2020). Surface waters have been known to act as a reservoir for antimicrobials, which can facilitate transmission to the environment and genetic transfer between microbes (Chen et al., 2024; Wilkinson et al., 2022).

Due to the increasing threat of enteric fever and antimicrobial resistance, global studies focusing on *Salmonella* in surface water are needed to better understand transmission and introduction sources. To address this, a team of researchers from Latin America, in a coordinated effort with the Joint Institute for Food Safety and Applied

Nutrition (JIFSAN), have been collecting water samples beginning in 2019. The overall prevalence in these three studies was 45.4%, where teams in Brazil, Chile, and Mexico found individual prevalence of 68.2%, 33.1%, and 62.9%, respectively (Chen et al., 2024). In these studies, antimicrobial resistance was prominent, where 33.8% of all isolates were resistant to at least one antimicrobial. (Chen et al., 2024). In Chile, *Salmonella* prevalence was identified at 28.1% between two major agricultural regions where crop presence was the most significant factor in predicting pathogen presence (Toro et al., 2022). Similar to studies in the United States (Deaven et al., 2021; Gorski et al., 2011; Haley et al., 2009; Murphy et al., 2022), this study found that seasonality was a driving factor for prevalence. While much of the research from this joint effort is forthcoming, the data made available from these studies thus far shows the need to continue surveillance of agricultural water internationally as a means of reflecting the food safety risks associated with contaminated water. This research and others around the world continue to improve our understanding of *Salmonella* distribution, diversity, and influencing factors. While not specific to *Salmonella*, the assessment of implemented measures to improve water quality found that conditions remained consistent, rather than degrading, following large urbanization (Johannessen et al., 2015). This demonstrates the need for improved management strategies as well as the benefit of targeted water quality regulations. Additionally, work completed in Mexico supported previous findings of high *Salmonella* levels in summer months (Estrada-Acosta et al., 2018; González-López et al., 2022). This work also identified a strain of serovar Oranienburg over multiple months, suggesting potential adaptations for persistence. Dynamics involved in the persistence

and transmission of *Salmonella* have not been well characterized prior to this work, highlighting the importance of elucidating these pathways.

Global urbanization, waste management practices, and improper use of waterways contribute to contamination of surface water across the (Tickner et al., 2017). This use of improperly managed water contributes to the estimated 93.8 million global cases of gastroenteritis and 59,100 deaths annually caused by *Salmonella* species (Majowicz et al., 2010; Roth et al., 2018). In their 2023 sustainable development goals report, the United Nations (UN) stated that 2.2 billion people around the world lacked access to safely managed drinking water. Because of this, the UN has listed access to clean water and sanitation as one of their 17 major goals to meet before 2030, indicating the importance of this natural resource to people around the world (Burden of Disease Attributable to Unsafe Drinking-Water, Sanitation and Hygiene, n.d.). This global effort to improve availability and access to safe water will rely on international research efforts to provide data used to meet these goals.

1.5 Methods of *Salmonella* isolation

Traditional *Salmonella* isolation includes four steps: 1) pre-enrichment, 2) selective enrichment, 3) indicator plating, and 4) confirmation (Figure 1.8). Depending on the sample type being tested, the procedure may vary between protocols following the Bacterial Analytical Manual (BAM) method (FDA, 2024a) or Microbiology Laboratory Guidebook (MLG) (USDA-FSIS, 2024b). In each protocol, these steps seek to promote *Salmonella* growth by inhibiting competing microbes while developing an environment suitable for *Salmonella* growth. The BAM is followed by the FDA for produce and environmental sampling, while MLG is followed by the USDA-FSIS for the culture of

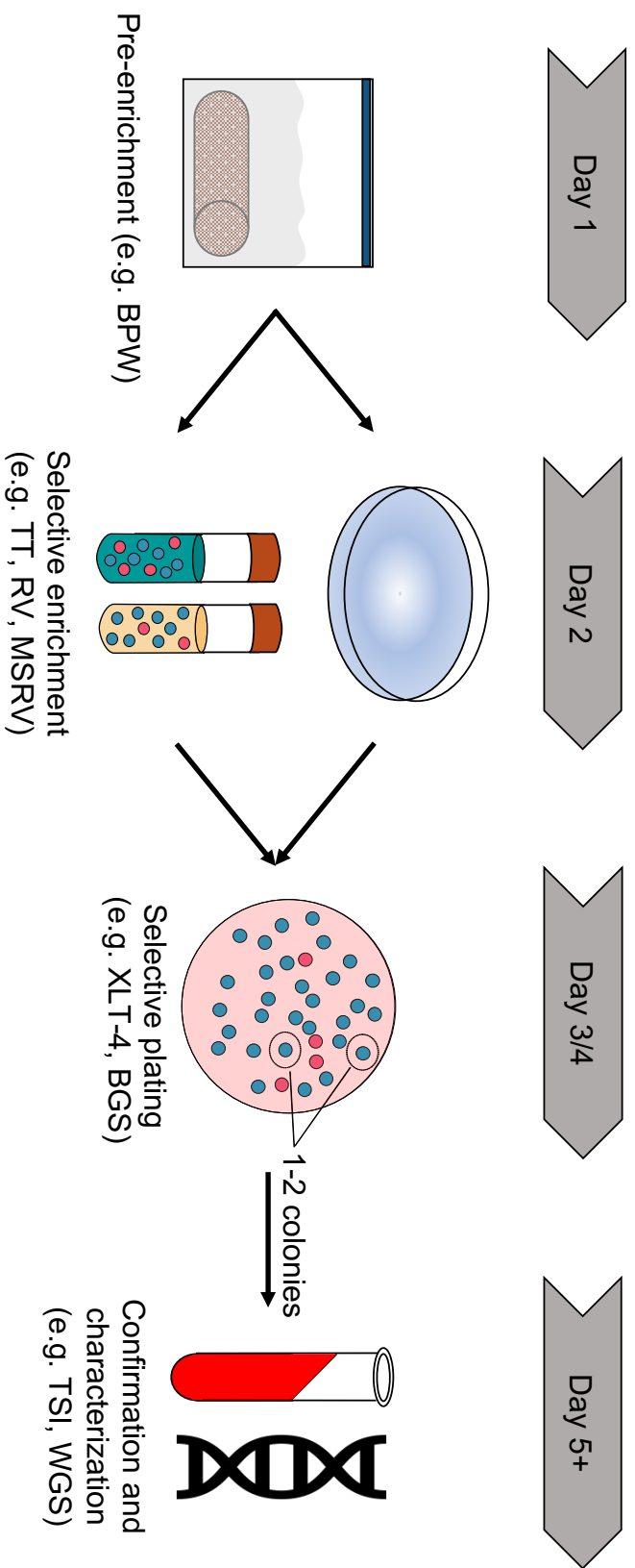


Figure 1.8. *Salmonella* isolation flow diagram. The general flow of *Salmonella* isolation begins with a pre-enrichment, followed by selective enrichment and selective plating. Confirmation of presumptive positive isolates concludes isolation.

meat, poultry, and pasteurized egg products. Chemical and physical differences in sample composition, such as pH, water activity, and fat content can drastically affect the recovery of *Salmonella* due to potential interactions with media or other microbial growth (Gibson et al., 1988). Additionally, recovery can be influenced by growth media, selecting for serovars better adapted to the conditions in one media but selecting against serovars without those adaptations (Gorski, 2012), Therefore, ensuring that the right protocol for a given matrix is key to effective isolation of *Salmonella*.

A non-selective pre-enrichment or “recovery” medium is often used as the first step of *Salmonella* isolation, as cells recovered from environmental sampling are often damaged and unable to compete (Budu-Amoako et al., 1992). Buffered peptone water (BPW) is often used for this step, as it contains peptones as a nutrient source and sodium chloride for electrolytes, but no selective agents. In samples from produce or food animal processing where antimicrobials are used, neutralizing agents are added to the recovery medium to prevent the continuation of antimicrobial properties from further injuring *Salmonella*. These neutralizing media including Dey-Engley (DE) broth used during FDA for environmental sampling during outbreak investigations and neutralizing BPW (nBPW) used by the USDA-FSIS, to prevent the continuation of antimicrobial properties from further injuring *Salmonella* (FDA, 2024b; Gamble et al., 2017). Samples are incubated between 37°C and 42°C for up to 24 hours to allow for the recovery of sub-lethally injured cells. As this medium is non-selective, background and competing microbes will also grow and potentially outcompete *Salmonella*, therefore an additional step is necessary to isolate *Salmonella* from the non-target bacteria (Busse, 1995; Rasamsetti et al., 2021).

Following pre-enrichment, samples are then transferred into one or more selective enrichment media. The two most common media used for *Salmonella* isolation are tetrathionate (TT) and Rappaport-Vassiliadis (RV) broths. TT broth takes advantage of a metabolic pathway present in many *Salmonella* serovars but missing in other species (Winter et al., 2010). This pathway allows *Salmonella* to metabolize tetrathionate as an energy source while providing no additional energy source for other species, preventing the growth of competing bacteria (D'aoust, 1981; Knox, 1945). Additionally, bile salts are used to inhibit gram-positive bacterial growth along with the presence of calcium chloride to balance pH, as sulfuric acid is produced by metabolic activity (W. Moats, 1981; Winter et al., 2010). An alternative TT formulation, TT Hajna, includes further selective ingredients including brilliant green and sodium desoxycholate, both of which further inhibit the growth of gram-positive bacteria (Hajna & Damon, 1956). RV broth contains a combination of low pH, malachite green, and magnesium chloride. *Salmonella* survives well at lower pH, while malachite green inhibits the growth of coliforms that are often identified in environments with *Salmonella* (Rappaport et al., 1956; Vassiliadis, 1983; Vassiliadis et al., 1981). Magnesium chloride acts to increase the osmotic pressure, creating a hypertonic solution that gram-negative organisms can tolerate, but gram-positive organisms cannot (Peterz et al., 1989). While both media are acceptable for recovery of *Salmonella*, TT medium has improved recovery when low microbial loads are present (Hammack et al., 1999). RV medium can often be altered depending on sample type, including the addition of novobiocin to further reduce gram-positives, or the addition of agar to form a semi-solid medium that retains selectivity while also allowing for the separation of motile and non-motile organisms (Aspinall et al., 1992; KOMATSU

& RESTAINO, 1981). After inoculation into selective enrichment media, cultures are incubated for an additional 24 hours.

Following enrichment, samples are then streaked onto an indicator plate that continues to selectively enrich for *Salmonella* while utilizing a visual change in the color of the colony or plate to identify presumptive-positive *Salmonella* colonies. Xylose Lysine Tergitol-4 (XLT-4), Brilliant Green Sulfate (BGS), and Xylose Lysine Deoxycholate (XLD) agar are commonly used. Both XLD and XLT-4 agar utilize a trait in most *Salmonella* serovars, the ability to produce hydrogen sulfide (H₂S) during fermentation, to turn H₂S (+) colonies black while leaving H₂S (-) colonies white (Taylor, 1965). XLT-4 utilizes tergitol, a surfactant that is bacteriostatic against gram-positive bacteria and *Proteus* species (Miller et al., 1991). XLD does not utilize tergitol; however, novobiocin is often added at concentrations ranging from 5ug/mL to 80ug/mL to reduce gram-positive bacteria (Restaino et al., 1982). BGS agar utilizes brilliant green to reduce gram-positive bacteria and the inability of *Salmonella* to ferment lactose by introducing phenol red to turn fermenting bacteria yellow, leaving non-fermenting bacteria pinkish red (W. A. Moats & Kinner, 1974). Finally, confirmation of *Salmonella* is needed after indicator plating to mitigate false positives from visually similar colonies such as surviving *Proteus* species on XLD or *Citrobacter* species on XLT-4. These confirmation assays can include biochemical tests with triple sugar iron slants (Krumwiede & Kohn, 1917), physical reactions with serum agglutination assays (Olopoenia & King, 2000), or molecular confirmation using polymerase chain reaction (PCR) for the *invA* gene (Rahn et al., 1992).

Variations of the culture methods described above include delayed secondary enrichment (DSE) and consecutive selective enrichment. DSE is the process of inoculating pre-enrichment culture into TT broth, the medium is left at room temperature for seven days, where the culture is then transferred to fresh TT and incubated for 24 hours (Waltman et al., 1991). Consecutive selective enrichment consists of pre-enrichment culture inoculation into TT broth and, after 24 hours of incubation, passaging this culture into RV broth for an additional 24 hours before plating (Rigby & Pettit, 1980). Additionally, alternative selective media have been used for the growth of *Salmonella*, including selenite broth and its variable formulations which utilize sodium selenite to inhibit the growth of gram positives and some gram negatives, while permitting *Salmonella* growth (Zimboro et al., 2009). A semisolid agar formulation of RV medium is also commonly used to maintain the same selective pressure as traditional RV while further selecting for motility from the inoculation site (Aspinall et al., 1992).

1.6 Characterizing *Salmonella* isolates

Following *Salmonella* confirmation, the next step in characterization includes serotyping to identify the serovar, which is then followed by subtyping to provide more granular strain information.

Serotyping

Traditional serotyping is performed via serum agglutination (Wattiau et al., 2011). This process utilizes the antigenic structure of *Salmonella* described in the White-Kauffman-Le minor scheme (P. A. D. Grimont & Weill, 2007). A complete serotyping analysis requires the possession of antisera for 46 somatic, 119 flagellar, and the capsular antigen (B. Liu et al., 2014; McQuiston et al., 2004). For each isolate, a separate test is

conducted for the somatic antigen along with testing for both H1 and H2 flagellar antigens, which can exist due to the ability of *Salmonella* to phase switch between the expression the *fliC* and *fljB* genes, and finally testing for the presence or absence of the capsular Vi antigen (P. A. D. Grimont & Weill, 2007; Silverman et al., 1979). Due to the cost associated with obtaining all necessary antisera and the difficulty in applying these tests for multiple isolates, this analysis is often conducted by specialty diagnostic laboratories. Alternatively, the identification of the O-antigen alone, referred to as serogrouping, can be used to gather general information from isolates and can be a useful research tool for screening isolates.

Alternative serotyping methods include molecular approaches: one of the first methods used was the amplification of abequose (*rfbJ*) and paratose synthase genes (*rfbS*) of common *Salmonella* O-groups (A, B, C2, and D). The presence of these group-specific sequences can be used to identify the serotype of an isolate (Luk et al., 1993). Additional work found group-specific sequences within the *rfb* gene cluster for groups E (Wang et al., 1992), H (Fitzgerald et al., 2003), O:35 (Wang & Reeves, 2000), and O:54 (Keenleyside et al., 1994). Due to the variability of the *rfb* gene cluster, a multiplex PCR was created to differentiate the presence of common serogroups (B, C1, C2, D, E, and O:13) using a Luminex platform (Dunbar et al., 2003; Fitzgerald et al., 2007). Further inclusion of group-specific and flagellar genes led to the development of a *Salmonella* multiplex assay for rapid typing (SMART) analysis to serogroup the 50 most common serovars in the United States (Leader et al., 2009). These methods utilize variability within surface antigen genes to separate serotypes; however, this can be accomplished

through other methods as well and many subtyping approaches described below can also accurately indicate serovar identity.

Non-antigenic molecular serotyping methods include insertion sequences, ribosomal variation, and other genomic targets. Insertion sequences (IS), primarily IS200, were identified as a prominent trait within a majority of human clinical isolates, where mutational variation within this sequence has been used for serotype identification and the additional distinction between some strains (Gibert et al., 1990; Lam & Roth, 1983; RUBINO et al., 1998). Additionally, utilizing variation within ribosomal DNA, by restriction enzyme mapping of 23S and 16S genes (Altwegg et al., 1989; Esteban et al., 1993) and the intergenic space, known as intergenic sequence ribotyping (ISR) (Lagatolla et al., 1996) can be used to differentiate serovars. Another method for serogrouping is phage-typing, which utilizes the highly specific viral receptor proteins to identify the O-group of an isolate (Anderson et al., 1977; Anderson & Williams, 1956; Demczuk et al., 2003). These methods, while efficiently serotyping isolates, cannot obtain significant strain information that is useful for outbreak tracebacks.

Subtyping

To better differentiate strain information, pulsed-field gel electrophoresis (PFGE) was developed to give isolates a genetic “fingerprint” that can be compared to other isolates and identify potential outbreaks. This method utilizes restriction enzymes (typically XbaI, BlnI, SpeI, SfiI, PacI, and NotI for *Salmonella*) to cut the chromosome at various locations, resulting in a series of DNA bands of different lengths (Schwartz & Cantor, 1964). An oscillating electric current is then used to separate these large bands on an agarose gel; the resulting gel image can then be compared to a PFGE database (e.g.,

CDC's PulseNet) to determine the genetic proximity of related isolates (Ricke, 2014). This method can discern strains within a serovar, but as this subtyping method only shows lengths of DNA and not sequence, it is not always sufficient for outbreak tracking, especially in the case of highly clonal serovars (Allard et al., 2013).

To satisfy the need to track outbreaks using DNA sequence information, multilocus sequence typing (MLST) was developed to elucidate the sequence of seven highly conserved housekeeping genes for serovar Typhi: *aroC*, *dnaN*, *hemD*, *hisD*, *purE*, *sucA*, and *thrA* (Kidgell et al., 2002). In this process, all seven genes are individually amplified by PCR and sequenced. Each gene is given an allelic identifier and the combination of the seven alleles is represented by a sequence type (ST). The strain is then given an overall sequence type (MLST) and compared to other strains, where a match in six out of seven genes and a change in the seventh gene would result in a different sequence type. This subtyping procedure is effective in characterizing non-typhoidal *Salmonella* strains; however, the depth of subtyping is limited to the sequence of these seven genes, therefore more closely related strains or highly clonal serovars, such as Enteritidis, are not able to be separated using this method (Achtman et al., 2012; Maiden, 2006).

Analogous to MLST, Clustered Regularly Interspaced Short palindromic repeats (CRISPR) typing is an alternate subtyping method. The CRISPR system in *Salmonella*, like many other prokaryotes, evolved to defend against conjugative plasmids and viral predation via acquired immunity (Grissa et al., 2007; Touchon & Rocha, 2010). In *Salmonella enterica* subspecies *enterica*, the presence of two CRISPR loci, separated by approximately 20kb, consist of highly conserved direct repeats 29 nucleotides long and

highly variable spacers 32 nucleotides long (Fabre et al., 2012; F. Liu, Barrangou, et al., 2011; F. Liu, Kariyawasam, et al., 2011). Following the survival from a conjugative plasmid or viral infection, a 32-nucleotide sequence from the foreign DNA is added at the leading end of the CRISPR loci. These loci, along with CRISPR-associated (*cas*) genes can then be expressed, where this sequence can be used to identify the presence of a repeated invasion by the same foreign DNA (Barrangou et al., 2007; Garneau et al., 2010; Pourcel et al., 2005). Interestingly, this system is no longer functional as a defense mechanism in *Salmonella*; however, the presence of the CRISPR loci remains consistent (Shariat et al., 2015).

CRISPR-typing utilizes the hypervariability of the spacer sequences between isolates belonging to the same serovar as a subtyping approach in *Salmonella* (F. Liu, Barrangou, et al., 2011). This process involves the independent amplification of both CRISPR loci and the comparison of their sequences, which, depending on the application, can be compared to a database of analyzed isolates (Fabre et al., 2012; Shariat & Dudley, 2014). This method was developed originally in *Mycobacterium tuberculosis* (Kamerbeek et al., 1997) and was called spoligotyping, and has since been adapted to Group A *Streptococcus* (Hoe et al., 1999), *Campylobacter* (Schouls et al., 2003), and *E. coli* (Delannoy et al., 2012). Notably, the CRISPR system remains active in *M. tuberculosis* and Group A *Streptococcus*, and is no longer active in *Campylobacter* and *E. coli*; however, these systems are commonly identified across tested isolates (Wei et al., 2019, Nozawa et al., 2011, Yeh et al., 2024, Touchon et al., 2011). Similarly to the hypervariability of CRISPR regions, virulence genes experience significant selective pressure and therefore evolve at a higher rate (Endo et al., 1996). Combining sequence

analysis of *Salmonella* virulence genes *fimH* and *sseL* with CRISPR loci improved the subtyping capability of nine clinically relevant serovars in a process called CRISPR-multi-virulence-locus sequence typing (CRISPR-MVLST) (F. Liu, Kariyawasam, et al., 2011). This method has since been adapted to many other *Salmonella* serovars (Almeida et al., 2017; Shariat, DiMarzio, et al., 2013; Shariat, Kirchner, et al., 2013; Shariat, Sandt, et al., 2013; Vilela et al., 2024; Vosik et al., 2018). Similar to MLST, this method is restricted in its subtyping discrimination, as analysis based on a small number of variable genomic regions does not account for the full genomic diversity between strains. Interestingly, the conservation of these molecular targets within a serovar also allows for CRISPR-typing and MLST to be used for serotyping. For example, CRISPR spacer content is highly associated with serovar identity (Deng et al., 2015; Fabre et al., 2012), as is MLST (Achtman et al., 2012).

Advances in sequencing technology have allowed scientists to utilize bacterial genomes for strain discrimination rather than individual components, known as whole genome sequencing (WGS). This technology is used to characterize individual genomes that can be compared to one another or to entire databases to assess genomic differences, such as gene presence or absence, insertion of genetic elements including plasmids and transposons, or single nucleotide polymorphisms (SNPs). These differences can then be used to characterize the phylogenetic relationship of strains, where closely related strains can be identified during outbreak investigations. This highly discriminatory subtyping tool is the foundation for GenomeTrakr, an open-source database of whole genome sequences used to detect closely related strains and identify sources of transmission (Allard et al., 2016; Timme et al., 2019). In these cases, WGS can identify SNPs and

more targeted genetic characteristics to trace outbreaks (M. Hoffmann et al., 2016). SNP-based characterization using WGS can include as few as the seven genes used in traditional MLST, or as much as the whole genome (wgMLST), with varying levels of depth, including ribosomal (rMLST) and core genome (cgMLST) (Mohammed & Thapa, 2020). The utility of WGS for outbreak tracebacks and higher-resolution characterization has been demonstrated by its ability to outperform PFGE, MLST, and other tools in recent outbreaks (Allard et al., 2012, 2013; den Bakker et al., 2014; Deng et al., 2015; M. Hoffmann et al., 2016; Lienau et al., 2011). Additionally, WGS allows for more robust serovar characterization by mapping serovar-specific genes beyond those responsible for surface antigens, such as the *Salmonella in silico* Typing Resource (SISTR) and SeqSero 2.0 (Yachison et al., 2017; Yoshida et al., 2016; S. Zhang et al., 2019a). SISTR identifies serovars via O-antigen flipase (*wzx*) and polymerase (*wzy*) genes and the *fliC* and *fljB* genes for the H1 and H2 determination, while SeqSero 2.0 targets the *rfb* gene cluster for the O-antigen and the *fliC* and *fljB* genes for H-antigen determination. For these reasons, WGS has become firmly established as the gold standard for *Salmonella* serotyping and subtyping. Having first been developed by the Center for Food Safety and Applied Nutrition at the US-FDA, WGS was adopted in 2014 by the CDC and USDA-FSIS for their *Salmonella* analyses.

1.7 Limitations of *Salmonella* isolation and characterization

All methods previously described above rely on the isolation of an individual isolate following *Salmonella* enrichment. Traditional isolation techniques rely on the picking of one to five colonies for characterization (FDA, 2024a; USDA-FSIS, 2024b); however, prior observations have suggested that six colonies would need to be picked to

reliably identify two serovars at equal proportions in a population (Cason J. et al., 2011). If a serovar composes only 10% of a population, 32 colonies would need to be picked. This depth of colony picking presents a serious limitation in *Salmonella* surveillance and source tracking analysis. Considering the upper limit of five picked colonies from a positive sample, this will often result in picking only the most abundant serovar or serovars in a mixed population. Consequently, serovars that account for a smaller proportion of the population are less likely to be detected, where these less abundant serovars have the potential to contribute to human illness (Berghaus et al., 2013; Cason J. et al., 2011).

To overcome this problem, a tool was developed that would be able to detect the presence of multiple serovars in a sample. CRISPR-SeroSeq is a deep serotyping method that utilizes the serovar-specific CRISPR spacers to differentiate serovars and provide a relative abundance of each serovar in a single sample (Thompson et al., 2018). In contrast to CRISPR-Typing, where each entire CRISPR locus is amplified by PCR, the primers used in CRISPR-SeroSeq target the highly conserved direct repeat sequences. This generates short sequences that collectively present the spacer content of *Salmonella* present in a given sample. By comparing the short sequence reads to a database of known spacer sequences, the serovars can be determined. Further, the total number of sequence reads belonging to each serovar within a sample can be used to determine the relative frequency of each serovar. Further, polyphyletic deliniations can be identified where a serovar can contain two or more distinct CRISPR profiles, indicating a potential evolutionary recombination event (Cherchame et al., 2022; Worley et al., 2018). This process has been used to identify multiseroovar populations within animal agriculture and

environmental samples, (Deaven et al., 2021; Siceloff et al., 2021; Thompson et al., 2018). Results from Thompson et al. 2018 showed that this technology can confidently identify serovars constituting as little as 0.01% of the population, though this is dependent on the depth of sequencing. In the first application of this technology, 91% of positive samples were found to contain multiple serovars, and up to four serovars were identified in a single sample. Siceloff et al. 2021 provided a parallel comparison of traditional isolation and CRISPR-SeroSeq deep serotyping in cattle fecal samples, where deep serotyping identified tetracycline-resistant serovar Reading in multiple isolates that were not identified by picking colonies. This provides context to the importance of identifying and characterizing all serovars in a population, as a tetracycline treatment given to these cattle would have allowed serovar Reading to survive. Relevant to the work here, Deaven et al. 2021 identified highly complex *Salmonella* populations in surface water, including up to ten serovars in a single sample. In this study, 80% of positive samples contained multiple serovars, most of which (78%) contained one of the top ten serovars associated with human illness. Importantly, serovars Typhimurium and Enteritidis were masked by other serovars in 78% and 71% of samples when present, respectively. This suggests a decreased likelihood of detecting these important serovars using traditional isolation methods. Additionally, this work began to characterize potential influences of *Salmonella* population complexity, such as precipitation and subsequent river discharge. The width of the Susquehanna River introduced a challenge in this study, limiting the ability to assess the full complexity of this water source. To overcome this challenge, more regular sampling of smaller creeks would better represent downstream and temporal changes. Collectively, this data suggests that *Salmonella*

populations in animals and the environment are highly complex and that these mixed populations need to be fully characterized to properly evaluate food safety risks. This method has transformed what we know about *Salmonella* ecology and serovar dynamics, and these findings suggest a need to re-evaluate and more deeply characterize the diversity and dynamics of *Salmonella* serovars in different environments and hosts.

1.8 Objectives for this dissertation

Salmonella enterica is a leading cause of bacterial foodborne illness in the United States. In addition to its high prevalence in food animals, *Salmonella* is commonly found in environmental sources such as wildlife and water; however, prevalence within these sources can be variable. Because of this, potential sources of contamination exist within a variety of scenarios (CDC/FDA/USDA, 2023). *Salmonella* is comprised of over 2,600 serovars, characterized by somatic (O) and flagellar (H) antigens (P. A. D. Grimont & Weill, 2007). In addition to antigenic differentiation, phenotypic and genotypic differences including host restriction (Pascopella et al., 1995), virulence factors (Cheng et al., 2019), and antimicrobial resistance (D. H. Shah et al., 2017) exemplify the genomic plasticity of *Salmonella*. Only 80 serovars account for 99% of human clinical cases in the US, and 13 of these contribute to over 75% of cases (CDC, 2024a). Many environmental sources of *Salmonella* are understudied such as wildlife and water, resulting in the inability to mitigate risks associated with these sources. It is critical to identify potential routes of contamination and prevent their entry into our food systems to reduce the number of illnesses attributed to *Salmonella*.

One challenge in this field of research is the lack of studies on the environmental prevalence of *Salmonella*. While animal agriculture and processing account for the bulk

of the literature about prevalence and contamination, environmental sources including wildlife and water require further investigation. The prevalence and diversity of serovars found within these understudied sources can provide important information about the routes of contamination that are not being protected. Previous work from our lab has demonstrated that *Salmonella* often exists in multiseroval populations, where traditional isolation techniques often only identify the dominant serovar in a population (Thompson et al., 2018). This creates a unique challenge in risk assessment, as less abundant serovars may contribute to negative health outcomes if left unidentified and untreated (Siceloff et al., 2021).

To address these obstacles and develop a further understanding of *Salmonella* prevalence and population dynamics in the environment, my doctoral research aims to evaluate the prevalence and diversity of *Salmonella* populations within previously understudied environmental reservoirs. This was completed by first assessing the risk associated with wild bird transmission of *Salmonella* in produce fields in the southeastern United States. In the first survey of *Salmonella* within wild birds in this region, I found a notable decrease in prevalence compared to studies on the West Coast, where pathogen viability was strongly associated with the freshness of feces. This work demonstrates a regional difference in risk assessment that growers can utilize for individual management strategies. I then characterized the prevalence and population of *Salmonella* in four distinct creeks over two years in the southeastern United States. This study found high *Salmonella* recovery and complexity associated with increased precipitation and humidity, while proximal land use was found to have a less significant influence. Additionally, many serovars not commonly found in food animal surveillance testing

were identified in water samples, suggesting a potential reservoir for clinically relevant *Salmonella*. Finally, I explored the population and prevalence of *Salmonella* in the rivers of both agricultural and urban rivers in Pinchinca, Ecuador to identify similarities and differences between these water sources. There I found significantly increased complexity within sites of the urban river compared to the agricultural river, including many clinically relevant serovars. Additionally, signs of persistence and downstream transmission were identified, suggesting a need to monitor and manage these important sources for safe, clean water. In conclusion, my work combines longitudinal studies with genomic and population sequencing to assess the prevalence, diversity, and risk associated with different environmental sources of *Salmonella*.

To establish a more robust understanding of the transmission of *Salmonella* between humans, food systems, and the environment to develop more effective food safety mitigation strategies, I employed two specific aims:

1. Assess the prevalence and complexity of *Salmonella* populations in wild birds and surface water sources.
2. Evaluate risk factors that influence *Salmonella* viability within wild bird and surface water environments.

1.9 References

National Outbreak Reporting System (NORS). (2022, February 3). National Outbreak Reporting System (NORS). Accessed on December 18, 2024.

<https://www.cdc.gov/nors/data/index.html>

Acheamfour, C. L., Parveen, S., Gutierrez, A., Handy, E. T., Behal, S., Kim, D., Kim, S., East, C., Xiong, R., Haymaker, J. R., Micallef, S. A., Rosenberg Goldstein, R. E.,

- Kniel, K. E., Sapkota, A. R., Hashem, F., & Sharma, M. (2024). Detection of *Salmonella enterica* and *Listeria monocytogenes* in alternative irrigation water by culture and qPCR-based methods in the Mid-Atlantic U.S. *Microbiology Spectrum*, 12(4). <https://doi.org/10.1128/spectrum.03536-23>
- Achtman, M., Wain, J., Weill, F.-X., Nair, S., Zhou, Z., Sangal, V., Krauland, M. G., Hale, J. L., Harbottle, H., Uesbeck, A., Dougan, G., Harrison, L. H., & Brisse, S. (2012). Multilocus Sequence Typing as a Replacement for Serotyping in *Salmonella enterica*. *PLoS Pathogens*, 8(6), e1002776. <https://doi.org/10.1371/journal.ppat.1002776>
- Adesiyun, A. A., Stewart-Johnson, A., & Thompson, N. N. (2009). Isolation Of Enteric Pathogens from Bats in Trinidad. *Journal of Wildlife Diseases*, 45(4), 952–961. <https://doi.org/10.7589/0090-3558-45.4.952>
- Afema, J. A., Byarugaba, D. K., Shah, D. H., Atukwase, E., Nambi, M., & Sischo, W. M. (2016). Potential Sources and Transmission of *Salmonella* and Antimicrobial Resistance in Kampala, Uganda. *PLOS ONE*, 11(3), e0152130. <https://doi.org/10.1371/journal.pone.0152130>
- Alegbeleye, O. O., Singleton, I., & Sant’Ana, A. S. (2018). Sources and contamination routes of microbial pathogens to fresh produce during field cultivation: A review. *Food Microbiology*, 73, 177–208. <https://doi.org/10.1016/j.fm.2018.01.003>
- Allard, M. W., Luo, Y., Strain, E., Li, C., Keys, C. E., Son, I., Stones, R., Musser, S. M., & Brown, E. W. (2012). High resolution clustering of *Salmonella enterica* serovar Montevideo strains using a next-generation sequencing approach. *BMC Genomics*, 13(1), 32. <https://doi.org/10.1186/1471-2164-13-32>

- Allard, M. W., Luo, Y., Strain, E., Pettengill, J., Timme, R., Wang, C., Li, C., Keys, C. E., Zheng, J., Stones, R., Wilson, M. R., Musser, S. M., & Brown, E. W. (2013). On the Evolutionary History, Population Genetics and Diversity among Isolates of *Salmonella* Enteritidis PFGE Pattern JEGX01.0004. PLoS ONE, 8(1), e55254. <https://doi.org/10.1371/journal.pone.0055254>
- Allard, M. W., Strain, E., Melka, D., Bunning, K., Musser, S. M., Brown, E. W., & Timme, R. (2016). Practical Value of Food Pathogen Traceability through Building a Whole-Genome Sequencing Network and Database. Journal of Clinical Microbiology, 54(8), 1975–1983. <https://doi.org/10.1128/JCM.00081-16>
- Alley, M., Connolly, J., Fenwick, S., Mackereth, G., Leyland, M., Rogers, L., Haycock, M., Nicol, C., & Reed, C. (2002). An epidemic of salmonellosis caused by *Salmonella* Typhimurium DT160 in wild birds and humans in New Zealand. New Zealand Veterinary Journal, 50(5), 170–176. <https://doi.org/10.1080/00480169.2002.36306>
- Almeida, F., Medeiros, M. I. C., Rodrigues, D. dos P., Allard, M. W., & Falcão, J. P. (2017). Molecular characterization of *Salmonella* Typhimurium isolated in Brazil by CRISPR-MVLST. Journal of Microbiological Methods, 133, 55–61. <https://doi.org/10.1016/j.mimet.2016.12.020>
- Altwegg, M., Hickman-Brenner, F. W., & Farmer III, J. J. (1989). Ribosomal RNA Gene Restriction Patterns Provide Increased Sensitivity for Typing *Salmonella* Typhi Strains. Journal of Infectious Diseases, 160(1), 145–149. <https://doi.org/10.1093/infdis/160.1.145>

- Amin, N., Liu, P., Foster, T., Rahman, M., Miah, M. R., Ahmed, G. B., Kabir, M., Raj, S., Moe, C. L., & Willetts, J. (2020). Pathogen flows from on-site sanitation systems in low-income urban neighborhoods, Dhaka: A quantitative environmental assessment. *International Journal of Hygiene and Environmental Health*, 230, 113619. <https://doi.org/10.1016/j.ijheh.2020.113619>
- Anderson, E. S., Ward, L. R., de Saxe, M. J., & de Sa, J. D. H. (1977). Bacteriophage-typing designations of *Salmonella* Typhimurium. *Journal of Hygiene*, 78(2), 297–300. <https://doi.org/10.1017/S0022172400056187>
- Anderson, E. S., & Williams, R. E. O. (1956). Bacteriophage Typing of Enteric Pathogens and Staphylococci and its Use in Epidemiology: A Review. *Journal of Clinical Pathology*, 9(2), 94–127. <https://doi.org/10.1136/jcp.9.2.94>
- Andersson, A. (1970). Food Habits and Predation of an Inland-Breeding Population of the Herring Gull *Larus argentatus* in Southern Sweden. *Ornis Scandinavica*, 1(2), 75. <https://doi.org/10.2307/3676026>
- Andino, A., & Hanning, I. (2015). *Salmonella enterica*: Survival, Colonization, and Virulence Differences among Serovars. *The Scientific World Journal*, 2015(1). <https://doi.org/10.1155/2015/520179>
- Andrewes, F. W. (1922). Studies on group-agglutination. I. The *Salmonella* group and its antigenic structure. *J. Pathol. Bacteriol*, 25, 1509–1514.
- Antilles, N., García-Bocanegra, I., Alba-Casals, A., López-Soria, S., Pérez-Méndez, N., Saco, M., González-Solís, J., & Cerdà-Cuéllar, M. (2021). Occurrence and antimicrobial resistance of zoonotic enteropathogens in gulls from southern

- Europe. Science of The Total Environment, 763, 143018.
<https://doi.org/10.1016/j.scitotenv.2020.143018>
- Arvanitidou, M., Kanellou, K., & Vagiona, D. G. (2005). Diversity of *Salmonella* spp. and fungi in northern Greek rivers and their correlation to fecal pollution indicators. Environmental Research, 99(2), 278–284.
<https://doi.org/10.1016/j.envres.2005.01.002>
- Aspinall, S. T., Hindle, M. A., & Hutchinson, D. N. (1992). Improved isolation of salmonellae from faeces using a semisolid rappaport-vassiliadis medium. European Journal of Clinical Microbiology & Infectious Diseases, 11(10), 936–939. <https://doi.org/10.1007/BF01962379>
- Barak, J. D., Gorski, L., Naraghi-Arani, P., & Charkowski, A. O. (2005). *Salmonella enterica* Virulence Genes Are Required for Bacterial Attachment to Plant Tissue. Applied and Environmental Microbiology, 71(10), 5685–5691.
<https://doi.org/10.1128/AEM.71.10.5685-5691.2005>
- Barak, J. D., & Liang, A. S. (2008). Role of Soil, Crop Debris, and a Plant Pathogen in *Salmonella enterica* Contamination of Tomato Plants. PLoS ONE, 3(2), e1657.
<https://doi.org/10.1371/journal.pone.0001657>
- Barrangou, R., Fremaux, C., Deveau, H., Richards, M., Boyaval, P., Moineau, S., Romero, D. A., & Horvath, P. (2007). CRISPR Provides Acquired Resistance Against Viruses in Prokaryotes. Science, 315(5819), 1709–1712.
<https://doi.org/10.1126/science.1138140>
- Batz, M. B., Richardson, L. C., Bazaco, M. C., Parker, C. C., Chirtel, S. J., Cole, D., Golden, N. J., Griffin, P. M., Gu, W., Schmitt, S. K., Wolpert, B. J., Kufel, J. S.

- Z., & Hoekstra, R. M. (2021). Recency-Weighted Statistical Modeling Approach to Attribute Illnesses Caused by 4 Pathogens to Food Sources Using Outbreak Data, United States. *Emerging Infectious Diseases*, 27(1), 214–222.
<https://doi.org/10.3201/eid2701.203832>
- Baumler, A., & Fang, F. C. (2013). Host Specificity of Bacterial Pathogens. *Cold Spring Harbor Perspectives in Medicine*, 3(12), a010041–a010041.
<https://doi.org/10.1101/cshperspect.a010041>
- Bäumler, A. J., Tsolis, R. M., Ficht, T. A., & Adams, L. G. (1998). Evolution of Host Adaptation in *Salmonella enterica* Infection and Immunity, 66(10), 4579–4587.
<https://doi.org/10.1128/IAI.66.10.4579-4587.1998>
- Beattie, R. E., Walsh, M., Cruz, M. C., McAliley, L. R., Dodgen, L., Zheng, W., & Hristova, K. R. (2018). Agricultural contamination impacts antibiotic resistance gene abundances in river bed sediment temporally. *FEMS Microbiology Ecology*, 94(9). <https://doi.org/10.1093/femsec/fiy131>
- Bell, R. L., Zheng, J., Burrows, E., Allard, S., Wang, C. Y., Keys, C. E., Melka, D. C., Strain, E., Luo, Y., Allard, M. W., Rideout, S., & Brown, E. W. (2015). Ecological prevalence, genetic diversity, and epidemiological aspects of *Salmonella* isolated from tomato agricultural regions of the Virginia Eastern Shore. *Frontiers in Microbiology*, 6. <https://doi.org/10.3389/fmicb.2015.00415>
- Belloy, L., Decrausaz, L., Boujon, P., Hächler, H., & Waldvogel, A. S. (2009). Diagnosis by culture and PCR of *Salmonella* Abortusovis infection under clinical conditions in aborting sheep in Switzerland. *Veterinary Microbiology*, 138(3–4), 373–377.
<https://doi.org/10.1016/j.vetmic.2009.03.026>

- Benjamin, L., Atwill, E. R., Jay-Russell, M., Cooley, M., Carychao, D., Gorski, L., & Mandrell, R. E. (2013). Occurrence of generic *Escherichia coli*, *E. coli* O157 and *Salmonella* spp. in water and sediment from leafy green produce farms and streams on the Central California coast. *International Journal of Food Microbiology*, 165(1), 65–76. <https://doi.org/10.1016/j.ijfoodmicro.2013.04.003>
- Berge, A. C. B., Dueger, E. L., & Sischo, W. M. (2006). Comparison of *Salmonella enterica* serovar distribution and antibiotic resistance patterns in wastewater at municipal water treatment plants in two California cities. *Journal of Applied Microbiology*, 101(6), 1309–1316. <https://doi.org/10.1111/j.1365-2672.2006.03031.x>
- Berghaus, R. D., Thayer, S. G., Law, B. F., Mild, R. M., Hofacre, C. L., & Singer, R. S. (2013). Enumeration of *Salmonella* and *Campylobacter* spp. in Environmental Farm Samples and Processing Plant Carcass Rinses from Commercial Broiler Chicken Flocks. *Applied and Environmental Microbiology*, 79(13), 4106–4114. <https://doi.org/10.1128/AEM.00836-13>
- Bintsis, T. (2017). Foodborne pathogens. *AIMS Microbiology*, 3(3), 529–563. <https://doi.org/10.3934/microbiol.2017.3.529>
- Bisha, B., Perez-Mendez, A., Danyluk, M. D., & Goodridge, L. D. (2011). Evaluation of Modified Moore Swabs and Continuous Flow Centrifugation for Concentration of *Salmonella* and *Escherichia coli* O157:H7 from Large Volumes of Water. *Journal of Food Protection*, 74(11), 1934–1937. <https://doi.org/10.4315/0362-028X.JFP-11-133>

- Brandl, M. T., Rosenthal, B. M., Haxo, A. F., & Berk, S. G. (2005). Enhanced Survival of *Salmonella enterica* in Vesicles Released by a Soilborne Tetrahymena Species. *Applied and Environmental Microbiology*, 71(3), 1562–1569.
<https://doi.org/10.1128/AEM.71.3.1562-1569.2005>
- Brenner, F. W., Villar, R. G., Angulo, F. J., Tauxe, R., & Swaminathan, B. (2000). *Salmonella* Nomenclature. *Journal of Clinical Microbiology*, 38(7), 2465–2467.
<https://doi.org/10.1128/JCM.38.7.2465-2467.2000>
- Budu-Amoako, E., Toora, S., Ablett, R. F., & Smith, J. (1992). Evaluation of the ability of primary selective enrichment to resuscitate heat-injured and freeze-injured *Listeria monocytogenes* cells. *Applied and Environmental Microbiology*, 58(9), 3177–3179. <https://doi.org/10.1128/aem.58.9.3177-3179.1992>
- Bullis, K. L. (1977). The History of Avian Medicine in the U.S. II. Pullorum Disease and Fowl Typhoid. *Avian Diseases*, 21(3), 422. <https://doi.org/10.2307/1589326>
- Burden of disease attributable to unsafe drinking-water, sanitation and hygiene (2023). Accessed on December 17, 2024).
<https://www.who.int/publications/i/item/9789240075610>
- Burris, K. P., Simmons, O. D., Webb, H. M., Deese, L. M., Moore, R. G., Jaykus, L.-A., Zheng, J., Reed, E., Ferreira, C. M., Brown, E. W., & Bell, R. L. (2020). Colonization and Internalization of *Salmonella enterica* and Its Prevalence in Cucumber Plants. *Frontiers in Microbiology*, 11.
<https://doi.org/10.3389/fmicb.2020.01135>
- Busse, M. (1995). Chapter 11 Media for *Salmonella* (pp. 187–201).
[https://doi.org/10.1016/S0079-6352\(05\)80013-X](https://doi.org/10.1016/S0079-6352(05)80013-X)

- Byappanahalli, M. N., Sawdey, R., Ishii, S., Shively, D. A., Ferguson, J. A., Whitman, R. L., & Sadowsky, M. J. (2009). Seasonal stability of *Cladophora*-associated *Salmonella* in Lake Michigan watersheds. *Water Research*, 43(3), 806–814.
<https://doi.org/10.1016/j.watres.2008.11.012>
- Callahan, M. T., Van Kessel, J. A., & Micallef, S. A. (2019). *Salmonella enterica* recovery from river waters of the Maryland Eastern Shore reveals high serotype diversity and some multidrug resistance. *Environmental Research*, 168, 7–13.
<https://doi.org/10.1016/j.envres.2018.09.012>
- Callaway, T. R., Edrington, T. S., & Nisbet, D. J. (2014). Isolation of *Escherichia coli* O157:H7 and *Salmonella* from Migratory Brown-Headed Cowbirds (*Molothrus ater*), Common Grackles (*Quiscalus quiscula*), and Cattle Egrets (*Bubulcus ibis*). *Foodborne Pathogens and Disease*, 11(10), 791–794.
<https://doi.org/10.1089/fpd.2014.1800>
- Cardoso, M. D., Santos, A. F. D. M., Rodrigues, M. D. S., Pribul, B. R., Grael, A. S., Pedroso, V. M., Pires, J. R., Travassos, C. E. P. F., Domit, C., Vieira-Da-Motta, O., Rodrigues, D. D. P., & Siciliano, S. (2021). *Salmonella* spp. profiles isolated from seabird samples from the Brazilian coast. *Preventive Veterinary Medicine*, 193, 105413. <https://doi.org/10.1016/j.prevetmed.2021.105413>
- Carlson, J. C., Chandler, J. C., Bisha, B., LeJeune, J. T., & Wittum, T. E. (2020). Bird-livestock interactions associated with increased cattle fecal shedding of ciprofloxacin-resistant *Escherichia coli* within feedlots in the United States. *Scientific Reports*, 10(1), 10174. <https://doi.org/10.1038/s41598-020-66782-4>

- Carlson, J. C., Franklin, A. B., Hyatt, D. R., Pettit, S. E., & Linz, G. M. (2011). The role of starlings in the spread of *Salmonella* within concentrated animal feeding operations. *Journal of Applied Ecology*, 48(2), 479–486.
<https://doi.org/10.1111/j.1365-2664.2010.01935.x>
- Carroll, D., Wang, J., Fanning, S., & McMahon, B. J. (2015). Antimicrobial Resistance in Wildlife: Implications for Public Health. *Zoonoses and Public Health*, 62(7), 534–542. <https://doi.org/10.1111/zph.12182>
- Carstens, C. K., Salazar, J. K., & Darkoh, C. (2019). Multistate Outbreaks of Foodborne Illness in the United States Associated With Fresh Produce From 2010 to 2017. *Frontiers in Microbiology*, 10. <https://doi.org/10.3389/fmicb.2019.02667>
- Casanova, L. M., Hill, V. R., & Sobsey, M. D. (2020). Antibiotic-resistant *Salmonella* in swine wastes and farm surface waters. *Letters in Applied Microbiology*, 71(1), 117–123. <https://doi.org/10.1111/lam.13242>
- Cason J., Cox N., Buhr R., Bourassa D., & Richardson L. (2011). Probability of identifying different *Salmonella* serotypes in poultry samples. *International Poultry Science Forum*, 75–76.
- CDC. (2015). 2015 NARMS Human Isolates Surveillance Report. Assessed on November 15, 2024. <https://www.cdc.gov/narms/reports/annual-human-isolates-report-2015.html>
- CDC. (2024a). BEAM Dashboard. Accessed on December 16, 2024). <https://www.cdc.gov/ncezid/dfwed/BEAM-dashboard.html>

- CDC. (2024b). CDC Yellow Book: Health Information for International Travel. Oxford University Press. Accessed on January 7, 2025.
<https://www.cdc.gov/travel/page/yellowbook-home>
- CDC. (2024c). Investigation Update: *Salmonella* Outbreak, Cucumbers - June 2024. Accessed on December 16, 2024.
<https://www.cdc.gov/salmonella/outbreaks/africana-06-24/index.html>
- CDC. (2024d). Reports of Selected *Salmonella* Outbreak Investigations. Accessed on December 16, 2024. <https://www.cdc.gov/salmonella/outbreaks.html>.
- CDC. (2025). National Antimicrobial Resistance Monitoring System (NARMS) Now: Human Data. Accessed on January 16, 2025. <https://www.cdc.gov/narmsnow>.
- CDC/FDA/USDA. (2024). Interagency Food Safety Analytics Collaboration. Foodborne illness source attribution estimates for 2022 for *Salmonella*, *Escherichia coli* O157, and *Listeria monocytogenes* using multi-year outbreak surveillance data. Accessed January 14, 2025. <https://www.cdc.gov/ifsac/php/data-research/annual-report-2022.html>
- Centers for Disease Control and Prevention. (2022). FoodNet Fast. Pathogen Surveillance. Accessed January 17, 2025. <https://wwwn.cdc.gov/foodnetfast/>
- Centers for Disease Control and Prevention. (2023). *Salmonella*: Questions and Answers. Accessed November 16, 2024.
<https://www.cdc.gov/salmonella/general/index.html>
- Chen, Z., Moreno-Switt, A. I., Reyes-Jara, A., Delgado Suarez, E., Adell, A. D., Oliveira, C. J. B., Bonelli, R. R., Huang, X., Brown, E., Allard, M., Grim, C., Bell, R., Meng, J., & Toro, M. (2024). A multicenter genomic epidemiological

- investigation in Brazil, Chile, and Mexico reveals the diversity and persistence of *Salmonella* populations in surface waters. *MBio*, 15(7), e0077724.
<https://doi.org/10.1128/mbio.00777-24>
- Cheng, R. A., Eade, C. R., & Wiedmann, M. (2019). Embracing Diversity: Differences in Virulence Mechanisms, Disease Severity, and Host Adaptations Contribute to the Success of Nontyphoidal *Salmonella* as a Foodborne Pathogen. *Frontiers in Microbiology*, 10, 1368. <https://doi.org/10.3389/fmicb.2019.01368>
- Chiu, C.-H., Wu, T.-L., Su, L.-H., Chu, C., Chia, J.-H., Kuo, A.-J., Chien, M.-S., & Lin, T.-Y. (2002). The Emergence in Taiwan of Fluoroquinolone Resistance in *Salmonella enterica* Serotype Choleraesuis. *New England Journal of Medicine*, 346(6), 413–419. <https://doi.org/10.1056/NEJMoa012261>
- Cho, S., Hiott, L. M., House, S. L., Woodley, T. A., McMillan, E. A., Sharma, P., Barrett, J. B., Adams, E. S., Brandenburg, J. M., Hise, K. B., Bateman McDonald, J. M., Ottesen, E. A., Lipp, E. K., Jackson, C. R., & Frye, J. G. (2022). Analysis of *Salmonella enterica* Isolated from a Mixed-Use Watershed in Georgia, USA: Antimicrobial Resistance, Serotype Diversity, and Genetic Relatedness to Human Isolates. *Applied and Environmental Microbiology*, 88(10).
<https://doi.org/10.1128/aem.00393-22>
- Cho, S., Jackson, C. R., & Frye, J. G. (2020). The prevalence and antimicrobial resistance phenotypes of *Salmonella*, *Escherichia coli* and *Enterococcus* sp. in surface water. *Letters in Applied Microbiology*, 71(1), 3–25.
<https://doi.org/10.1111/lam.13301>

- Cohen, E., Azriel, S., Auster, O., Gal, A., Mikhlin, S., Crauwels, S., Rahav, G., & Gal-Mor, O. (2024). A new *Salmonella enterica* serovar that was isolated from a wild sparrow presents a distinct genetic, metabolic and virulence profile. *Microbes and Infection*, 26(3), 105249. <https://doi.org/10.1016/j.micinf.2023.105249>
- Cohen, E., Azriel, S., Auster, O., Gal, A., Zitronblat, C., Mikhlin, S., Scharte, F., Hensel, M., Rahav, G., & Gal-Mor, O. (2021). Pathoadaptation of the passerine-associated *Salmonella enterica* serovar Typhimurium lineage to the avian host. *PLOS Pathogens*, 17(3), e1009451. <https://doi.org/10.1371/journal.ppat.1009451>
- Cooley, M. B., Quiñones, B., Oryang, D., Mandrell, R. E., & Gorski, L. (2014). Prevalence of shiga toxin producing *Escherichia coli*, *Salmonella enterica*, and *Listeria monocytogenes* at public access watershed sites in a California Central Coast agricultural region. *Frontiers in Cellular and Infection Microbiology*, 4. <https://doi.org/10.3389/fcimb.2014.00030>
- Cooley, M., Carychao, D., Crawford-Miksza, L., Jay, M. T., Myers, C., Rose, C., Keys, C., Farrar, J., & Mandrell, R. E. (2007). Incidence and Tracking of *Escherichia coli* O157:H7 in a Major Produce Production Region in California. *PLoS ONE*, 2(11), e1159. <https://doi.org/10.1371/journal.pone.0001159>
- Crump, J. A., Luby, S. P., & Mintz, E. D. (2004). The global burden of typhoid fever. *Bulletin of the World Health Organization*, 82(5), 346–353.
- D'aoust, J.-Y. (1981). Update on Preenrichment and Selective Enrichment Conditions for Detection of *Salmonella* in Foods. *Journal of Food Protection*, 44(5), 369–375. <https://doi.org/10.4315/0362-028X-44.5.369>

- Deaven, A. M., Ferreira, C. M., Reed, E. A., Chen See, J. R., Lee, N. A., Almaraz, E., Rios, P. C., Marogi, J. G., Lamendella, R., Zheng, J., Bell, R. L., & Shariat, N. W. (2021). *Salmonella* Genomics and Population Analyses Reveal High Inter- and Intraserovar Diversity in Freshwater. *Applied and Environmental Microbiology*, 87(6). <https://doi.org/10.1128/AEM.02594-20>
- Delannoy, S., Beutin, L., & Fach, P. (2012). Use of Clustered Regularly Interspaced Short Palindromic Repeat Sequence Polymorphisms for Specific Detection of Enterohemorrhagic *Escherichia coli* Strains of Serotypes O26:H11, O45:H2, O103:H2, O111:H8, O121:H19, O145:H28, and O157:H7 by Real-Time PCR. *Journal of Clinical Microbiology*, 50(12), 4035–4040. <https://doi.org/10.1128/JCM.02097-12>
- Demczuk, W., Soule, G., Clark, C., Ackermann, H.-W., Easy, R., Khakhria, R., Rodgers, F., & Ahmed, R. (2003). Phage-Based Typing Scheme for *Salmonella enterica* Serovar Heidelberg, a Causative Agent of Food Poisonings in Canada. *Journal of Clinical Microbiology*, 41(9), 4279–4284. <https://doi.org/10.1128/JCM.41.9.4279-4284.2003>
- den Bakker, H. C., Allard, M. W., Bopp, D., Brown, E. W., Fontana, J., Iqbal, Z., Kinney, A., Limberger, R., Musser, K. A., Shudt, M., Strain, E., Wiedmann, M., & Wolfgang, W. J. (2014). Rapid Whole-Genome Sequencing for Surveillance of *Salmonella enterica* Serovar Enteritidis. *Emerging Infectious Diseases*, 20(8), 1306–1314. <https://doi.org/10.3201/eid2008.131399>
- Deng, X., Shariat, N., Driebe, E. M., Roe, C. C., Tolar, B., Trees, E., Keim, P., Zhang, W., Dudley, E. G., Fields, P. I., & Engelthaler, D. M. (2015). Comparative

- Analysis of Subtyping Methods against a Whole-Genome-Sequencing Standard for *Salmonella enterica* Serotype Enteritidis. *Journal of Clinical Microbiology*, 53(1), 212–218. <https://doi.org/10.1128/JCM.02332-14>
- Devarajan, N., Weller, D. L., Jones, M., Adell, A. D., Adhikari, A., Allende, A., Arnold, N. L., Baur, P., Beno, S. M., Clements, D., Olimpí, E. M., Critzer, F., Green, H., Gorski, L., Ferelli Gruber, A., Kovac, J., McGarvey, J., Murphy, C. M., Murphy, S. I., ... Karp, D. S. (2023). Evidence for the efficacy of pre-harvest agricultural practices in mitigating food-safety risks to fresh produce in North America. *Frontiers in Sustainable Food Systems*, 7, 1101435. <https://doi.org/10.3389/fsufs.2023.1101435>
- Dias-Alves, A., Espunyes, J., Ayats, T., Sente, C., Sebulime, P., Muro, J., Tushabe, J., Asiimwe, C., Fernandez Aguilar, X., Aruho, R., Marco, I., Planellas, M., Cardells, J., Cabezón, O., & Cerdà-Cuéllar, M. (2023). Foodborne Pathogens at the Livestock–Wildlife–Human Interface in Rural Western Uganda. *EcoHealth*, 20(2), 144–149. <https://doi.org/10.1007/s10393-023-01639-6>
- Díaz-Torres, O., Lugo-Melchor, O. Y., de Anda, J., Gradilla-Hernández, M. S., Amézquita-López, B. A., & Meza-Rodríguez, D. (2020). Prevalence, Distribution, and Diversity of *Salmonella* Strains Isolated From a Subtropical Lake. *Frontiers in Microbiology*, 11. <https://doi.org/10.3389/fmicb.2020.521146>
- Doden, G., Gartlan, B., Klein, K., Maddox, C. W., Adamovicz, L. A., & Allender, M. C. (2021). Prevalence And Antimicrobial Resistance Patterns Of *Salmonella* Spp. In Two Free-Ranging Populations Of Eastern Box Turtles (*Terrapene Carolina*

- Carolina*). Journal of Zoo and Wildlife Medicine, 52(3).
<https://doi.org/10.1638/2020-0061>
- Dolejská, M., Bierošová, B., Kohoutová, L., Literák, I., & Čížek, A. (2009). Antibiotic-resistant *Salmonella* and *Escherichia coli* isolates with integrons and extended-spectrum beta-lactamases in surface water and sympatric black-headed gulls. Journal of Applied Microbiology, 106(6), 1941–1950.
<https://doi.org/10.1111/j.1365-2672.2009.04155.x>
- Dunbar, S. A., Vander Zee, C. A., Oliver, K. G., Karem, K. L., & Jacobson, J. W. (2003). Quantitative, multiplexed detection of bacterial pathogens: DNA and protein applications of the Luminex LabMAP™ system. Journal of Microbiological Methods, 53(2), 245–252. [https://doi.org/10.1016/S0167-7012\(03\)00028-9](https://doi.org/10.1016/S0167-7012(03)00028-9)
- Editorial team, C., Bertrand, S., Rimhanen-Finne, R., Weill, F. X., Rabsch, W., Thornton, L., Perevoščikovs, J., van Pelt, W., & Heck, M. (2008). *Salmonella* infections associated with reptiles: the current situation in Europe. Eurosurveillance, 13(24).
<https://doi.org/10.2807/ese.13.24.18902-en>
- Elsohaby, I., Samy, A., Elmoslemany, A., Alorabi, M., Alkafafy, M., Aldoweriej, A., Al-Marri, T., Elbehiry, A., & Fayez, M. (2021). Migratory Wild Birds as a Potential Disseminator of Antimicrobial-Resistant Bacteria around Al-Asfar Lake, Eastern Saudi Arabia. Antibiotics, 10(3), 260. <https://doi.org/10.3390/antibiotics10030260>
- Endo, T., Ikeo, K., & Gojobori, T. (1996). Large-scale search for genes on which positive selection may operate. Molecular Biology and Evolution, 13(5), 685–690.
<https://doi.org/10.1093/oxfordjournals.molbev.a025629>

- Entis, P. (1990). Improved Hydrophobic Grid Membrane Filter Method, Using EF-18 Agar, for Detection of *Salmonella* in Foods: Collaborative Study. *Journal of AOAC International*, 73(5), 734–742. <https://doi.org/10.1093/jaoac/73.5.734>
- Ernst, R. K., Dombroski, D. M., & Merrick, J. M. (1990). Anaerobiosis, type 1 fimbriae, and growth phase are factors that affect invasion of HEp-2 cells by *Salmonella* Typhimurium. *Infection and Immunity*, 58(6), 2014–2016. <https://doi.org/10.1128/iai.58.6.2014-2016.1990>
- Ernst, R. K., Guina, T., & Miller, S. I. (2001). outer membrane remodeling: role in resistance to host innate immunity. *Microbes and Infection*, 3(14–15), 1327–1334. [https://doi.org/10.1016/S1286-4579\(01\)01494-0](https://doi.org/10.1016/S1286-4579(01)01494-0)
- Espunyes, J., Cabezón, O., Dias-Alves, A., Miralles, P., Ayats, T., & Cerdà-Cuéllar, M. (2021). Assessing the role of livestock and sympatric wild ruminants in spreading antimicrobial resistant *Campylobacter* and *Salmonella* in alpine ecosystems. *BMC Veterinary Research*, 17(1), 79. <https://doi.org/10.1186/s12917-021-02784-2>
- Esteban, E., Snipes, K., Hird, D., Kasten, R., & Kinde, H. (1993). Use of ribotyping for characterization of *Salmonella* serotypes. *Journal of Clinical Microbiology*, 31(2), 233–237. <https://doi.org/10.1128/jcm.31.2.233-237.1993>
- Estrada-Acosta, M. D., Ramirez, K., Medrano-Félix, J. A., Castro-Del Campo, N., López-Moreno, H. S., Jimenez Edeza, M., Martínez-Urtaza, J., & Chaidez, C. (2018). Effect of river water exposition on adhesion and invasion abilities of *Salmonella* Oranienburg and Saintpaul. *International Journal of Environmental Health Research*, 28(1), 43–54. <https://doi.org/10.1080/09603123.2017.1415308>

- Ewing, W., & Edwards, P. (1986). *Edwards and Ewing's Identification of Enterobacteriaceae* (4th ed.).
- Fabre, L., Zhang, J., Guigon, G., Le Hello, S., Guibert, V., Accou-Demartin, M., de Romans, S., Lim, C., Roux, C., Passet, V., Diancourt, L., Guibourdenche, M., Issenhuth-Jeanjean, S., Achtman, M., Brisse, S., Sola, C., & Weill, F.-X. (2012). CRISPR Typing and Subtyping for Improved Laboratory Surveillance of *Salmonella* Infections. *PLoS ONE*, 7(5), e36995.
<https://doi.org/10.1371/journal.pone.0036995>
- FDA. (2011). FSMA Final Rule on Produce Safety. Accessed November 17, 2024.
<https://www.fda.gov/food/food-safety-modernization-act-fsma/fsma-final-rule-produce-safety>.
- FDA. (2015). Veterinary Feed Directive (VFD). Accessed November 17, 2024.
<https://www.fda.gov/animal-veterinary/development-approval-process/veterinary-feed-directive-vfd>
- FDA. (2021). Factors Potentially Contributing to the Contamination of Red Onions Implicated in the Summer 2020 Outbreak of *Salmonella* Newport. Accessed November 18, 2024. <https://www.fda.gov/food/outbreaks-foodborne-illness/factors-potentially-contributing-contamination-red-onions-implicated-summer-2020-outbreak-salmonella>.
- FDA. (2023a). Factors Potentially Contributing to the Contamination of Cantaloupe Implicated in the Outbreak of *Salmonella* Typhimurium During the Summer of 2022. Accessed December 4, 2024. <https://www.fda.gov/food/outbreaks->

foodborne-illness/factors-potentially-contributing-contamination-cantaloupe-implicated-outbreak-salmonella-typhimurium

FDA. (2023b). FDA Issues Report Highlighting *Salmonella* Outbreak in Cantaloupe During Summer of 2022. Accessed December 18, 2024.

<https://www.fda.gov/food/hfp-constituent-updates/fda-issues-report-highlighting-salmonella-outbreak-cantaloupe-during-summer-2022>.

FDA. (2024a). Bacterial Analytical Manual (BAM) Chapter 5: *Salmonella*. Accessed December 8, 2024. <https://www.fda.gov/food/laboratory-methods-food/bam-chapter-5-salmonella>

FDA. (2024b). Investigations Operations Manual. Accessed January 7, 2025.

<https://www.fda.gov/inspections-compliance-enforcement-and-criminal-investigations/inspection-references/investigations-operations-manual>

Figueira, R., & Holden, D. W. (2012). Functions of the *Salmonella* pathogenicity island 2 (SPI-2) type III secretion system effectors. *Microbiology*, 158(5), 1147–1161.

<https://doi.org/10.1099/mic.0.058115-0>

Fitzgerald, C., Collins, M., van Duyne, S., Mikoleit, M., Brown, T., & Fields, P. (2007).

Multiplex, Bead-Based Suspension Array for Molecular Determination of Common *Salmonella* Serogroups. *Journal of Clinical Microbiology*, 45(10), 3323–3334. <https://doi.org/10.1128/JCM.00025-07>

Fitzgerald, C., Sherwood, R., Gheesling, L. L., Brenner, F. W., & Fields, P. I. (2003).

Molecular Analysis of the rfb O Antigen Gene Cluster of *Salmonella enterica* Serogroup O:6,14 and Development of a Serogroup-Specific PCR Assay. *Applied*

and Environmental Microbiology, 69(10), 6099–6105.

<https://doi.org/10.1128/AEM.69.10.6099-6105.2003>

Ford, L., Ellison, Z., Schwensohn, C., Griffin, I., Birhane, M. G., Cote, A., Fortenberry, G. Z., Tecle, S., Higa, J., Spencer, S., Patton, B., Patel, J., Dow, J., Maroufi, A., Robbins, A., Donovan, D., Fitzgerald, C., Burrell, S., Tolar, B., ... Francois Watkins, L. K. (2023). Strain of Multidrug-Resistant *Salmonella* Newport Remains Linked to Travel to Mexico and U.S. Beef Products — United States, 2021–2022. MMWR. Morbidity and Mortality Weekly Report, 72(45), 1225–1229. <https://doi.org/10.15585/mmwr.mm7245a3>

Franklin, A. B., & VerCauteren, K. C. (2016). Keeping Wildlife Out of Your Food: Mitigation and Control Strategies to Reduce the Transmission Risk of Food-Borne Pathogens.

Friend, M., Wild, M., & Gibbs, S. (2015). Field Manual of Wildlife Diseases. <https://doi.org/10.3133/tm15>

Fu, Y., Smith, J. C., Shariat, N. W., M'ikanatha, N. M., & Dudley, E. G. (2022). Evidence for common ancestry and microevolution of passerine-adapted *Salmonella enterica* serovar Typhimurium in the UK and USA. Microbial Genomics, 8(2). <https://doi.org/10.1099/mgen.0.000775>

Fukui, D., Takahashi, K., Kubo, M., Une, Y., Kato, Y., Izumiya, H., Teraoka, H., Asakawa, M., Yanagida, K., & Bando, G. (2014). Mass Mortality Of Eurasian Tree Sparrows (*Passer Montanus*) From *Salmonella* Typhimurium DT40 In Japan, Winter 2008–09. Journal of Wildlife Diseases, 50(3), 484–495. <https://doi.org/10.7589/2012-12-321>

- Gaertner, J. P., Mendoza, J. A., Forstner, M. R. J., & Hahn, D. (2011). Recovery of *Salmonella* from biofilms in a headwater spring ecosystem. *Journal of Water and Health*, 9(3), 458–466. <https://doi.org/10.2166/wh.2011.173>
- Gal-Mor, O., Boyle, E. C., & Grassl, G. A. (2014). Same species, different diseases: how and why typhoidal and non-typhoidal *Salmonella enterica* serovars differ. *Frontiers in Microbiology*, 5. <https://doi.org/10.3389/fmicb.2014.00391>
- Gal-Mor, O., Valinsky, L., Weinberger, M., Guy, S., Jaffe, J., Schorr, Y. I., Raisfeld, A., Agmon, V., & Nissan, I. (2010). Multidrug-Resistant *Salmonella enterica* Serovar Infantis, Israel. *Emerging Infectious Diseases*, 16(11), 1754–1757. <https://doi.org/10.3201/eid1611.100100>
- Gamble, G. R., Berrang, M. E., Buhr, R. J., Hinton, A., Bourassa, D. V., Ingram, K. D., Adams, E. S., Feldner, P. W., & Johnston, J. J. (2017). Neutralization of Bactericidal Activity Related to Antimicrobial Carryover in Broiler Carcass Rinse Samples. *Journal of Food Protection*, 80(4), 685–691. <https://doi.org/10.4315/0362-028X.JFP-16-412>
- Gardner, T. J., Fitzgerald, C., Xavier, C., Klein, R., Pruckler, J., Stroika, S., & McLaughlin, J. B. (2011). Outbreak of Campylobacteriosis Associated With Consumption of Raw Peas. *Clinical Infectious Diseases*, 53(1), 26–32. <https://doi.org/10.1093/cid/cir249>
- Garneau, J. E., Dupuis, M.-È., Villion, M., Romero, D. A., Barrangou, R., Boyaval, P., Fremaux, C., Horvath, P., Magadán, A. H., & Moineau, S. (2010). The CRISPR/Cas bacterial immune system cleaves bacteriophage and plasmid DNA. *Nature*, 468(7320), 67–71. <https://doi.org/10.1038/nature09523>

- Gibert, I., Barbe, J., & Casadesus, J. (1990). Distribution of insertion sequence IS200 in *Salmonella* and *Shigella*. *Journal of General Microbiology*, 136(12), 2555–2560. <https://doi.org/10.1099/00221287-136-12-2555>
- Gibson, A. M., Bratchell, N., & Roberts, T. A. (1988). Predicting microbial growth: growth responses of salmonellae in a laboratory medium as affected by pH, sodium chloride and storage temperature. *International Journal of Food Microbiology*, 6(2), 155–178. [https://doi.org/10.1016/0168-1605\(88\)90051-7](https://doi.org/10.1016/0168-1605(88)90051-7)
- González-López, I., Medrano-Félix, J. A., Castro-del Campo, N., López-Cuevas, O., González-Gómez, J. P., Valdez-Torres, J. B., Aguirre-Sánchez, J. R., Martínez-Urtaza, J., Gómez-Gil, B., Lee, B. G., Quiñones, B., & Chaidez, C. (2022). Prevalence and Genomic Diversity of *Salmonella enterica* Recovered from River Water in a Major Agricultural Region in Northwestern Mexico. *Microorganisms*, 10(6), 1214. <https://doi.org/10.3390/microorganisms10061214>
- Gordus, A. G. , R. E. M. and E. R. A. (2011). Wildlife Survey for *E. coli* O157:H7 and *Salmonella* in the Central Coastal Counties of California.
- Gorski, L. (2012). Selective Enrichment Media Bias the Types of *Salmonella enterica* Strains Isolated from Mixed Strain Cultures and Complex Enrichment Broths. *PLoS ONE*, 7(4), e34722. <https://doi.org/10.1371/journal.pone.0034722>
- Gorski, L., Liang, A. S., Walker, S., Carychao, D., Aviles Noriega, A., Mandrell, R. E., & Cooley, M. B. (2022). *Salmonella enterica* Serovar Diversity, Distribution, and Prevalence in Public-Access Waters from a Central California Coastal Leafy Green-Growing Region from 2011 to 2016. *Applied and Environmental Microbiology*, 88(3). <https://doi.org/10.1128/aem.01834-21>

- Gorski, L., & Noriega, A. A. (2023). Comparison of Phenotype Nutritional Profiles and Phosphate Metabolism Genes in Four Serovars of *Salmonella enterica* from Water Sources. *Microorganisms*, 11(8), 2109.
<https://doi.org/10.3390/microorganisms11082109>
- Gorski, L., Parker, C. T., Liang, A., Cooley, M. B., Jay-Russell, M. T., Gordus, A. G., Atwill, E. R., & Mandrell, R. E. (2011). Prevalence, Distribution, and Diversity of *Salmonella enterica* in a Major Produce Region of California. *Applied and Environmental Microbiology*, 77(8), 2734–2748.
<https://doi.org/10.1128/AEM.02321-10>
- Goyal, S. M., Gerba, C. P., & Melnick, J. L. (1977). Occurrence and distribution of bacterial indicators and pathogens in canal communities along the Texas coast. *Applied and Environmental Microbiology*, 34(2), 139–149.
<https://doi.org/10.1128/aem.34.2.139-149.1977>
- Gray, J. T., Fedorka-Cray, P. J., Stabel, T. J., & Kramer, T. T. (1996). Natural transmission of *Salmonella choleraesuis* in swine. *Applied and Environmental Microbiology*, 62(1), 141–146. <https://doi.org/10.1128/aem.62.1.141-146.1996>
- Grimont, P. A. D., & Weill, F.-X. (2007). WHO Collaborating Centre for Reference and Research on *Salmonella* Antigenic Formulae Of The *Salmonella* Serovars 2007 9th edition.
- Grissa, I., Vergnaud, G., & Pourcel, C. (2007). The CRISPRdb database and tools to display CRISPRs and to generate dictionaries of spacers and repeats. *BMC Bioinformatics*, 8(1), 172. <https://doi.org/10.1186/1471-2105-8-172>

- Grivokostopoulos, N. C., Makariti, I. P., Hilaj, N., Apostolidou, Z., & Skandamis, P. N. (2022). Internalization of *Salmonella* in Leafy Greens and Impact on Acid Tolerance. *Applied and Environmental Microbiology*, 88(6).
<https://doi.org/10.1128/aem.02249-21>
- Gruszynski, K., Pao, S., Kim, C., Toney, D., Wright, K., Ross, P. G., Colon, A., & Levine, S. (2014). Evaluating Wildlife as a Potential Source of *Salmonella* serotype Newport (JJPX01.0061) Contamination for Tomatoes on the Eastern Shore of Virginia. *Zoonoses and Public Health*, 61(3), 202–207.
<https://doi.org/10.1111/zph.12061>
- Gu, G., Strawn, L. K., Ottesen, A. R., Ramachandran, P., Reed, E. A., Zheng, J., Boyer, R. R., & Rideout, S. L. (2021). Correlation of *Salmonella enterica* and *Listeria monocytogenes* in Irrigation Water to Environmental Factors, Fecal Indicators, and Bacterial Communities. *Frontiers in Microbiology*, 11.
<https://doi.org/10.3389/fmicb.2020.557289>
- Gu, G., Strawn, L. K., Zheng, J., Reed, E. A., & Rideout, S. L. (2019). Diversity and Dynamics of *Salmonella enterica* in Water Sources, Poultry Litters, and Field Soils Amended With Poultry Litter in a Major Agricultural Area of Virginia. *Frontiers in Microbiology*, 10. <https://doi.org/10.3389/fmicb.2019.02868>
- Hadjinicolaou, A. V., Demetriou, V. L., Emmanuel, M. A., Kakoyiannis, C. K., & Kostrikis, L. G. (2009). Molecular beacon-based real-time PCR detection of primary isolates of *Salmonella* Typhimurium and *Salmonella* Enteritidis in environmental and clinical samples. *BMC Microbiology*, 9(1), 97.
<https://doi.org/10.1186/1471-2180-9-97>

- Hajna, A. A., & Damon, S. R. (1956). New Enrichment and Plating Media for the Isolation of *Salmonella* and Shigella Organisms. *Applied Microbiology*, 6, 341–345.
- Haley, B. J., Cole, D. J., & Lipp, E. K. (2009). Distribution, Diversity, and Seasonality of Waterborne Salmonellae in a Rural Watershed. *Applied and Environmental Microbiology*, 75(5), 1248–1255. <https://doi.org/10.1128/AEM.01648-08>
- Hall, A. J., & Saito, E. K. (2008). Avian Wildlife Mortality Events Due To Salmonellosis In The United States, 1985–2004. *Journal of Wildlife Diseases*, 44(3), 585–593. <https://doi.org/10.7589/0090-3558-44.3.585>
- Hammack, T. S., Amaguaña, R. M., June, G. A., Sherrod, P. S., & Andrews, W. H. (1999). Relative Effectiveness of Selenite Cystine Broth, Tetrathionate Broth, and Rappaport-Vassiliadis Medium for the Recovery of *Salmonella* spp. from Foods with a Low Microbial Load. *Journal of Food Protection*, 62(1), 16–21. <https://doi.org/10.4315/0362-028X-62.1.16>
- Havelaar, A. H., Vazquez, K. M., Topalcengiz, Z., Muñoz-Carpena, R., & DANYLUK, M. D. (2017). Evaluating the U.S. Food Safety Modernization Act Produce Safety Rule Standard for Microbial Quality of Agricultural Water for Growing Produce. *Journal of Food Protection*, 80(11), 1832–1841. <https://doi.org/10.4315/0362-028X.JFP-17-122>
- Hendriksen, R. S., Vieira, A. R., Karlsmose, S., Lo Fo Wong, D. M. A., Jensen, A. B., Wegener, H. C., & Aarestrup, F. M. (2011). Global Monitoring of *Salmonella* Serovar Distribution from the World Health Organization Global Foodborne Infections Network Country Data Bank: Results of Quality Assured Laboratories

- from 2001 to 2007. *Foodborne Pathogens and Disease*, 8(8), 887–900.
<https://doi.org/10.1089/fpd.2010.0787>
- Hensel, M. (2000). *Salmonella* Pathogenicity Island 2. *Molecular Microbiology*, 36(5), 1015–1023. <https://doi.org/10.1046/j.1365-2958.2000.01935.x>
- Hernandez, S. M., Keel, K., Sanchez, S., Trees, E., Gerner-Smidt, P., Adams, J. K., Cheng, Y., Ray, A., Martin, G., Presotto, A., Ruder, M. G., Brown, J., Blehert, D. S., Cottrell, W., & Maurer, J. J. (2012). Epidemiology of a *Salmonella enterica* subsp. *enterica* Serovar Typhimurium Strain Associated with a Songbird Outbreak. *Applied and Environmental Microbiology*, 78(20), 7290–7298.
<https://doi.org/10.1128/AEM.01408-12>
- Hernandez, S. M., Welch, C. N., Peters, V. E., Lipp, E. K., Curry, S., Yabsley, M. J., Sanchez, S., Presotto, A., Gerner-Smidt, P., Hise, K. B., Hammond, E., Kistler, W. M., Madden, M., Conway, A. L., Kwan, T., & Maurer, J. J. (2016). Urbanized White Ibises (*Eudocimus albus*) as Carriers of *Salmonella enterica* of Significance to Public Health and Wildlife. *PLOS ONE*, 11(10), e0164402.
<https://doi.org/10.1371/journal.pone.0164402>
- Hintz, L. D., Boyer, R. R., Ponder, M. A., Williams, R. C., & Rideout, S. L. (2010). Recovery of *Salmonella enterica* Newport Introduced through Irrigation Water from Tomato (*Lycopersicum esculentum*) Fruit, Roots, Stems, and Leaves. *HortScience*, 45(4), 675–678. <https://doi.org/10.21273/HORTSCI.45.4.675>
- Ho, Y.-N., Tsai, H.-C., Hsu, B.-M., & Chiou, C.-S. (2018). The association of *Salmonella enterica* from aquatic environmental and clinical samples in Taiwan. *Science of*

The Total Environment, 624, 106–113.

<https://doi.org/10.1016/j.scitotenv.2017.12.122>

Hoe, N., Nakashima, K., Grigsby, D., Pan, X., Dou, S. J., Naidich, S., Garcia, M., Kahn, E., Bergmire-Sweat, D., & Musser, J. M. (1999). Rapid Molecular Genetic Subtyping of Serotype M1 Group A *Streptococcus* Strains. *Emerging Infectious Diseases*, 5(2), 254–263. <https://doi.org/10.3201/eid0502.990210>

Hoffmann, M., Luo, Y., Monday, S. R., Gonzalez-Escalona, N., Ottesen, A. R., Muruvanda, T., Wang, C., Kastanis, G., Keys, C., Janies, D., Senturk, I. F., Catalyurek, U. V., Wang, H., Hammack, T. S., Wolfgang, W. J., Schoonmaker-Bopp, D., Chu, A., Myers, R., Haendiges, J., ... Brown, E. W. (2016). Tracing Origins of the *Salmonella* Bareilly Strain Causing a Food-borne Outbreak in the United States. *Journal of Infectious Diseases*, 213(4), 502–508. <https://doi.org/10.1093/infdis/jiv297>

Hoffmann, S., Batz, M. B., & Morris, J. G. (2012). Annual Cost of Illness and Quality-Adjusted Life Year Losses in the United States Due to 14 Foodborne Pathogens. *Journal of Food Protection*, 75(7), 1292–1302. <https://doi.org/10.4315/0362-028X.JFP-11-417>

Huang, K.-H., Hsu, B.-M., Chou, M.-Y., Tsai, H.-L., Kao, P.-M., Wang, H.-J., Hsiao, H.-Y., Su, M.-J., & Huang, Y.-L. (2014). Application of molecular biological techniques to analyze *Salmonella* seasonal distribution in stream water. *FEMS Microbiology Letters*, 352(1), 87–96. <https://doi.org/10.1111/1574-6968.12381>

Hutchison, M. L., Walters, L. D., Moore, A., Crookes, K. M., & Avery, S. M. (2004). Effect of Length of Time before Incorporation on Survival of Pathogenic Bacteria

- Present in Livestock Wastes Applied to Agricultural Soil. *Applied and Environmental Microbiology*, 70(9), 5111–5118.
- <https://doi.org/10.1128/AEM.70.9.5111-5118.2004>
- IFSAC. (2024). Foodborne Illness Source Attribution Estimates – United States, 2022. Accessed January 7, 2025. <https://www.cdc.gov/ifsac/php/data-research/annual-report-2022.html>.
- IHME. (2021a). Paratyphoid fever—Level 4 cause. Accessed December 7, 2024. <https://www.healthdata.org/research-analysis/diseases-injuries-risks/factsheets/2021-paratyphoid-fever-level-4-disease>.
- IHME. (2021b). Typhoid fever—Level 4 cause. Accessed December 7, 2024. <https://www.healthdata.org/research-analysis/diseases-injuries-risks/factsheets/2021-typhoid-fever-level-4-disease>.
- Islam, M., Doyle, M. P., Phatak, S. C., Millner, P., & Jiang, X. (2005). Survival of *Escherichia coli* O157:H7 in soil and on carrots and onions grown in fields treated with contaminated manure composts or irrigation water. *Food Microbiology*, 22(1), 63–70. <https://doi.org/10.1016/j.fm.2004.04.007>
- Issenhuth-Jeanjean, S., Roggentin, P., Mikoleit, M., Guibourdenche, M., de Pinna, E., Nair, S., Fields, P. I., & Weill, F.-X. (2014). Supplement 2008–2010 (no. 48) to the White–Kauffmann–Le Minor scheme. *Research in Microbiology*, 165(7), 526–530. <https://doi.org/10.1016/j.resmic.2014.07.004>
- Janecko, N., Čížek, A., Halová, D., Karpíšková, R., Myšková, P., & Literák, I. (2015). Prevalence, Characterization and Antibiotic Resistance of *Salmonella* Isolates in

- Large Corvid Species of Europe and North America Between 2010 and 2013. *Zoonoses and Public Health*, 62(4), 292–300. <https://doi.org/10.1111/zph.12149>
- Jay-Russell, M. T., & Justice-Allen, A. (2014). Reducing the Risk for Transfer of Zoonotic Foodborne Pathogens from Domestic and Wild Animals to Vegetable Crops in the Southwest Desert.
- Jechalke, S., Schierstaedt, J., Becker, M., Flemer, B., Grosch, R., Smalla, K., & Schikora, A. (2019). *Salmonella* Establishment in Agricultural Soil and Colonization of Crop Plants Depend on Soil Type and Plant Species. *Frontiers in Microbiology*, 10. <https://doi.org/10.3389/fmicb.2019.00967>
- Johannessen, G., Wennberg, A., Nesheim, I., & Tryland, I. (2015). Diverse Land Use and the Impact on (Irrigation) Water Quality and Need for Measures — A Case Study of a Norwegian River. *International Journal of Environmental Research and Public Health*, 12(6), 6979–7001. <https://doi.org/10.3390/ijerph120606979>
- Jokinen, C. C., Koot, J., Cole, L., Desruisseau, A., Edge, T. A., Khan, I. U. H., Koning, W., Lapen, D. R., Pintar, K. D. M., Reid-Smith, R., Thomas, J. L., Topp, E., Wang, L. Y., Wilkes, G., Ziebell, K., van Bochove, E., & Gannon, V. P. J. (2015). The distribution of *Salmonella enterica* serovars and subtypes in surface water from five agricultural regions across Canada. *Water Research*, 76, 120–131. <https://doi.org/10.1016/j.watres.2015.02.038>
- Jung, Y., Jang, H., & Matthews, K. R. (2014). Effect of the food production chain from farm practices to vegetable processing on outbreak incidence. *Microbial Biotechnology*, 7(6), 517–527. <https://doi.org/10.1111/1751-7915.12178>

- Kabir, S. M. L. (2010). Avian Colibacillosis and Salmonellosis: A Closer Look at Epidemiology, Pathogenesis, Diagnosis, Control and Public Health Concerns. *International Journal of Environmental Research and Public Health*, 7(1), 89–114. <https://doi.org/10.3390/ijerph7010089>
- Kamerbeek, J., Schouls, L., Kolk, A., van Agterveld, M., van Soolingen, D., Kuijper, S., Bunschoten, A., Molhuizen, H., Shaw, R., Goyal, M., & van Embden, J. (1997). Simultaneous detection and strain differentiation of *Mycobacterium tuberculosis* for diagnosis and epidemiology. *Journal of Clinical Microbiology*, 35(4), 907–914. <https://doi.org/10.1128/jcm.35.4.907-914.1997>
- Karp, D. S., Gennet, S., Kilonzo, C., Partyka, M., Chaumont, N., Atwill, E. R., & Kremen, C. (2015). Comanaging fresh produce for nature conservation and food safety. *Proceedings of the National Academy of Sciences*, 112(35), 11126–11131. <https://doi.org/10.1073/pnas.1508435112>
- Kauffmann, F. (1964). Das Kaufmann-White schema: The world problem of salmonellosis.
- Keenleyside, W. J., Perry, M., Maclean, L., Poppe, C., & Whitfield, C. (1994). A plasmid-encoded rfb O:54 gene cluster is required for biosynthesis of the O:54 antigen in *Salmonella enterica* serovar Borreze. *Molecular Microbiology*, 11(3), 437–448. <https://doi.org/10.1111/j.1365-2958.1994.tb00325.x>
- Khouja, B. A., Salazar, J. K., Babaria, H., Fay, M. L., & Stewart, D. S. (2024). Method of Inoculation Influences the Survival of *Salmonella enterica* on Retail and Orchard Peaches. *Journal of Food Protection*, 87(6), 100289. <https://doi.org/10.1016/j.jfp.2024.100289>

- Kidgell, C., Reichard, U., Wain, J., Linz, B., Torpdahl, M., Dougan, G., & Achtman, M. (2002). *Salmonella* Typhi, the causative agent of typhoid fever, is approximately 50,000 years old. *Infection, Genetics and Evolution*, 2(1), 39–45.
[https://doi.org/10.1016/S1567-1348\(02\)00089-8](https://doi.org/10.1016/S1567-1348(02)00089-8)
- Kim, M., Barnett-Neefs, C., Chavez, R. A., Kealey, E., Wiedmann, M., & Stasiewicz, M. J. (2024). Risk Assessment Predicts Most of the Salmonellosis Risk in Raw Chicken Parts is Concentrated in Those Few Products with High Levels of High-Virulence Serotypes of *Salmonella*. *Journal of Food Protection*, 87(7), 100304.
<https://doi.org/10.1016/j.jfp.2024.100304>
- Kim, S., Pachepsky, Y., Micallef, S. A., Rosenberg Goldstein, R., Sapkota, A. R., Hashem, F., Parveen, S., Kniel, K. E., & Sharma, M. (2023). Temporal Stability of *Salmonella enterica* and *Listeria monocytogenes* in Surface Waters Used for Irrigation in the Mid-Atlantic United States. *Journal of Food Protection*, 86(4), 100058. <https://doi.org/10.1016/j.jfp.2023.100058>
- Kinde, H., Adelson, M., Ardans, A., Little, E. H., Willoughby, D., Berchtold, D., Read, D. H., Breitmeyer, R., Kerr, D., Tarbell, R., & Hughes, E. (1997). Prevalence of *Salmonella* in Municipal Sewage Treatment Plant Effluents in Southern California. *Avian Diseases*, 41(2), 392. <https://doi.org/10.2307/1592195>
- Kirk, J. H., Holmberg, C. A., & Jeffrey, J. S. (2002). Prevalence of *Salmonella* spp. in selected birds captured on California dairies. *Journal of the American Veterinary Medical Association*, 220(3), 359–362.
<https://doi.org/10.2460/javma.2002.220.359>

- Knodler, L. A., Vallance, B. A., Celli, J., Winfree, S., Hansen, B., Montero, M., & Steele-Mortimer, O. (2010). Dissemination of invasive *Salmonella* via bacterial-induced extrusion of mucosal epithelia. *Proceedings of the National Academy of Sciences*, 107(41), 17733–17738. <https://doi.org/10.1073/pnas.1006098107>
- Knox, R. (1945). The Effect of Tetrathionate on Bacterial Growth. *The British Journal of Experimental Pathology*, 26(3), 146–150.
- Komatsu, K. K., & Restaino, L. (1981). Determination Of The Effectiveness Of Novobiocin Added To Two Agar Plating Media For The Isolation Of *Salmonella* From Fresh Meat Products. *Journal of Food Safety*, 3(3), 183–192. <https://doi.org/10.1111/j.1745-4565.1981.tb00420.x>
- Kraft, A. L., Wells, J. E., Frye, J. G., Ibekwe, A. M., Durso, L. M., Hiott, L., East, C., McConn, B. R., Franklin, A. M., Boczek, L. A., Garland, J. L., Kabera, C., McDermott, P. F., Ottesen, A. R., Zheng, J., Cook, K. L., & Sharma, M. (2023). A comparison of methods to detect low levels of *Salmonella enterica* in surface waters to support antimicrobial resistance surveillance efforts performed in multiple laboratories. *Science of The Total Environment*, 905, 167189. <https://doi.org/10.1016/j.scitotenv.2023.167189>
- Krawiec, M., Kuczkowski, M., Kruszewicz, A. G., & Wieliczko, A. (2015). Prevalence and genetic characteristics of *Salmonella* in free-living birds in Poland. *BMC Veterinary Research*, 11(1), 15. <https://doi.org/10.1186/s12917-015-0332-x>
- Krishnan, A., Kogan, C., Peters, R. T., Thomas, E. L., & Critzer, F. (2021). Microbial and physicochemical assessment of irrigation water treatment methods. *Journal of Applied Microbiology*, 131(3), 1555–1562. <https://doi.org/10.1111/jam.15043>

- Krishnan, A., Xu, X., Tamayo, M. S., Mishra, A., & Critzer, F. (2023). Impact of chlorine or peracetic acid on inactivation of *Salmonella*, *Escherichia coli*, and *Listeria monocytogenes* in agricultural water. *Science of The Total Environment*, 885, 163884. <https://doi.org/10.1016/j.scitotenv.2023.163884>
- Krumwiede, C., & Kohn, L. A. (1917). A Triple-Sugar Modification of the Russell Double-Sugar Medium. *The Journal of Medical Research*, 37(2), 225–227.
- Laborda, P., Sanz-García, F., Ochoa-Sánchez, L. E., Gil-Gil, T., Hernando-Amado, S., & Martínez, J. L. (2022). Wildlife and Antibiotic Resistance. *Frontiers in Cellular and Infection Microbiology*, 12. <https://doi.org/10.3389/fcimb.2022.873989>
- Lagatolla, C., Dolzani, L., Tonin, E., Lavenia, A., Di Michele, M., Tommasini, T., & Monti-Bragadin, C. (1996). PCR ribotyping for characterizing *Salmonella* isolates of different serotypes. *Journal of Clinical Microbiology*, 34(10), 2440–2443. <https://doi.org/10.1128/jcm.34.10.2440-2443.1996>
- Lam, S., & Roth, J. R. (1983). IS200: A *Salmonella*-specific insertion sequence. *Cell*, 34(3), 951–960. [https://doi.org/10.1016/0092-8674\(83\)90552-4](https://doi.org/10.1016/0092-8674(83)90552-4)
- Langholz, J., & Jay-Russell, M. (2013). Potential role of wildlife in pathogenic contamination of fresh produce. *Human-Wildlife Interactions*, 7(1), 140–157.
- Langridge, G. C., Fookes, M., Connor, T. R., Feltwell, T., Feasey, N., Parsons, B. N., Seth-Smith, H. M. B., Barquist, L., Stedman, A., Humphrey, T., Wigley, P., Peters, S. E., Maskell, D. J., Corander, J., Chabalgoity, J. A., Barrow, P., Parkhill, J., Dougan, G., & Thomson, N. R. (2015). Patterns of genome evolution that have accompanied host adaptation in *Salmonella*. *Proceedings of the National Academy of Sciences*, 112(3), 863–868. <https://doi.org/10.1073/pnas.1416707112>

- Lawson, B., Hughes, L. A., Peters, T., de Pinna, E., John, S. K., Macgregor, S. K., & Cunningham, A. A. (2011). Pulsed-Field Gel Electrophoresis Supports the Presence of Host-Adapted *Salmonella enterica* subsp. *enterica* Serovar Typhimurium Strains in the British Garden Bird Population. *Applied and Environmental Microbiology*, 77(22), 8139–8144.
<https://doi.org/10.1128/AEM.00131-11>
- Leader, B. T., Frye, J. G., Hu, J., Fedorka-Cray, P. J., & Boyle, D. S. (2009). High-Throughput Molecular Determination of *Salmonella enterica* Serovars by Use of Multiplex PCR and Capillary Electrophoresis Analysis. *Journal of Clinical Microbiology*, 47(5), 1290–1299. <https://doi.org/10.1128/JCM.02095-08>
- Lee, D., Tertuliano, M., Harris, C., Vellidis, G., Levy, K., & Coolong, T. (2019). *Salmonella* Survival in Soil and Transfer Onto Produce via Splash Events. *Journal of Food Protection*, 82(12), 2023–2037. <https://doi.org/10.4315/0362-028X.JFP-19-066>
- Li, B., Jackson, S. A., Gangiredla, J., Wang, W., Liu, H., Tall, B. D., Beaubrun, J. J.-G., Jay-Russell, M., Vellidis, G., & Elkins, C. A. (2015). Genomic Evidence Reveals Numerous *Salmonella enterica* Serovar Newport Reintroduction Events in Suwannee Watershed Irrigation Ponds. *Applied and Environmental Microbiology*, 81(24), 8243–8253. <https://doi.org/10.1128/AEM.02179-15>
- Li, B., Vellidis, G., Liu, H., Jay-Russell, M., Zhao, S., Hu, Z., Wright, A., & Elkins, C. A. (2014). Diversity and Antimicrobial Resistance of *Salmonella enterica* Isolates from Surface Water in Southeastern United States. *Applied and Environmental Microbiology*, 80(20), 6355–6365. <https://doi.org/10.1128/AEM.02063-14>

- Liao, C.-H., & Shollenberger, L. M. (2003). Survivability and long-term preservation of bacteria in water and in phosphate-buffered saline. *Letters in Applied Microbiology*, 37(1), 45–50. <https://doi.org/10.1046/j.1472-765X.2003.01345.x>
- Lienau, E. K., Strain, E., Wang, C., Zheng, J., Ottesen, A. R., Keys, C. E., Hammack, T. S., Musser, S. M., Brown, E. W., Allard, M. W., Cao, G., Meng, J., & Stones, R. (2011). Identification of a Salmonellosis Outbreak by Means of Molecular Sequencing. *New England Journal of Medicine*, 364(10), 981–982. <https://doi.org/10.1056/NEJMc1100443>
- Lindquist, H. D. A., Harris, S., Lucas, S., Hartzel, M., Riner, D., Rochele, P., & DeLeon, R. (2007). Using ultrafiltration to concentrate and detect *Bacillus anthracis*, *Bacillus atrophaeus* subspecies *globigii*, and *Cryptosporidium parvum* in 100-liter water samples. *Journal of Microbiological Methods*, 70(3), 484–492. <https://doi.org/10.1016/j.mimet.2007.06.007>
- Liu, B., Knirel, Y. A., Feng, L., Perepelov, A. V., Senchenkova, S. N., Reeves, P. R., & Wang, L. (2014). Structural diversity in *Salmonella* O antigens and its genetic basis. *FEMS Microbiology Reviews*, 38(1), 56–89. <https://doi.org/10.1111/1574-6976.12034>
- Liu, F., Barrangou, R., Gerner-Smidt, P., Ribot, E. M., Knabel, S. J., & Dudley, E. G. (2011). Novel Virulence Gene and Clustered Regularly Interspaced Short Palindromic Repeat (CRISPR) Multilocus Sequence Typing Scheme for Subtyping of the Major Serovars of *Salmonella enterica* subsp. *enterica*. *Applied and Environmental Microbiology*, 77(6), 1946–1956. <https://doi.org/10.1128/AEM.02625-10>

- Liu, F., Kariyawasam, S., Jayarao, B. M., Barrangou, R., Gerner-Smidt, P., Ribot, E. M., Knabel, S. J., & Dudley, E. G. (2011). Subtyping *Salmonella enterica* Serovar Enteritidis Isolates from Different Sources by Using Sequence Typing Based on Virulence Genes and Clustered Regularly Interspaced Short Palindromic Repeats (CRISPRs). *Applied and Environmental Microbiology*, 77(13), 4520–4526.
<https://doi.org/10.1128/AEM.00468-11>
- Liu, H., Whitehouse, C. A., & Li, B. (2018). Presence and Persistence of *Salmonella* in Water: The Impact on Microbial Quality of Water and Food Safety. *Frontiers in Public Health*, 6. <https://doi.org/10.3389/fpubh.2018.00159>
- Liu, P., Hill, V. R., Hahn, D., Johnson, T. B., Pan, Y., Jothikumar, N., & Moe, C. L. (2012). Hollow-fiber ultrafiltration for simultaneous recovery of viruses, bacteria and parasites from reclaimed water. *Journal of Microbiological Methods*, 88(1), 155–161. <https://doi.org/10.1016/j.mimet.2011.11.007>
- Liu, P., Ibaraki, M., Kapoor, R., Amin, N., Das, A., Miah, R., Mukhopadhyay, A. K., Rahman, M., Dutta, S., & Moe, C. L. (2021). Development of Moore Swab and Ultrafiltration Concentration and Detection Methods for *Salmonella* Typhi and *Salmonella* Paratyphi A in Wastewater and Application in Kolkata, India and Dhaka, Bangladesh. *Frontiers in Microbiology*, 12.
<https://doi.org/10.3389/fmicb.2021.684094>
- Lou, L., Zhang, P., Piao, R., & Wang, Y. (2019). *Salmonella* Pathogenicity Island 1 (SPI-1) and Its Complex Regulatory Network. *Frontiers in Cellular and Infection Microbiology*, 9. <https://doi.org/10.3389/fcimb.2019.00270>

- Luk, J. M., Kongmuang, U., Reeves, P. R., & Lindberg, A. A. (1993). Selective amplification of abequose and paratose synthase genes (rfb) by polymerase chain reaction for identification of *Salmonella* major serogroups (A, B, C2, and D). *Journal of Clinical Microbiology*, 31(8), 2118–2123.
<https://doi.org/10.1128/jcm.31.8.2118-2123.1993>
- Luo, Z., Gu, G., Ginn, A., Giurcanu, M. C., Adams, P., Vellidis, G., van Bruggen, A. H. C., Danyluk, M. D., & Wright, A. C. (2015). Distribution and Characterization of *Salmonella enterica* Isolates from Irrigation Ponds in the Southeastern United States. *Applied and Environmental Microbiology*, 81(13), 4376–4387.
<https://doi.org/10.1128/AEM.04086-14>
- Macnab, R. M. (1996). “Flagella and motility.” *Escherichia coli* and *Salmonella*: cellular and molecular biology.
- Mafizur, R. M., Sangjin, L., & Chul, P. Y. (2024). Prevalence of *Salmonella* spp. and *Escherichia coli* in the feces of free-roaming wildlife throughout South Korea. *PLOS ONE*, 19(2), e0281006. <https://doi.org/10.1371/journal.pone.0281006>
- Mahagamage, M. G. Y. L., Pathirage, M. V. S. C., & Manage, P. M. (2020). Contamination Status of *Salmonella* spp., *Shigella* spp. and *Campylobacter* spp. in Surface and Groundwater of the Kelani River Basin, Sri Lanka. *Water*, 12(8), 2187. <https://doi.org/10.3390/w12082187>
- Maiden, M. C. J. (2006). Multilocus Sequence Typing of Bacteria. *Annual Review of Microbiology*, 60(1), 561–588.
<https://doi.org/10.1146/annurev.micro.59.030804.121325>

- Majowicz, S. E., Musto, J., Scallan, E., Angulo, F. J., Kirk, M., O'Brien, S. J., Jones, T. F., Fazil, A., & Hoekstra, R. M. (2010). The Global Burden of Nontyphoidal *Salmonella* Gastroenteritis. *Clinical Infectious Diseases*, 50(6), 882–889.
<https://doi.org/10.1086/650733>
- Mandal, B. K., & Brennand, J. (1988). Bacteraemia in salmonellosis: a 15 year retrospective study from a regional infectious diseases unit. *BMJ*, 297(6658), 1242–1243. <https://doi.org/10.1136/bmj.297.6658.1242>
- Maron, D. F., Smith, T. J., & Nachman, K. E. (2013). Restrictions on antimicrobial use in food animal production: an international regulatory and economic survey. *Globalization and Health*, 9(1), 48. <https://doi.org/10.1186/1744-8603-9-48>
- Mather, A. E., Lawson, B., de Pinna, E., Wigley, P., Parkhill, J., Thomson, N. R., Page, A. J., Holmes, M. A., & Paterson, G. K. (2016). Genomic Analysis of *Salmonella enterica* Serovar Typhimurium from Wild Passerines in England and Wales. *Applied and Environmental Microbiology*, 82(22), 6728–6735.
<https://doi.org/10.1128/AEM.01660-16>
- Maurer, J. J., Martin, G., Hernandez, S., Cheng, Y., Gerner-Smidt, P., Hise, K. B., Tobin D'Angelo, M., Cole, D., Sanchez, S., Madden, M., Valeika, S., Presotto, A., & Lipp, E. K. (2015). Diversity and Persistence of *Salmonella enterica* Strains in Rural Landscapes in the Southeastern United States. *PLOS ONE*, 10(7), e0128937. <https://doi.org/10.1371/journal.pone.0128937>
- McDougall, F., & Power, M. (2021). Occurrence of *Salmonella enterica* in grey-headed flying foxes from New South Wales. *Australian Veterinary Journal*, 99(12), 517–521. <https://doi.org/10.1111/avj.13116>

- McEgan, R., Chandler, J. C., Goodridge, L. D., & Danyluk, M. D. (2014). Diversity of *Salmonella* Isolates from Central Florida Surface Waters. *Applied and Environmental Microbiology*, 80(21), 6819–6827.
<https://doi.org/10.1128/AEM.02191-14>
- McEgan, R., Mootian, G., Goodridge, L. D., Schaffner, D. W., & Danyluk, M. D. (2013). Predicting *Salmonella* Populations from Biological, Chemical, and Physical Indicators in Florida Surface Waters. *Applied and Environmental Microbiology*, 79(13), 4094–4105. <https://doi.org/10.1128/AEM.00777-13>
- McMillan, E. A., Wasilenko, J. L., Tagg, K. A., Chen, J. C., Simmons, M., Gupta, S. K., Tillman, G. E., Folster, J., Jackson, C. R., & Frye, J. G. (2020). Carriage and Gene Content Variability of the pESI-Like Plasmid Associated with *Salmonella* *Infantis* Recently Established in United States Poultry Production. *Genes*, 11(12), 1516. <https://doi.org/10.3390/genes11121516>
- McQuiston, J. R., Parrenas, R., Ortiz-Rivera, M., Gheesling, L., Brenner, F., & Fields, P. I. (2004). Sequencing and Comparative Analysis of Flagellin Genes *fliC* , *fljB* , and *flpA* from *Salmonella*. *Journal of Clinical Microbiology*, 42(5), 1923–1932.
<https://doi.org/10.1128/JCM.42.5.1923-1932.2004>
- Meinersmann, R. J., Berrang, M. E., Jackson, C. R., Fedorka-Cray, P., Ladely, S., Little, E., Frye, J. G., & Mattsson, B. (2008). *Salmonella*, *Campylobacter* and *Enterococcus* spp.: Their Antimicrobial Resistance Profiles and their Spatial Relationships in a Synoptic Study of the Upper Oconee River Basin. *Microbial Ecology*, 55(3), 444–452. <https://doi.org/10.1007/s00248-007-9290-6>

- Micallef, S. A., Rosenberg Goldstein, R. E., George, A., Kleinfelter, L., Boyer, M. S., McLaughlin, C. R., Estrin, A., Ewing, L., Jean-Gilles Beaubrun, J., Hanes, D. E., Kothary, M. H., Tall, B. D., Razeq, J. H., Joseph, S. W., & Sapkota, A. R. (2012). Occurrence and antibiotic resistance of multiple *Salmonella* serotypes recovered from water, sediment and soil on mid-Atlantic tomato farms. *Environmental Research*, 114, 31–39. <https://doi.org/10.1016/j.envres.2012.02.005>
- Michael, G. B., & Schwarz, S. (2016). Antimicrobial resistance in zoonotic nontyphoidal *Salmonella*: an alarming trend? *Clinical Microbiology and Infection*, 22(12), 968–974. <https://doi.org/10.1016/j.cmi.2016.07.033>
- Millan, J., Aduriz, G., Moreno, B., Juste, R. A., & Barral, M. (2004). *Salmonella* isolates from wild birds and mammals in the Basque Country (Spain). *Revue Scientifique et Technique de l'OIE*, 23(3), 905–911. <https://doi.org/10.20506/rst.23.3.1529>
- Miller, R. G., Tate, C. R., Mallinson, E. T., & Scherrer, J. A. (1991). Xylose-Lysine-Tergitol 4: An Improved Selective Agar Medium for the Isolation of *Salmonella*. *Poultry Science*, 70(12), 2429–2432. <https://doi.org/10.3382/ps.0702429>
- Moats, W. (1981). Update on *Salmonella* in Foods: Selective Plating Media and Other Diagnostic Media. *Journal of Food Protection*, 44(5), 375–380.
- Moats, W. A., & Kinner, J. A. (1974). Factors Affecting Selectivity of Brilliant Green-Phenol Red Agar for *Salmonellae*. *Applied Microbiology*, 27(1), 118–123. <https://doi.org/10.1128/am.27.1.118-123.1974>
- Mohammed, M., & Thapa, S. (2020). Evaluation of WGS-subtyping methods for epidemiological surveillance of foodborne salmonellosis. *One Health Outlook*, 2(1), 13. <https://doi.org/10.1186/s42522-020-00016-5>

- Monack, D. M., Raupach, B., Hromockyj, A. E., & Falkow, S. (1996). *Salmonella* typhimurium invasion induces apoptosis in infected macrophages. *Proceedings of the National Academy of Sciences*, 93(18), 9833–9838.
<https://doi.org/10.1073/pnas.93.18.9833>
- Moore, B. (1948). The Detection of Paratyphoid Carriers in Towns by means of Sewage Examination. *Monthly Bull. Ministry of Health & Pub. Health Lab. Service* (Directed by Med. Res. Council), 7, 241–248.
- Mughini-Gras, L., Heck, M., & van Pelt, W. (2016). Increase in reptile-associated human salmonellosis and shift toward adulthood in the age groups at risk, the Netherlands, 1985 to 2014. *Eurosurveillance*, 21(34).
<https://doi.org/10.2807/1560-7917.ES.2016.21.34.30324>
- Murphy, C. M., Strawn, L. K., Chapin, T. K., McEgan, R., Gopidi, S., Friedrich, L., Goodridge, L. D., Weller, D. L., Schneider, K. R., & Danyluk, M. D. (2022). Factors Associated With *E. coli* Levels in and *Salmonella* Contamination of Agricultural Water Differed Between North and South Florida Waterways. *Frontiers in Water*, 3. <https://doi.org/10.3389/frwa.2021.750673>
- Murphy, C. M., Weller, D. L., & Strawn, L. K. (2024). Scale and detection method impacted *Salmonella* prevalence and diversity in ponds. *Science of The Total Environment*, 907, 167812. <https://doi.org/10.1016/j.scitotenv.2023.167812>
- Murray, C. J. L., Ikuta, K. S., Sharara, F., Swetschinski, L., Robles Aguilar, G., Gray, A., Han, C., Bisignano, C., Rao, P., Wool, E., Johnson, S. C., Browne, A. J., Chipeta, M. G., Fell, F., Hackett, S., Haines-Woodhouse, G., Kashef Hamadani, B. H., Kumaran, E. A. P., McManigal, B., ... Naghavi, M. (2022). Global burden of

- bacterial antimicrobial resistance in 2019: a systematic analysis. *The Lancet*, 399(10325), 629–655. [https://doi.org/10.1016/S0140-6736\(21\)02724-0](https://doi.org/10.1016/S0140-6736(21)02724-0)
- Murray, M. H., Hernandez, S. M., Rozier, R. S., Kidd, A. D., Hepinstall-Cymerman, J., Curry, S. E., Yabsley, M. J., Adams, H., Ellison, T., Welch, C. N., & Lipp, E. K. (2021). Site Fidelity is Associated with Food Provisioning and *Salmonella* in an Urban Wading Bird. *EcoHealth*, 18(3), 345–358. <https://doi.org/10.1007/s10393-021-01543-x>
- Navarro-Gonzalez, N., Wright, S., Aminabadi, P., Gwinn, A., Suslow, T. V., & Jay-Russell, M. T. (2020). Carriage and Subtypes of Foodborne Pathogens Identified in Wild Birds Residing near Agricultural Lands in California: a Repeated Cross-Sectional Study. *Applied and Environmental Microbiology*, 86(3). <https://doi.org/10.1128/AEM.01678-19>
- Nozawa, T., Furukawa N., Aikawa, C., Watanabe, T., Haobam, B., Kurokawa, K., Maruyama, F., Nakagawa, I (2011). CRISPR Inhibition of Prophage Acquisition in *Streptococcus pyogenes*. *PLOS One*. 6(5).
- Nguyen, D. T. A., Awasthi, S. P., Hoang, P. H., Nguyen, P. Do, Jayedul, H., Hatanaka, N., Hinenoya, A., Van Dang, C., Faruque, S. M., & Yamasaki, S. (2021). Prevalence, Serovar, and Antimicrobial Resistance of Nontyphoidal *Salmonella* in Vegetable, Fruit, and Water Samples in Ho Chi Minh City, Vietnam. *Foodborne Pathogens and Disease*, 18(5), 354–363. <https://doi.org/10.1089/fpd.2020.2891>
- Odjadjare, E., & Olaniran, A. (2015). Prevalence of Antimicrobial Resistant and Virulent *Salmonella* spp. in Treated Effluent and Receiving Aquatic Milieu of Wastewater Treatment Plants in Durban, South Africa. *International Journal of Environmental*

Research and Public Health, 12(8), 9692–9713.

<https://doi.org/10.3390/ijerph120809692>

Oguadinma, I. C., Mishra, A., & Kumar, G. D. (2022). Sunlight Parameters Influence the Survival and Decline of *Salmonella* and *Escherichia coli* in Water. *Journal of Food Protection*, 85(11), 1614–1624. <https://doi.org/10.4315/JFP-22-042>

Olds, H. T., Corsi, S. R., Dila, D. K., Halmo, K. M., Bootsma, M. J., & McLellan, S. L. (2018). High levels of sewage contamination released from urban areas after storm events: A quantitative survey with sewage specific bacterial indicators. *PLOS Medicine*, 15(7), e1002614. <https://doi.org/10.1371/journal.pmed.1002614>

Olimpi, E. M., Garcia, K., Gonthier, D. J., Kremen, C., Snyder, W. E., Wilson-Rankin, E. E., & Karp, D. S. (2022). Semi-natural habitat surrounding farms promotes multifunctionality in avian ecosystem services. *Journal of Applied Ecology*, 59(4), 898–908. <https://doi.org/10.1111/1365-2664.14124>

Olimpi, E. M., Ke, A., Baur, P., Carlisle, L., Esquivel, K. E., Glaser, T., Snyder, W. E., Waterhouse, H., Bowles, T. M., Kremen, C., & Karp, D. S. (2024). Ungrazed seminatural habitats around farms benefit bird conservation without enhancing foodborne pathogen risks. *Landscape Ecology*, 39(7), 128.

<https://doi.org/10.1007/s10980-024-01907-y>

Olopoenia, L. A., & King, A. L. (2000). Widal agglutination test – 100 years later: still plagued by controversy. *Postgraduate Medical Journal*, 76(892), 80–84.

<https://doi.org/10.1136/pmj.76.892.80>

Olsen, K. M., & Larsson, H. (2004). *Gulls of Europe*.

- Oni, R. A., Sharma, M., & Buchanan, R. L. (2015). Survival of *Salmonella enterica* in Dried Turkey Manure and Persistence on Spinach Leaves. *Journal of Food Protection*, 78(10), 1791–1799. <https://doi.org/10.4315/0362-028X.JFP-15-047>
- Ormsby, M. J., Woodford, L., White, H. L., Fellows, R., & Quilliam, R. S. (2024). The plastisphere can protect *Salmonella* Typhimurium from UV stress under simulated environmental conditions. *Environmental Pollution*, 358, 124464. <https://doi.org/10.1016/j.envpol.2024.124464>
- Pal, A., Bailey, M. A., Talorico, A. A., Krehling, J. T., Macklin, K. S., Price, S. B., Buhr, R. J., & Bourassa, D. V. (2021). Impact of poultry litter *Salmonella* levels and moisture on transfer of *Salmonella* through associated in vitro generated dust. *Poultry Science*, 100(8), 101236. <https://doi.org/10.1016/j.psj.2021.101236>
- Palmer, T. (1976). Pest Bird Damage Control In Cattle Feedlots: The Integrated Systems Approach. 7th Vertebrate Pest Conference Proceedings Collection, 1721.
- Palmgren, H., Aspan, A., Broman, T., Bengtsson, K., Blomquist, L., Bergström, S., Sellin, M., Wollin, R., & Olsen, B. (2006). *Salmonella* in Black-headed gulls (*Larus ridibundus*); prevalence, genotypes and influence on *Salmonella* epidemiology. *Epidemiology and Infection*, 134(3), 635–644. <https://doi.org/10.1017/S0950268805005261>
- Partyka, M. L., Bond, R. F., Chase, J. A., & Atwill, E. R. (2018). Spatial and temporal variability of bacterial indicators and pathogens in six California reservoirs during extreme drought. *Water Research*, 129, 436–446. <https://doi.org/10.1016/j.watres.2017.11.038>

- Pascopella, L., Raupach, B., Ghorri, N., Monack, D., Falkow, S., & Small, P. L. C. (1995). Host Restriction Phenotypes of *Salmonella* typhi and *Salmonella* gallinarum. *American Society for Microbiology*, 63(17).
- Patchanee, P., Molla, B., White, N., Line, D. E., & Gebreyes, W. A. (2010). Tracking *Salmonella* Contamination in Various Watersheds and Phenotypic and Genotypic Diversity. *Foodborne Pathogens and Disease*, 7(9), 1113–1120.
<https://doi.org/10.1089/fpd.2010.0572>
- Patel, K., Stapleton, G. S., Trevejo, R. T., Tellier, W. T., Higa, J., Adams, J. K., Hernandez, S. M., Sanchez, S., Nemeth, N. M., Debess, E. E., Rogers, K. H., Mete, A., Watson, K. D., Foss, L., Low, M. S. F., Gollarza, L., & Nichols, M. (2023). Human Salmonellosis Outbreak Linked to *Salmonella* Typhimurium Epidemic in Wild Songbirds, United States, 2020–2021. *Emerging Infectious Diseases*, 29(11). <https://doi.org/10.3201/eid2911.230332>
- Pathogen Reduction; Hazard Analysis and Critical Control Point (HACCP) Systems; Final Rule (1996). Accessed on January 7, 2025.
<https://www.fsis.usda.gov/policy/federal-register-rulemaking/federal-register-rules/pathogen-reduction-hazard-analysis-and>
- Pees, M., Brockmann, M., Steiner, N., & Marschang, R. E. (2023). *Salmonella* in reptiles: a review of occurrence, interactions, shedding and risk factors for human infections. *Frontiers in Cell and Developmental Biology*, 11.
<https://doi.org/10.3389/fcell.2023.1251036>
- Peterz, M., Wiberg, C., & Norberg, P. (1989). The effect of incubation temperature and magnesium chloride concentration on growth of *Salmonella* in home-made and in

- commercially available dehydrated Rappaport-Vassiliadis broths. *Journal of Applied Bacteriology*, 66(6), 523–528. <https://doi.org/10.1111/j.1365-2672.1989.tb04573.x>
- Polo, F., Figueras, M. J., Inza, I., Sala, J., Fleisher, J. M., & Guarro, J. (1998). Relationship between presence of *Salmonella* and indicators of faecal pollution in aquatic habitats. *FEMS Microbiology Letters*, 160(2), 253–256. <https://doi.org/10.1111/j.1574-6968.1998.tb12919.x>
- Pourcel, C., Salvignol, G., & Vergnaud, G. (2005). CRISPR elements in *Yersinia pestis* acquire new repeats by preferential uptake of bacteriophage DNA, and provide additional tools for evolutionary studies. *Microbiology*, 151(3), 653–663. <https://doi.org/10.1099/mic.0.27437-0>
- Produce Safety Rule, 74354 (2015). Accessed on December 17, 2024. <https://www.fda.gov/food/food-safety-modernization-act-fsma/fsma-final-rule-produce-safety>
- Prum, R. O., Berv, J. S., Dornburg, A., Field, D. J., Townsend, J. P., Lemmon, E. M., & Lemmon, A. R. (2015). A comprehensive phylogeny of birds (Aves) using targeted next-generation DNA sequencing. *Nature*, 526(7574), 569–573. <https://doi.org/10.1038/nature15697>
- Pulford, C. V., Wenner, N., Redway, M. L., Rodwell, E. V., Webster, H. J., Escudero, R., Kröger, C., Canals, R., Rowe, W., Lopez, J., Hall, N., Rowley, P. D., Timofte, D., Harrison, R. A., Baker, K. S., & Hinton, J. C. D. (2019). The diversity, evolution and ecology of *Salmonella* in venomous snakes. *PLOS Neglected Tropical Diseases*, 13(6), e0007169. <https://doi.org/10.1371/journal.pntd.0007169>

- Rahn, K., De Grandis, S. A., Clarke, R. C., McEwen, S. A., Galán, J. E., Ginocchio, C., Curtiss, R., & Gyles, C. L. (1992). Amplification of an *invA* gene sequence of *Salmonella* Typhimurium by polymerase chain reaction as a specific method of detection of *Salmonella*. *Molecular and Cellular Probes*, 6(4), 271–279.
[https://doi.org/10.1016/0890-8508\(92\)90002-F](https://doi.org/10.1016/0890-8508(92)90002-F)
- Rajabi, M., Jones, M., Hubbard, M., Rodrick, G., & Wright, A. C. (2011). Distribution and Genetic Diversity of *Salmonella enterica* in the Upper Suwannee River. *International Journal of Microbiology*, 2011, 1–9.
<https://doi.org/10.1155/2011/461321>
- Ramey, A. M., & Ahlstrom, C. A. (2020). Antibiotic Resistant Bacteria In Wildlife: Perspectives On Trends, Acquisition And Dissemination, Data Gaps, And Future Directions. *Journal of Wildlife Diseases*, 56(1), 1. <https://doi.org/10.7589/2019-04-099>
- Ramos, R., Cerdà-Cuéllar, M., Ramírez, F., Jover, L., & Ruiz, X. (2010). Influence of Refuse Sites on the Prevalence of *Campylobacter* spp. and *Salmonella* Serovars in Seagulls. *Applied and Environmental Microbiology*, 76(9), 3052–3056.
<https://doi.org/10.1128/AEM.02524-09>
- Rappaport, F., Konforti, N., & Navon, B. (1956). A New Enrichment Medium for Certain *Salmonellae*. *Journal of Clinical Pathology*, 9(3), 261–266.
<https://doi.org/10.1136/jcp.9.3.261>
- Rasamsetti, S., Berrang, M., Cox, N. A., & Shariat, N. W. (2021). Selective pre-enrichment method to lessen time needed to recover *Salmonella* from commercial

- poultry processing samples. *Food Microbiology*, 99, 103818.
<https://doi.org/10.1016/j.fm.2021.103818>
- Refsum, T., Vikøren, T., Handeland, K., Kapperud, G., & Holstad, G. (2003).
 Epidemiologic And Pathologic Aspects Of *Salmonella* Typhimurium Infection In
 Passerine Birds In Norway. *Journal of Wildlife Diseases*, 39(1), 64–72.
<https://doi.org/10.7589/0090-3558-39.1.64>
- Renter, D. G., Gnad, D. P., Sargeant, J. M., & Hygnstrom, S. E. (2006). Prevalence and
 Serovars of *Salmonella* in the Feces of Free-Ranging White-Tailed Deer
 (*Odocoileus virginianus*) in Nebraska. *Journal of Wildlife Diseases*, 42(3), 699–
 703. <https://doi.org/10.7589/0090-3558-42.3.699>
- Restaino, L., Komatsu, K. K., & Syracuse, M. J. (1982). A note on novobiocin in XLD
 and HE agars: the optimum levels required in two commercial sources of media to
 improve isolation of salmonellas. *Journal of Applied Bacteriology*, 53, 285–288.
- Ricke, S. C. (2014). Application of Molecular Approaches for Understanding Foodborne
Salmonella Establishment in Poultry Production . *Advances in Biology*, 2014, 1–
 25. <https://doi.org/10.1155/2014/813275>
- Rigby, C. E., & Pettit, J. R. (1980). Delayed secondary enrichment for the isolation of
 salmonellae from broiler chickens and their environment. *Applied and
 Environmental Microbiology*, 40(4), 783–786.
<https://doi.org/10.1128/aem.40.4.783-786.1980>
- Rivadeneira, P., Hilson, C., Justice-Allen, A., & Jay-Russell, M. (2016). Pathogen Risks
 Related to the Movement of Birds Frequenting Livestock and Fresh Produce

- Growing Areas in the Southwestern U.S. Proceedings of the Vertebrate Pest Conference, 27, 258–263. <https://doi.org/10.5070/V427110583>
- Rivadeneira, P., Kross, S., Navarro-Gonzalez, N., & Jay-Russell, M. (2018). A Review of Bird Deterrents Used in Agriculture. Proceedings of the Vertebrate Pest Conference, 28, 218–223. <https://doi.org/10.5070/V42811040>
- Rodrigues, C., da Silva, A. L. B. R., & Dunn, L. L. (2019). Factors Impacting the Prevalence of Foodborne Pathogens in Agricultural Water Sources in the Southeastern United States. *Water*, 12(1), 51. <https://doi.org/10.3390/w12010051>
- Roszak, D. B., Grimes, D. J., & Colwell, R. R. (1984). Viable but nonrecoverable stage of *Salmonella* Enteritidis in aquatic systems. *Canadian Journal of Microbiology*, 30(3), 334–338. <https://doi.org/10.1139/m84-049>
- Roth, G. A., Abate, D., Abate, K. H., Abay, S. M., Abbafati, C., Abbasi, N., Abastabar, H., Abd-Allah, F., Abdela, J., Abdelalim, A., Abdollahpour, I., Abdulkader, R. S., Abebe, H. T., Abebe, M., Abebe, Z., Abejie, A. N., Abera, S. F., Abil, O. Z., Abraha, H. N., ... Murray, C. J. L. (2018). Global, regional, and national age-sex-specific mortality for 282 causes of death in 195 countries and territories, 1980–2017: a systematic analysis for the Global Burden of Disease Study 2017. *The Lancet*, 392(10159), 1736–1788. [https://doi.org/10.1016/S0140-6736\(18\)32203-7](https://doi.org/10.1016/S0140-6736(18)32203-7)
- Rubino, S., Muresu, E., Solinas, M., Santona, M., Paglietti, B., Azara, A., Schiaffino, A., Santona, A., Maida, A., & Cappuccinelli, P. (1998). IS 200 fingerprint of *Salmonella enterica* serotype Typhimurium human strains isolated in Sardinia. *Epidemiology and Infection*, 120(3), 215–222. <https://doi.org/10.1017/S095026889800867X>

- Ruiz, B. G.-V., Espinar, A. C., & Carmona, M. J. B. (1987). A comparative study of strains of *Salmonella* isolated from irrigation waters, vegetables and human infections. *Epidemiology and Infection*, 98(3), 271–276.
<https://doi.org/10.1017/S0950268800062026>
- Saelinger, C. A., Lewbart, G. A., Christian, L. S., & Lemons, C. L. (2006). Prevalence of *Salmonella* spp in cloacal, fecal, and gastrointestinal mucosal samples from wild North American turtles. *Journal of the American Veterinary Medical Association*, 229(2), 266–268. <https://doi.org/10.2460/javma.229.2.266>
- Salas-Rosas, L. M., López González, C. A., Sánchez-Cervantes, A., Vázquez-Peláez, C. G., Rodríguez-Torres, A., Cervantes-Chávez, J. A., & Olvera-Ramírez, A. M. (2020). Prevalence of *Salmonella* spp. and *Escherichia coli* O157 in a red deer herd (*Cervus elaphus scoticus*) in central Mexico. *Applied Animal Science*, 36(5), 622–629. <https://doi.org/10.15232/aas.2020-02032>
- Salmon, D. E., & Smith, T. (1891). Special report on the cause and prevention of swine plague. Gov't print. off. <https://doi.org/10.5962/bhl.title.36649>
- Santander, J., Roland, K. L., & Curtiss, R. (2008). Regulation of Vi capsular polysaccharide synthesis in *Salmonella enterica* serotype Typhi. *Journal of Infection in Developing Countries*, 2(6), 412–420.
<https://doi.org/10.3855/jidc.154>
- Santiago, P., Jiménez-Belenguer, A., García-Hernández, J., Estellés, R. M., Hernández Pérez, M., Castillo López, M. A., Ferrús, M. A., & Moreno, Y. (2018). High prevalence of *Salmonella* spp. in wastewater reused for irrigation assessed by

- molecular methods. *International Journal of Hygiene and Environmental Health*, 221(1), 95–101. <https://doi.org/10.1016/j.ijheh.2017.10.007>
- Santiago-Rodriguez, T. M., Tremblay, R. L., Toledo-Hernandez, C., Gonzalez-Nieves, J. E., Ryu, H., Santo Domingo, J. W., & Toranzos, G. A. (2012). Microbial Quality of Tropical Inland Waters and Effects of Rainfall Events. *Applied and Environmental Microbiology*, 78(15), 5160–5169. <https://doi.org/10.1128/AEM.07773-11>
- Santo Domingo, J., Harmon, S., & Bennett, J. (2000). Survival of *Salmonella* Species in River Water. *Current Microbiology*, 40, 409–417. <https://doi.org/https://doi.org/10.1007/s002840010079>
- Santos, R. L., Zhang, S., Tsois, R. M., Kingsley, R. A., Garry Adams, L., & Bäuml, A. J. (2001). Animal models of infections: enteritis versus typhoid fever. *Microbes and Infection*, 3(14–15), 1335–1344. [https://doi.org/10.1016/S1286-4579\(01\)01495-2](https://doi.org/10.1016/S1286-4579(01)01495-2)
- Scallan, E., Hoekstra, R. M., Angulo, F. J., Tauxe, R. V., Widdowson, M.-A., Roy, S. L., Jones, J. L., & Griffin, P. M. (2011). Foodborne Illness Acquired in the United States—Major Pathogens. *Emerging Infectious Diseases*, 17(1), 7–15. <https://doi.org/10.3201/eid1701.P11101>
- Schouls, L. M., Reulen, S., Duim, B., Wagenaar, J. A., Willems, R. J. L., Dingle, K. E., Colles, F. M., & Van Embden, J. D. A. (2003). Comparative Genotyping of *Campylobacter jejuni* by Amplified Fragment Length Polymorphism, Multilocus Sequence Typing, and Short Repeat Sequencing: Strain Diversity, Host Range,

- and Recombination. *Journal of Clinical Microbiology*, 41(1), 15–26.
<https://doi.org/10.1128/JCM.41.1.15-26.2003>
- Schwartz, D. C., & Cantor, C. (1964). Separation of Yeast Chromosome-Sized DNAs by Pulsed Field Gradient Gel Electrophoresis. In *Cell* (Vol. 37).
- Setti, I., Rodriguez-Castro, A., Pata, M. P., Cadarso-Suarez, C., Yacoubi, B., Bensmael, L., Moukrim, A., & Martinez-Urtaza, J. (2009). Characteristics and Dynamics of *Salmonella* Contamination along the Coast of Agadir, Morocco. *Applied and Environmental Microbiology*, 75(24), 7700–7709.
<https://doi.org/10.1128/AEM.01852-09>
- Sha, Q., Vatter, D. A., Forstner, M. R. J., & Hahn, D. (2013). Quantifying *Salmonella* Population Dynamics in Water and Biofilms. *Microbial Ecology*, 65(1), 60–67.
<https://doi.org/10.1007/s00248-012-0106-y>
- Shah, D. H., Paul, N. C., Sisco, W. C., Crespo, R., & Guard, J. (2017). Population dynamics and antimicrobial resistance of the most prevalent poultry-associated *Salmonella* serotypes. *Poultry Science*, 96(3), 687–702.
<https://doi.org/10.3382/ps/pew342>
- Shah, J., Desai, P. T., Chen, D., Stevens, J. R., & Weimer, B. C. (2013). Preadaptation to Cold Stress in *Salmonella enterica* Serovar Typhimurium Increases Survival during Subsequent Acid Stress Exposure. *Applied and Environmental Microbiology*, 79(23), 7281–7289. <https://doi.org/10.1128/AEM.02621-13>
- Shariat, N., DiMarzio, M. J., Yin, S., Dettinger, L., Sandt, C. H., Lute, J. R., Barrangou, R., & Dudley, E. G. (2013). The combination of CRISPR-MVLST and PFGE provides increased discriminatory power for differentiating human clinical

- isolates of *Salmonella enterica* subsp. *enterica* serovar Enteritidis. Food Microbiology, 34(1), 164–173. <https://doi.org/10.1016/j.fm.2012.11.012>
- Shariat, N., & Dudley, E. G. (2014). CRISPRs: Molecular Signatures Used for Pathogen Subtyping. Applied and Environmental Microbiology, 80(2), 430–439. <https://doi.org/10.1128/AEM.02790-13>
- Shariat, N., Kirchner, M. K., Sandt, C. H., Trees, E., Barrangou, R., & Dudley, E. G. (2013). Subtyping of *Salmonella enterica* Serovar Newport Outbreak Isolates by CRISPR-MVLST and Determination of the Relationship between CRISPR-MVLST and PFGE Results. Journal of Clinical Microbiology, 51(7), 2328–2336. <https://doi.org/10.1128/JCM.00608-13>
- Shariat, N., Sandt, C. H., DiMarzio, M. J., Barrangou, R., & Dudley, E. G. (2013). CRISPR-MVLST subtyping of *Salmonella enterica* subsp. *enterica* serovars Typhimurium and Heidelberg and application in identifying outbreak isolates. BMC Microbiology, 13(1), 254. <https://doi.org/10.1186/1471-2180-13-254>
- Shariat, N., Timme, R. E., Pettengill, J. B., Barrangou, R., & Dudley, E. G. (2015). Characterization and evolution of *Salmonella* CRISPR-Cas systems. Microbiology, 161(2), 374–386. <https://doi.org/10.1099/mic.0.000005>
- Sharma, M., Handy, E. T., East, C. L., Kim, S., Jiang, C., Callahan, M. T., Allard, S. M., Micallef, S., Craighead, S., Anderson-Coughlin, B., Gartley, S., Vanore, A., Kniel, K. E., Haymaker, J., Duncan, R., Foust, D., White, C., Taabodi, M., Hashem, F., ... Sapkota, A. R. (2020). Prevalence of *Salmonella* and *Listeria monocytogenes* in non-traditional irrigation waters in the Mid-Atlantic United

- States is affected by water type, season, and recovery method. PLOS ONE, 15(3), e0229365. <https://doi.org/10.1371/journal.pone.0229365>
- Siceloff, A. T., Ohta, N., Norman, K. N., Loneragan, G. H., Norby, B., Scott, H. M., & Shariat, N. W. (2021). Antimicrobial Resistance Hidden within Multiserovar *Salmonella* Populations. *Antimicrobial Agents and Chemotherapy*, 65(6), e00048-21. <https://doi.org/10.1128/AAC.00048-21>
- Siceloff, A. T., Waltman, D., & Shariat, N. W. (2022). Regional *Salmonella* Differences in United States Broiler Production from 2016 to 2020 and the Contribution of Multiserovar Populations to *Salmonella* Surveillance. *Applied and Environmental Microbiology*, 88(8). <https://doi.org/10.1128/aem.00204-22>
- Sikorski, M. J., & Levine, M. M. (2020). Reviving the “Moore Swab”: a Classic Environmental Surveillance Tool Involving Filtration of Flowing Surface Water and Sewage Water To Recover Typhoidal *Salmonella* Bacteria. *Applied and Environmental Microbiology*, 86(13). <https://doi.org/10.1128/AEM.00060-20>
- Silva, M. A., Fernandes, É. F. S. T., Santana, S. C., Marvulo, M. F. V., Barros, M. R., Vilela, S. M. O., Reis, E. M. F., Mota, R. A., & Silva, J. C. R. (2018). Isolation of *Salmonella* spp. in cattle egrets (*Bubulcus ibis*) from Fernando de Noronha Archipelago, Brazil. *Brazilian Journal of Microbiology*, 49(3), 559–563. <https://doi.org/10.1016/j.bjm.2018.01.004>
- Silverman, M., Zieg, J., Hilmen, M., & Simon, M. (1979). Phase variation in *Salmonella*: genetic analysis of a recombinational switch. *Proceedings of the National Academy of Sciences*, 76(1), 391–395. <https://doi.org/10.1073/pnas.76.1.391>

- Simpson, K. M. J., Hill-Cawthorne, G. A., Ward, M. P., & Mor, S. M. (2018). Diversity of *Salmonella* serotypes from humans, food, domestic animals and wildlife in New South Wales, Australia. *BMC Infectious Diseases*, 18(1), 623. <https://doi.org/10.1186/s12879-018-3563-1>
- Sirsat, S. A., Burkholder, K. M., Muthaiyan, A., Dowd, S. E., Bhunia, A. K., & Ricke, S. C. (2011). Effect of sublethal heat stress on *Salmonella* Typhimurium virulence. *Journal of Applied Microbiology*, 110(3), 813–822. <https://doi.org/10.1111/j.1365-2672.2011.04941.x>
- Smith, C. M., & Hill, V. R. (2009). Dead-End Hollow-Fiber Ultrafiltration for Recovery of Diverse Microbes from Water. *Applied and Environmental Microbiology*, 75(16), 5284–5289. <https://doi.org/10.1128/AEM.00456-09>
- Smith, O. M., Edworthy, A., Taylor, J. M., Jones, M. S., Tormanen, A., Kennedy, C. M., Fu, Z., Latimer, C. E., Cornell, K. A., Michelotti, L. A., Sato, C., Northfield, T., Snyder, W. E., & Owen, J. P. (2020). Agricultural intensification heightens food safety risks posed by wild birds. *Journal of Applied Ecology*, 57(11). <https://doi.org/10.1111/1365-2664.13723>
- Smith, O. M., Olimpi, E. M., & Karp, D. S. (2022). Feedlot Association Increases Food Safety Risk Associated with Wild Birds. *The Bulletin of the Ecological Society of America*, 103(2). <https://doi.org/10.1002/bes2.1974>
- Smith, O. M., Olimpi, E. M., Navarro-Gonzalez, N., Cornell, K. A., Frishkoff, L. O., Northfield, T. D., Bowles, T. M., Edworthy, M., Eilers, J., Fu, Z., Garcia, K., Gonthier, D. J., Jones, M. S., Kennedy, C. M., Latimer, C. E., Owen, J. P., Sato, C., Taylor, J. M., Wilson-Rankin, E. E., ... Karp, D. S. (2022). A trait-based

- framework for predicting foodborne pathogen risk from wild birds. *Ecological Applications*, 32(2). <https://doi.org/10.1002/eap.2523>
- Smith, O. M., Snyder, W. E., & Owen, J. P. (2020). Are we overestimating risk of enteric pathogen spillover from wild birds to humans? *Biological Reviews*, 95(3), 652–679. <https://doi.org/10.1111/brv.12581>
- Somda, N. S., Bonkougou, I. J. O., Sambe-Ba, B., Drabo, M. S., Wane, A. A., Sawadogo-Lingani, H., & Savadogo, A. (2021). Diversity and antimicrobial drug resistance of non-typhoid *Salmonella* serotypes isolated in lettuce, irrigation water and clinical samples in Burkina Faso. *Journal of Agriculture and Food Research*, 5, 100167. <https://doi.org/10.1016/j.jafr.2021.100167>
- Song, Q., Zhang, D., Gao, H., & Wu, J. (2018). *Salmonella* Species' Persistence and Their High Level of Antimicrobial Resistance in Flooded Man-Made Rivers in China. *Microbial Drug Resistance*, 24(9), 1404–1411. <https://doi.org/10.1089/mdr.2017.0316>
- Spector, M. P., & Kenyon, W. J. (2012). Resistance and survival strategies of *Salmonella enterica* to environmental stresses. *Food Research International*, 45(2), 455–481. <https://doi.org/10.1016/j.foodres.2011.06.056>
- Stenström, T., Conway, P., & Kjelleberg, S. (1989). Inhibition by antibiotics of the bacterial response to long-term starvation of *Salmonella* Typhimurium and the colon microbiota of mice. *Journal of Applied Bacteriology*, 67(1), 53–59. <https://doi.org/10.1111/j.1365-2672.1989.tb04954.x>
- Strawn, L. K., Gröhn, Y. T., Warchocki, S., Worobo, R. W., Bihn, E. A., & Wiedmann, M. (2013). Risk Factors Associated with *Salmonella* and *Listeria monocytogenes*

- Contamination of Produce Fields. *Applied and Environmental Microbiology*, 79(24), 7618–7627. <https://doi.org/10.1128/AEM.02831-13>
- Suhartono, S., Ismail, Y. S., & Aini, Z. (2021). Distribution of multidrug-resistant *Salmonella* spp. recovered from aquatic environment of Banda Aceh, Indonesia. *Biodiversitas Journal of Biological Diversity*, 22(2).
<https://doi.org/10.13057/biodiv/d220243>
- Tack, D. M., Marder, E. P., Griffin, P. M., Cieslak, P. R., Dunn, J., Hurd, S., Scallan, E., Lathrop, S., Muse, A., Ryan, P., Smith, K., Tobin-D'Angelo, M., Vugia, D. J., Holt, K. G., Wolpert, B. J., Tauxe, R., & Geissler, A. L. (2019). Preliminary incidence and trends of infections with pathogens transmitted commonly through food — Foodborne Diseases Active Surveillance Network, 10 U.S. sites, 2015–2018. *American Journal of Transplantation*, 19(6), 1859–1863.
<https://doi.org/10.1111/ajt.15412>
- Tardone, R., Rivera, D., Dueñas, F., Sallaberry-Pincheira, N., Hamilton-West, C., Adell, A. D., & Moreno-Switt, A. I. (2020). *Salmonella* in Raptors and Aquatic Wild Birds in Chile. *Journal of Wildlife Diseases*, 56(3), 707.
<https://doi.org/10.7589/2019-08-198>
- Taylor, W. (1965). Isolation of Shigellae. I. Xylose Lysine Agars; New Media for Isolation of Enteric Pathogens. *American Journal of Clinical Pathology*, 44(4), 471–475.
- Teunis, P. F. M., Kasuga, F., Fazil, A., Ogden, I. D., Rotariu, O., & Strachan, N. J. C. (2010). Dose–response modeling of *Salmonella* using outbreak data. *International*

- Journal of Food Microbiology, 144(2), 243–249.
<https://doi.org/10.1016/j.ijfoodmicro.2010.09.026>
- Thomas, J. L., Slawson, R. M., & Taylor, W. D. (2013). *Salmonella* serotype diversity and seasonality in urban and rural streams. Journal of Applied Microbiology, 114(3), 907–922. <https://doi.org/10.1111/jam.12079>
- Thompson, C. P., Doak, A. N., Amirani, N., Schroeder, E. A., Wright, J., Kariyawasam, S., Lamendella, R., & Shariat, N. W. (2018). High-Resolution Identification of Multiple *Salmonella* Serovars in a Single Sample by Using CRISPR-SeroSeq. Applied and Environmental Microbiology, 84(21).
<https://doi.org/10.1128/AEM.01859-18>
- Tickner, D., Parker, H., Moncrieff, C. R., Oates, N. E. M., Ludi, E., & Acreman, M. (2017). Managing Rivers for Multiple Benefits—A Coherent Approach to Research, Policy and Planning. Frontiers in Environmental Science, 5.
<https://doi.org/10.3389/fenvs.2017.00004>
- Timme, R. E., Sanchez Leon, M., & Allard, M. W. (2019). Utilizing the Public GenomeTrakr Database for Foodborne Pathogen Traceback.
https://doi.org/10.1007/978-1-4939-9000-9_17
- Tindall, B. J., Grimont, P. A. D., Garrity, G. M., & Euzéby, J. P. (2005). Nomenclature and taxonomy of the genus *Salmonella*. International Journal of Systematic and Evolutionary Microbiology, 55(1), 521–524. <https://doi.org/10.1099/ij.s.0.63580-0>
- Tizard, I. (2004). Salmonellosis in wild birds. Seminars in Avian and Exotic Pet Medicine, 13(2), 50–66. <https://doi.org/10.1053/j.saep.2004.01.008>

- Topalcengiz, Z., McEgan, R., & Danyluk, M. D. (2019). Fate of *Salmonella* in Central Florida Surface Waters and Evaluation of EPA Worst Case Water as a Standard Medium. *Journal of Food Protection*, 82(6), 916–925.
<https://doi.org/10.4315/0362-028X.JFP-18-331>
- Topalcengiz, Z., Spaninger, P. M., Jeamsripong, S., Persad, A. K., Buchanan, R. L., Saha, J., LeJeune, J., Jay-Russell, M. T., Kniel, K. E., & Danyluk, M. D. (2020). Survival of *Salmonella* in Various Wild Animal Feces That May Contaminate Produce. *Journal of Food Protection*, 83(4), 651–660.
<https://doi.org/10.4315/0362-028X.JFP-19-302>
- Toro, M., Weller, D., Ramos, R., Diaz, L., Alvarez, F. P., Reyes-Jara, A., Moreno-Switt, A. I., Meng, J., & Adell, A. D. (2022). Environmental and anthropogenic factors associated with the likelihood of detecting *Salmonella* in agricultural watersheds. *Environmental Pollution*, 306, 119298.
<https://doi.org/10.1016/j.envpol.2022.119298>
- Touchon, M., Charpentier, S., Clermont, O., Rocha, E., Denamur, E., & Branger, C (2011). CRISPR Distribution within the *Escherichia coli* Species Is Not Suggestive of Immunity-Associated Diversifying Selection. *Journal of Bacteriology*, 193 (10). <https://doi.org/10.1128/jb.01307-10>
- Touchon, M., & Rocha, E. P. C. (2010). The Small, Slow and Specialized CRISPR and Anti-CRISPR of *Escherichia* and *Salmonella*. *PLoS ONE*, 5(6), e11126.
<https://doi.org/10.1371/journal.pone.0011126>
- Truitt, L. N., Vazquez, K. M., Pfuntner, R. C., Rideout, S. L., Havelaar, A. H., & Strawn, L. K. (2018). Microbial Quality of Agricultural Water Used in Produce Preharvest

- Production on the Eastern Shore of Virginia. *Journal of Food Protection*, 81(10), 1661–1672. <https://doi.org/10.4315/0362-028X.JFP-18-185>
- Uelze, L., Bloch, A., Borowiak, M., Grobbel, M., Deneke, C., Fischer, M., Malorny, B., Pietsch, M., Simon, S., Szabó, I., Tausch, S. H., & Fischer, J. (2021). What WGS Reveals about *Salmonella enterica* subsp. *enterica* in Wildlife in Germany. *Microorganisms*, 9(9), 1911. <https://doi.org/10.3390/microorganisms9091911>
- USDA-ERS. (2023). Cost Estimates of Foodborne Illnesses. Accessed December 14, 2024. <https://www.ers.usda.gov/data-products/cost-estimates-of-foodborne-illnesses>
- USDA-FSIS. (2024a). Laboratory Sampling Data. Accessed January 14, 2025. <https://www.fsis.usda.gov/science-data/data-sets-visualizations/laboratory-sampling-data>.
- USDA-FSIS. (2024b). Microbiology Laboratory Guidebook. Accessed December 16, 2024. <https://www.fsis.usda.gov/news-events/publications/microbiology-laboratory-guidebook>
- Uzzau, S., Brown, D. J., Wallis, T., Rubino, S., Leori, G., Bernard, S., Casadesus, J., Platt, D. J., & Olsen, J. E. (2000). Host adapted serotypes of *Salmonella enterica*. *Epidemiology and Infection*, 125(2), 229–255. <https://doi.org/10.1017/S0950268899004379>
- Varriano, S., Snyder, W., Smith, J. C., Shariat, N. W., & Dunn, L. L. (2025). Detering Wild Birds During Fruit and Vegetable Production. *Food Protection Trends*, 45(1), 27–35.

- Vassiliadis, P. (1983). The Rappaport—Vassiliadis (RV) enrichment medium for the isolation of salmonellas: An overview. *Journal of Applied Bacteriology*, 54(1), 69–76. <https://doi.org/10.1111/j.1365-2672.1983.tb01302.x>
- Vassiliadis, P., Kalapothaki, V., Trichopoulos, D., Mavromatti, C., & Serie, C. (1981). Improved Isolation of Salmonellae from Naturally Contaminated Meat Products by Using Rappaport-Vassiliadis Enrichment Broth. *Applied and Environmental Microbiology*, 42(4), 615–618.
- Vereen, E., Lowrance, R. R., Jenkins, M. B., Adams, P., Rajeev, S., & Lipp, E. K. (2013). Landscape and seasonal factors influence *Salmonella* and *Campylobacter* prevalence in a rural mixed use watershed. *Water Research*, 47(16), 6075–6085. <https://doi.org/10.1016/j.watres.2013.07.028>
- Very, K. J., Kirchner, M. K., Shariat, N., Cottrell, W., Sandt, C. H., Dudley, E. G., Kariyawasam, S., & Jayarao, B. M. (2016). Prevalence and Spatial Distribution of *Salmonella* Infections in the Pennsylvania Raccoon (*Procyon lotor*). *Zoonoses and Public Health*, 63(3), 223–233. <https://doi.org/10.1111/zph.12222>
- Vilela, F. P., Rodrigues, D. dos P., Allard, M. W., & Falcão, J. P. (2024). CRISPR and CRISPR–MVLST reveal conserved spacer distribution and high similarity among *Salmonella enterica* serovar Infantis genomes from Brazil and other countries. *Molecular Genetics and Genomics*, 299(1), 61. <https://doi.org/10.1007/s00438-024-02147-0>
- Vosik, D., Tewari, D., Dettinger, L., M’ikanatha, N. M., & Shariat, N. W. (2018). CRISPR Typing and Antibiotic Resistance Correlates with Polyphyletic

- Distribution in Human Isolates of *Salmonella* Kentucky. Foodborne Pathogens and Disease, 15(2), 101–108. <https://doi.org/10.1089/fpd.2017.2298>
- Waltenburg, M. A., Perez, A., Salah, Z., Karp, B. E., Whichard, J., Tolar, B., Gollarza, L., Koski, L., Blackstock, A., Basler, C., & Nichols, M. (2022). Multistate reptile- and amphibian-associated salmonellosis outbreaks in humans, United States, 2009–2018. Zoonoses and Public Health, 69(8), 925–937. <https://doi.org/10.1111/zph.12990>
- Waltman, W. D., Horne, A. M., Pirkle, C., & Dickson, T. G. (1991). Use of Delayed Secondary Enrichment for the Isolation of *Salmonella* in Poultry and Poultry Environments. Avian Diseases, 35(1), 88. <https://doi.org/10.2307/1591299>
- Wang, L., & Reeves, P. R. (2000). The *Escherichia coli* O111 and *Salmonella enterica* O35 Gene Clusters: Gene Clusters Encoding the Same Colitose-Containing O Antigen Are Highly Conserved. Journal of Bacteriology, 182(18), 5256–5261. <https://doi.org/10.1128/JB.182.18.5256-5261.2000>
- Wang, L., Romana, L. K., & Reeves, P. R. (1992). Molecular analysis of a *Salmonella enterica* group E1 rfb gene cluster: O antigen and the genetic basis of the major polymorphism. Genetics, 130(3), 429–443. <https://doi.org/10.1093/genetics/130.3.429>
- Wattiau, P., Boland, C., & Bertrand, S. (2011). Methodologies for *Salmonella enterica* subsp. *enterica* Subtyping: Gold Standards and Alternatives. Applied and Environmental Microbiology, 77(22), 7877–7885. <https://doi.org/10.1128/AEM.05527-11>

- Wei, W., Zhang, S., Fleming, J., Chen, Y., Li, Z., Fan, S., Liu, Y., Wang, W., Wang, T., Liu, Y., Ren, B., Wang, M., Jiao, J., Chen, Y., Zhao, Y., Zhao, Y., Gu, S., Zhang X., Wan, L., Chen, T., Zhou, L., Chen, Y., Zhang, X., Li, C., Zhang, H., Bi, L., *Mycobacterium tuberculosis* type III-A CRISPR/Cas system crRNA and its maturation have atypical features. The FASEB Journal, 33(1), 1496-1509. <https://doi.org/10.1096/fj.201800557RR>
- Weller, D., Brassill, N., Rock, C., Ivanek, R., Mudrak, E., Roof, S., Ganda, E., & Wiedmann, M. (2020). Complex Interactions Between Weather, and Microbial and Physicochemical Water Quality Impact the Likelihood of Detecting Foodborne Pathogens in Agricultural Water. *Frontiers in Microbiology*, 11. <https://doi.org/10.3389/fmicb.2020.00134>
- WHO. (2023). Antibiotic Resistance. Accessed December 6, 2024. <https://www.who.int/news-room/fact-sheets/detail/antimicrobial-resistance>.
- Wilkinson, J. L., Boxall, A. B. A., Kolpin, D. W., Leung, K. M. Y., Lai, R. W. S., Galbán-Malagón, C., Adell, A. D., Mondon, J., Metian, M., Marchant, R. A., Bouzas-Monroy, A., Cuni-Sanchez, A., Coors, A., Carriquiriborde, P., Rojo, M., Gordon, C., Cara, M., Moermond, M., Luarte, T., ... Teta, C. (2022). Pharmaceutical pollution of the world's rivers. *Proceedings of the National Academy of Sciences*, 119(8). <https://doi.org/10.1073/pnas.2113947119>
- Winfield, M. D., & Groisman, E. A. (2003). Role of Nonhost Environments in the Lifestyles of *Salmonella* and *Escherichia coli*. *Applied and Environmental Microbiology*, 69(7), 3687–3694. <https://doi.org/10.1128/AEM.69.7.3687-3694.2003>

- Winter, S. E., Thiennimitr, P., Winter, M. G., Butler, B. P., Huseby, D. L., Crawford, R. W., Russell, J. M., Bevins, C. L., Adams, L. G., Tsolis, R. M., Roth, J. R., & Bäumlér, A. J. (2010). Gut inflammation provides a respiratory electron acceptor for *Salmonella*. *Nature*, 467(7314), 426–429. <https://doi.org/10.1038/nature09415>
- Wittenberger, K., & Dohlman, E. (2010). Impacts of the 2008-09 Foodborne Illness Outbreak Linked to *Salmonella* in Peanuts. Economic Research Service, 1–18.
- Yachison, C. A., Yoshida, C., Robertson, J., Nash, J. H. E., Kruczkiewicz, P., Taboada, E. N., Walker, M., Reimer, A., Christianson, S., Nichani, A., & Nadon, C. (2017). The Validation and Implications of Using Whole Genome Sequencing as a Replacement for Traditional Serotyping for a National *Salmonella* Reference Laboratory. *Frontiers in Microbiology*, 8. <https://doi.org/10.3389/fmicb.2017.01044>
- Yeh, H., Cox, N., Hinton, A., Berrang, M., Detection and Distribution of Clustered Regularly Interspaced Short Palindromic Repeats (CRISPRs) in *Campylobacter jejuni* Isolates from Chicken Livers. *Journal of Food Protection*, 87(4), 100250. <https://doi.org/10.1016/j.jfp.2024.100250>
- Yoshida, C. E., Kruczkiewicz, P., Laing, C. R., Lingohr, E. J., Gannon, V. P. J., Nash, J. H. E., & Taboada, E. N. (2016). The *Salmonella* In Silico Typing Resource (SISTR): An Open Web-Accessible Tool for Rapidly Typing and Subtyping Draft *Salmonella* Genome Assemblies. *PLOS ONE*, 11(1), e0147101. <https://doi.org/10.1371/journal.pone.0147101>
- You, Y., Rankin, S. C., Aceto, H. W., Benson, C. E., Toth, J. D., & Dou, Z. (2006). Survival of *Salmonella enterica* Serovar Newport in Manure and Manure-

- Amended Soils. *Applied and Environmental Microbiology*, 72(9), 5777–5783.
<https://doi.org/10.1128/AEM.00791-06>
- Zhang, S., den Bakker, H. C., Li, S., Chen, J., Dinsmore, B. A., Lane, C., Lauer, A. C., Fields, P. I., & Deng, X. (2019). SeqSero2: Rapid and Improved *Salmonella* Serotype Determination Using Whole-Genome Sequencing Data. *Applied and Environmental Microbiology*, 85(23). <https://doi.org/10.1128/AEM.01746-19>
- Zhou, Z., Alikhan, N.-F., Mohamed, K., Fan, Y., & Achtman, M. (2020). The EnteroBase user’s guide, with case studies on *Salmonella* transmissions, *Yersinia pestis* phylogeny, and *Escherichia* core genomic diversity. *Genome Research*, 30(1), 138–152. <https://doi.org/10.1101/gr.251678.119>
- Zimboro, M. J., Powers, D. A., Miller, S. M., Wilson, G. E., & Johnson, J. A. (2009). *Manual of Microbiological Culture Media*.

CHAPTER 2

PREVALENCE AND MOLECULAR CHARACTERIZATION OF *SALMONELLA* ISOLATED FROM WILD BIRDS IN FRESH PRODUCE ENVIRONMENTS¹

¹Jared C. Smith, Sofia Varriano, Kerrie Roach, Zach Snipes, Joshua L. Dawson, Justin Shealy, Laurel L. Dunn, William E. Snyder, Nikki W. Shariat. 2023. *Frontiers in Microbiology*. 14: 1272916. Reprinted here with permission of publisher.

2.1 Abstract

Wild birds pose a difficult food safety risk to manage because they can avoid traditional wildlife mitigation strategies, such as fences. Birds often use agricultural fields and structures as foraging and nesting areas, which can lead to defecation on crops and subsequent transfer of foodborne pathogens. To assess the food safety risk associated with these events, wild bird feces were collected from produce fields across the southeastern United States during the 2021 and 2022 growing seasons. In total 773 fecal samples were collected from 45 farms across Florida, Georgia, South Carolina, and Tennessee, and 2.1% ($n = 16$) of samples were *Salmonella*-positive. Importantly, 75% of *Salmonella* were isolated from moist feces, showing reduced *Salmonella* viability when feces dry out. 16S microbiome analysis showed that presence of culturable *Salmonella* in moist feces correlated to a higher proportion of the Enterobacteriaceae family. From the *Salmonella*-positive samples, 62.5% (10/16) contained multi-serovar *Salmonella* populations. Overall, 13 serovars were detected, including six most commonly attributed to human illness (Enteritidis, Newport, Typhimurium, Infantis, Saintpaul, and Muenchen). PCR screening identified an additional 59 *Salmonella*-positive fecal samples, which were distributed across moist ($n = 44$) and dried feces ($n = 15$). On-farm point counts and molecular identification from fecal samples identified 57 bird species, including for 10 *Salmonella*-positive fecal samples. Overall, there was a low prevalence of *Salmonella* in fecal samples, especially in dried feces, and we found no evidence of *Salmonella* transmission to proximal foliage or produce. Fecal samples collected in farms close together shared highly related isolates by whole genome sequencing and also had highly similar *Salmonella* populations with

comparable relative frequencies of the same serovars, suggesting the birds acquired *Salmonella* from a common source.

2.2 Introduction

Salmonella enterica is a leading contributor of bacterial foodborne illness in the United States (Scallan et al., 2011; Tack et al., 2019). While *Salmonella* is an enteric pathogen, it can be found in non-host environments, such as surface water and soil, as well as on produce (Critzler and Doyle, 2010; Gorski et al., 2011; Strawn et al., 2013; Reddy et al., 2016; Bardsley et al., 2021; Deaven et al., 2021), where it can survive and cause outbreaks (CDC, 2023). Consumption of contaminated produce causes an estimated 44.2% of salmonellosis cases in the United States [The Interagency Food Safety Analytics Collaboration (IFSAC), 2022]. In produce, contamination can occur through water, soil, equipment, personnel, and wildlife introduction events (Alegbeleye et al., 2018; Rodrigues et al., 2019; Devarajan et al., 2021, 2023). Because produce is often eaten raw and post-harvest kill steps are limited, there is a significant need to understand and mitigate potential sources of contamination. The Standards for the Growing, Harvesting, Packing, and Holding of Produce for Human Consumption (Produce Safety Rule, 74354, 2015) went into effect in 2016 as a part of the Food Safety Modernization Act. This rule set the first federally mandated standards for the safe production of fruits, vegetables, and nuts, and includes requirements for microbial quality of production and postharvest water, soil amendments, cleaning and sanitation practices, worker training and hygiene, and wildlife mitigation in order to reduce the likelihood of foodborne pathogen-contamination to produce. While many of these standards have clear guidelines, wildlife mitigation is often limited to physical barriers to prevent foraging from deer,

raccoons, and other land animals (Hamilton et al., 2015). These precautions do little to prevent the intrusion of birds, which can easily fly into fields to forage for plants, insects, or small rodents. Birds are a further challenge as they can become accustomed to deterrents and often fly long distances while migrating (Rivadeneira et al., 2018; Elsohaby et al., 2021).

Wild birds are known to carry foodborne pathogens, including *Salmonella enterica* subspecies *enterica* (Tizard, 2004). Studies performed in the western and southwestern United States found *Salmonella* prevalence in wild birds at 0.5–6.5% (Gorski et al., 2011; Rivadeneira et al., 2016; Navarro-Gonzalez et al., 2020). Additionally, flocks of wild bird can spread disease among individuals when congregating at common food and water sources (Hernandez et al., 2016). Outside of explosive mortality events caused by *Salmonella* serovar Typhimurium (Hernandez et al., 2012), *Salmonella* does not typically elicit symptoms in wild birds, so healthy carriers can transmit this pathogen without suffering from salmonellosis (Prosser et al., 2011). Transmission of pathogens from birds to produce can occur through defecation when birds are flying over fields or foraging for food. An outbreak of *Salmonella* serovar Typhimurium in 2009 found matching strains in birds, peanut crops, and human clinical cases (Hernandez et al., 2012). While birds can benefit farms by providing services like natural pest control (Karp et al., 2013), their habituation in production environments could play a role in the transmission of foodborne pathogens via fresh produce.

Salmonella enterica is a diverse species, consisting of over 2,600 distinct serovars that are categorized by their unique O (somatic) and H (flagellar) antigens (Grimont and Weill, 2007; Issenhuth-Jeanjean et al., 2014). Genomic diversity between these serovars

has led to differences in host specificity, pathogenesis, and antibiotic resistance profiles (Uzzau et al., 2000; Cheng et al., 2019). While some serovars are most typically found in a small number of reservoirs (e.g., serovar Enteritidis is most closely linked to poultry), others, such as serovar Typhimurium, are ubiquitous and found in a variety of different hosts. Further, *Salmonella* is often detected in food animal production systems and the environment as mixed populations of multiple serovars (Deaven et al., 2021; Siceloff et al., 2021, 2022; Obe et al., 2023). In some instances, low frequency serovars in these populations may have greater potential impacts on public health when they have clinically relevant antimicrobial resistance profiles (Siceloff et al., 2022) or are more often associated with human illness (Deaven et al., 2021). Traditional isolation techniques that rely on picking a small number of colonies from selective agar are unable to resolve complex multi-serovar *Salmonella* populations (Cason et al., 2011). This hurdle is overcome by deep serotyping approaches such as CRISPR-SeroSeq, which can resolve the relative frequencies of multiple serovars in a single sample (Thompson et al., 2018).

In this study, we investigated the role of wild birds in the transmission of *Salmonella* to produce foliage in the southeastern United States. This study region includes more than 12 million acres of cropland (CroplandCROS, 2022) where produce such as tomatoes, peppers, eggplant, and other fruit, vegetable, and nut crops are significant economic contributors. Wild bird feces were collected from produce fields over a two-year period and cultured for *Salmonella*. Deep serotyping and whole genome sequencing were performed to assess *Salmonella* populations and to estimate source

attribution. Additionally, wild bird species were identified with both physical and molecular techniques to associate pathogen transmission risk.

2.3 Materials and Methods

2.3.1 Site selection and overview of study design

To study the impact of wild bird activity upon produce contamination, 45 different farms across the southeastern United States (Tennessee, Georgia, Florida, and South Carolina) were visited between 1–6 times (average 2.4 visits/farm). Produce grown on these farms included peppers (bell, banana, and jalapeño), eggplant, cucumbers, tomatoes, squash, grapes, pole beans, and okra. These above ground produce were chosen because they pose a greater risk for human illness should they be contaminated, as many are often eaten raw. Additionally, selecting produce growing above ground reduced the incidence of identifying contamination from on-ground sources, such as soil, or rodents or other small wildlife that primarily forage on the soil surface. Farms in this study were diverse and included organic and conventional farms, commercial and family-run operations, mono- and polyculture farms, and some had livestock on and around the farm. To best measure the effect of seasonality on the prevalence of *Salmonella*, repeated sample collections were completed at farms, up to three times per sampling season (May–October), where possible. During each sampling visit, crops around the perimeter and the interior of the fields were inspected to identify wild bird fecal samples. When fecal samples were identified, the leaf containing the feces was removed and homogenized for culturing *Salmonella*. To evaluate the necessity of exclusion zones encouraged by groups such as the Leafy Green Marketing Agreement (LGMA), surface swabs of a piece of produce beneath the fecal sample and from the leaf of a neighboring plant downwind

were also collected. *Salmonella* was first identified by culture. Samples found positive for *Salmonella* culture were then further analyzed with additional molecular tools (e.g., PCR, WGS, and CRISPR-SeroSeq). The culture-negative samples were then analyzed by a *Salmonella* PCR.

2.3.2 Sample collection

Fecal samples were collected between sunrise and 11 am to capture on-field bird activity while also limiting UV exposure and reducing the opportunity for desiccation. Upon arriving at a farm, sampling was conducted around the perimeter of each field, followed by a step-wise sampling through the interior of the field. On smaller farms (or small (<1 acre) fields on a large farm), all individual rows were surveyed. When a fecal sample was identified, it was visually scored for moisture as either 1 (moist) or 0 (dry) as an indicator of freshness. Then, the leaf containing the fecal sample was removed, inserted into 2 mL buffered peptone water (BPW, Hardy Diagnostics, Ohio, USA) recovery media, and placed on ice until culturing (within 24 h). Because the fecal samples were small and because in some cases removing them from the leaf would lose some of the fecal material, the entire leaf was removed from the plant and then the portion containing just the feces and the leaf material directly under the feces were isolated and collected. To test for transmission of *Salmonella* from the fecal sample, the surface of a piece of produce under the leaf was swabbed, along with a leaf of a neighboring plant downwind from the fecal sample. These swabs were collected by soaking a sterile cotton ball in 3 mL of BPW and using sterile forceps to drag it across the top and bottom of the neighboring leaf and across the entire surface of the produce. Swabs were placed in a

cooler with ice packs and stored at 4°C for no more than 24 h or until culturing could begin in the laboratory.

2.3.3 *Salmonella* culturing

Fecal samples were homogenized by hand into the 2 mL of recovery media. For *Salmonella* isolation, 750uL of the homogenate was transferred into a culture tube containing 9.25 mL BPW and incubated at 42°C for 24 h. Then, this was sub-inoculated into 9 mL Tetrathionate (TT, Neogen Diagnostics, Michigan, USA) and 9.9 mL Rappaport-Vassiliadis (RV, Hardy Diagnostics, Ohio, USA) selective enrichment broths in parallel and incubated for 24 h at 37°C before being streaked onto Xylose Lysine Tergitol-4 agar plates (XLT-4, Hardy Diagnostics, Ohio, USA). The plates were incubated at 37°C for 24 h and inspected for black colonies as an indicator of presumptive *Salmonella* colonies. If no H₂S-positive colonies were present, the plates were re-incubated for another 24 h. Up to 2 colonies from each sample were selected and were re-streaked onto XLT-4 for isolation if needed. *Salmonella* isolates were grown in Luria Broth (LB, Hardy Diagnostics, Ohio, USA) where aliquots were used to make frozen glycerol stocks and for DNA isolation. If we observed presumptive *Salmonella* colonies, we then returned to the swabs from produce and neighboring foliage and cultured these using the same protocol.

2.3.4 DNA isolation and *Salmonella* PCR screen

The total genomic DNA was isolated from 500uL of the fecal/recovery media homogenate using the Genome Wizard kit (Promega, Wisconsin, USA), with the additional step of grinding the fecal pellet with a sterile mortar and pestle to disrupt the fecal particles before beginning the extraction. Prior to any PCR (i.e., for *Salmonella* or

for COI), DNA from this fecal/recovery media homogenate was screened with an internal amplification control (IAC) PCR to identify the presence PCR inhibitors (Rosenstrauss et al., 1998). The primers for IAC PCR were IAC_F (5'-AGTTGCAGTGTAACCGTCATGT-3') and IAC_R (5'-TCGACGAGACTCTGCTGTTAAG-3') and the IAC template control sequence was IAC (5'-AGTTGCAGTGTAACCGTCATGTACCAGTAATCTGCGTCGCACGTGTGCACCTAGTCTA ATCACTTATGACTCAGATAACTTAACAGCAGAGTCTCGTCGA-3'). For each reaction, the following components were mixed: 39.5uL sterile water, 5uL 10x Taq Buffer, 0.5uL 10uM forward primer, 0.5uL 10uM reverse primer, 0.3uL 100 mM dNTPs, and 1 U Taq polymerase, before 2uL of bird fecal DNA was added as template. Cycle conditions were as follows: 95°C for three minutes followed by 40 cycles of 95°C for 30 s, 56°C for 30 s, and 72°C for 30 s. This was followed by a final elongation of 72°C for two minutes and resting at 4°C. PCR products were visualized by gel electrophoresis. Where there was no amplification, suggesting the presence of PCR inhibitors, a 1:10 dilution of the bird fecal sample DNA was made, and the PCR repeated. In this study, nearly 10% (n = 75) of samples contained PCR inhibitors as shown by the IAC PCR. This inhibition was resolved when the template was diluted 10-fold in molecular grade water, and this dilution was used for all subsequent PCRs.

For the *Salmonella* screening, an *invA* PCR was used (Rahn et al., 1992). In this PCR, primers – InvA_F1 (5'-AACGTGTTTCCGTGCGTAAT-3') and InvA_R1 (5'-TCCATCAAATTAGCGGAGGC-3') were mixed with 38.5uL sterile water, 5uL 10x Taq Buffer, 2uL of 6.25ug/mL BSA, 1uL 10uM forward primer, 1uL 10uM reverse

primer, 0.25uL 100 mM dNTPs, and 1 U Taq polymerase before 2uL of bird fecal sample DNA was added as template. Cycling conditions began with an initial melting temperature of 95°C for three minutes followed by 40 cycles of the following: 95°C for 30 s, 56°C for 30 s, and 72°C for 30 s. A final elongation temperature of 72°C for two minutes was completed before resting at 4°C.

2.3.5 *Salmonella* weather analysis

For each site, weather data from the day prior to collection, including total precipitation, average wind, average humidity, and high temperature values were determined using the closest USGS weather stations. To identify relationships between weather variables and moist feces, we conducted a series of binomial generalized linear mixed models (GLMM) using the glmmTMB package V1.1.7 (Brooks et al., 2023) within R V4.1.1. All continuous variables were standardized prior to analysis. Visits nested within farm and year were used as random effects. We ran models using individual variables as a fixed effect in each model and considered different additive configurations of other weather variables. We assessed multicollinearity using the performance package V0.10.3 (Lüdtke et al., 2021) and homogeneity of variance using the DHARMA package V0.4.6 (Hartig and Lohse, 2022); models meeting these assumptions (i.e., VIF < 5 and equally distributed residuals, respectively) were retained for comparison. Models were compared using the Akaike Information Criterion adjusted for small sample sizes (AICc) using the R package AICcmodavg V2.3–2. We considered “top models” as those with $\Delta AICc \leq 2$ (Burnham and Anderson, 2002). The same weather stations were used to calculate the monthly average weather values (Supplementary Figure S2.1).

2.3.6 Whole genome sequencing

Total genomic DNA from *Salmonella* isolates was extracted using a Promega Genome Wizard DNA extraction kit (Promega, Wisconsin, USA) and sequenced on an Illumina MiSeq 500 cycle v2 chemistry kit (Illumina, California, USA). The sequence reads were assembled using SPAdes de-novo assembly (Version 3.15.5) (Bankevich et al., 2012) and the serovar determined using SeqSero 2.0 (Zhang et al., 2019). Sequences were uploaded to Enterobase (Zhou et al., 2020) where sequence types (ST) could be predicted and used to identify related isolates. Phylogenetic relatedness was visualized through GrapeTree and allelic differences were used to identify the closest related source type. The assembled genomes were uploaded to NCBI (Accession numbers SAMN33186945, SAMN33186956, SAMN33186963, SAMN33186964, SAMN33186971, SAMN33186984, SAMN33187804, SAMN33187835, SAMN33187836, SAMN33187842, SAMN33187843, SAMN33187878, SAMN33187961, SAMN33187962, SAMN33187972, SAMN33225914, SAMN37196586, SAMN37196587, and SAMN37196588).

2.3.7 *Salmonella* population analysis

To identify the populations of *Salmonella* within wild bird feces, TT and RV enrichments from *Salmonella* culture positive samples were processed individually by centrifuging 1 mL of each selective enrichment at 14,000 rpm for three minutes. Total genomic DNA was isolated from the resulting pellet using a Promega Genome Wizard Kit and resuspended in 200uL of molecular-grade water. A total of 2 µL of this template was used in the PCR for CRISPR-SeroSeq with primers targeting the conserved direct repeat sequences within *Salmonella* CRISPR arrays (Thompson et al., 2018; Siceloff et

al., 2022). Primers also included index sequences which facilitated multiplexed, high throughput sequencing. PCR products were purified using the Ampure system (Beckman Coulter, Indianapolis, IN) and pooled in approximate equimolar ratios. Pooled libraries were sequenced using the Illumina NextSeq 550 platform (Illumina, California, USA) mid output 150 cycle v2.5 kit with single-end reads. A water negative-control and a positive control containing *Salmonella* serovar Enteritidis genomic DNA with a known CRISPR profile were included in the library. Sequence reads were scanned and matched in a local BLAST search to a lab-curated database of over 150 serovars (Siceloff et al., 2022).

Serovars were called only if they contained multiple CRISPR spacers that were unique to that serovar. Where there were sufficient *Salmonella* sequence reads (>1,000 reads) for both the TT and RV enrichments the relative frequency of each serovar was normalized across both enrichments to provide a single serovar profile.

2.3.8 Microbiome analysis

All 16S rRNA Illumina-tag PCR reactions were performed on DNA extracts per the Earth Microbiome Project protocol (Walters et al., 2016). Negative controls (molecular grade water) were processed in parallel with the samples for PCR amplification. PCR products were pooled in batches of ~200 samples each and gel purified on a 2% agarose gel using the QIAquick Gel Purification Kit (Qiagen, Frederick, Maryland, USA). Before sequencing, purified pools were quality checked using an Agilent 2100 BioAnalyzer and Agilent DNA High Sensitivity DNA kit (Agilent Technologies, Santa Clara, California, USA). The purified pools were stored at -20°C, then sequenced using an Illumina MiSeq 500 cycle v2 chemistry kit (Illumina, California,

USA). Raw data were processed, analyzed, and quality checked with QIIME2 (Bolyen et al., 2019) before forward and reverse reads were merged and chimeras removed with DADA2 (Callahan et al., 2016). DADA2 was also used to assign sequences to amplicon sequence variants (ASVs) using a pre-trained Silva 132 Database (Quast et al., 2012). MAFFT (Kato and Standley, 2013) and FastTree (Price et al., 2010) were used to create a rooted phylogenetic tree using representative ASVs. Additionally, a biomarker analysis was completed to identify taxonomic groups that were differentially abundant within groupings of samples (*Salmonella* Culture, *Salmonella* PCR, and No *Salmonella*) using LEfSe (Segata et al., 2011) by normalizing the ASVs with the counts per million method and a differential abundance value of p of <0.05 and a log (LDA) score of at least 1.0.

2.3.9 Bird species identification

Wild birds were identified in two ways: physical identification (i.e., point counts) of birds present around and in fields, and molecular identification from feces. Point counts were conducted at all field locations on sample days between 6 and 10 am. One point count was done for every 10 hectares (ha) of sampled field when field conditions and harvesting schedules allowed. Points on the same farm were at least 200 m apart. Points were positioned approximately 90 m away from the edge of fields to overlap with bacterial sampling areas while still capturing birds moving in and out of produce. All birds seen and heard within a 100-m radius during a 10-min period were recorded, along with distance, detection method, and habitat. During the 10 min, birds were counted in sub-periods of three, three, and four minutes. Only new species were counted after the first sub-period to avoid counting the same individual multiple times. Birds flying overhead were excluded unless they were a species that forages aerially (e.g., swallows),

in which case a note was made that they were “aerial foraging.” The same observer conducted all counts for both years of sampling. Birds were categorized as in-field if they were observed interacting with produce (e.g., in tunnels, perching on produce stakes, or on produce plants) and other birds were categorized as off-field.

Molecular identification of wild bird species from fecal samples was completed using 2uL of DNA isolated from fecal samples as part of a PCR to amplify the Cytochrome C Oxidase Subunit I (COI). The sequence variability of the COI gene between bird species enables species identification. Many COI PCR assays were attempted, following published protocols (Hebert et al., 2003; Ivanova et al., 2007; Kerr et al., 2007; Joo and Park, 2012), but either did not yield amplicons or failed to produce quality sequences. This PCR used the following primers: COI_F1 (5'-CGCYTWAACAYTCYGCCATCTTACC-3') and COI_R1 (5'-ATTCCTATGTAGCCGAATGGTTCTTT-3') (Patel et al., 2010). For each reaction, the following were mixed into a 50uL reaction: 38.5uL sterile water, 5uL 10x Taq Buffer, 2uL 25 mM MgCl₂, 1uL 10uM forward primer, 1uL 10uM reverse primer, 0.3uL 100 mM dNTPs, and 1 U Taq polymerase along with adding 2uL of DNA template. The mix was run on the following PCR program: Initial melting of 95°C for four minutes was followed by five cycles of 95°C for 30 s, 59°C for 30 s, and 68°C for 45 s. This was followed by 40 cycles of 95°C for 30 s, 56°C for 30 s, and 68°C for 45 s and a final two-minute elongation step. Appropriately sized amplicons were sequenced in the forward and reverse direction by Eton Bioscience Inc. (Research Triangle Park, NC). SeqMan (Lasergene, DNA Star) was used to assemble the forward and reverse reads into a single

sequence, which was then compared to two databases: NCBI BLAST, and the Barcode of Life Database (Meiklejohn et al., 2019) with a 97% nucleotide identity threshold.

2.4 Results

During 2021 and 2022, 109 farm visits were performed across the southeastern United States, including Tennessee (n = 4 farms), North Georgia (n = 8), South Georgia (n = 20), South Carolina (n = 10), and North Florida (n = 3; Table 2.1). Farms ranged in size from 1.6–233 acres and included 13 small or independently owned farms (1.6–33.3 acres), as well as 32 large commercial farms (6.95–233 acres). Over the two seasons, 773 fecal samples were collected: 227 samples in 2021 and 546 in 2022. In total, 43.6% (337/773) of fecal samples were scored as moist, including 152 in 2021 and 185 in 2022, while 56.4% (436/773) were scored as dry, including 75 in 2021 and 361 in 2022 (Figure 2.1A).

By culture, *Salmonella* was isolated in 16 samples (16/773 total samples; 2.1%); 15 were identified in the first year of collection (15/227, 6.6%) and one was identified in the second year (1/546, 0.2%). Three quarters (12/16) of *Salmonella* samples were recovered from moist fecal samples (Figure 2.1B). *Salmonella*-positive samples were found in South Georgia (n = 10), Florida (n = 4), and North Georgia (n = 2). There was no recoverable incidence of transmission from fecal samples to produce below leaves with feces, nor to neighboring plants downwind. We screened all samples not confirmed positive by culture using a PCR targeting the *Salmonella* *invA* gene and detected *Salmonella* in 59 additional fecal samples, bringing the total *Salmonella*-positive samples to 75 (9.7%) (Table 2.2). Similar to culture-positive fecal samples, *Salmonella* was more commonly detected in the first year of collection, with 16.5% (35/212) of culture-

Table 2.1. Sampling distribution across the Southeast.

State	Number of farms (Number of visits)	Number of fecal samples collected
TN	4 (16)	218
SC	10 (29)	225
GA-N	8 (31)	235
GA-S	21 (27)	76
FL	3 (6)	19
Total	45 (109)	773

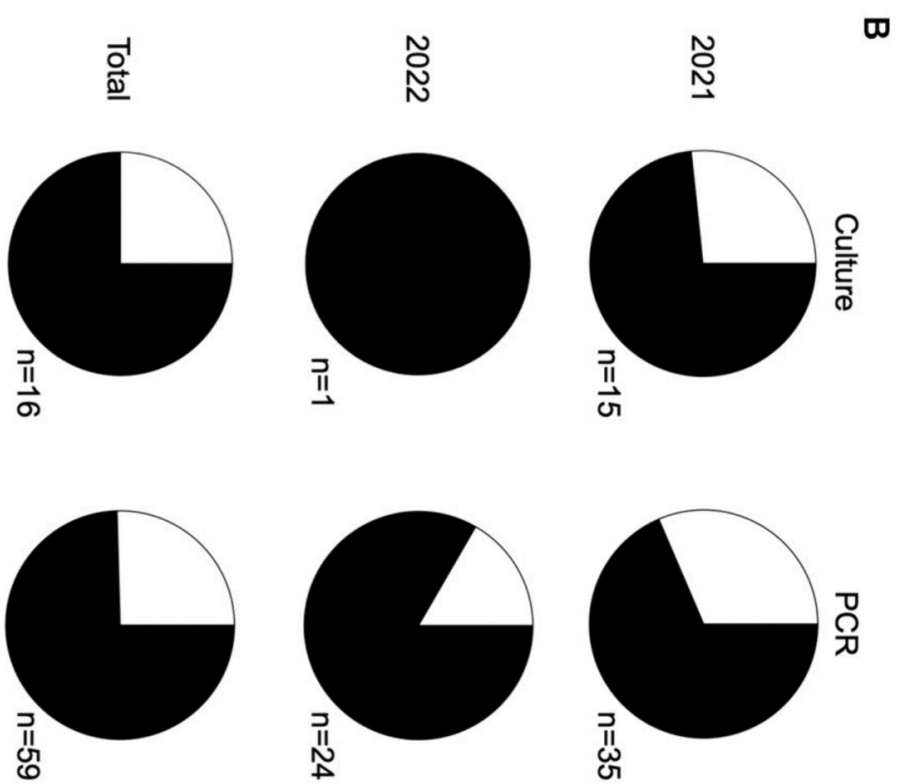
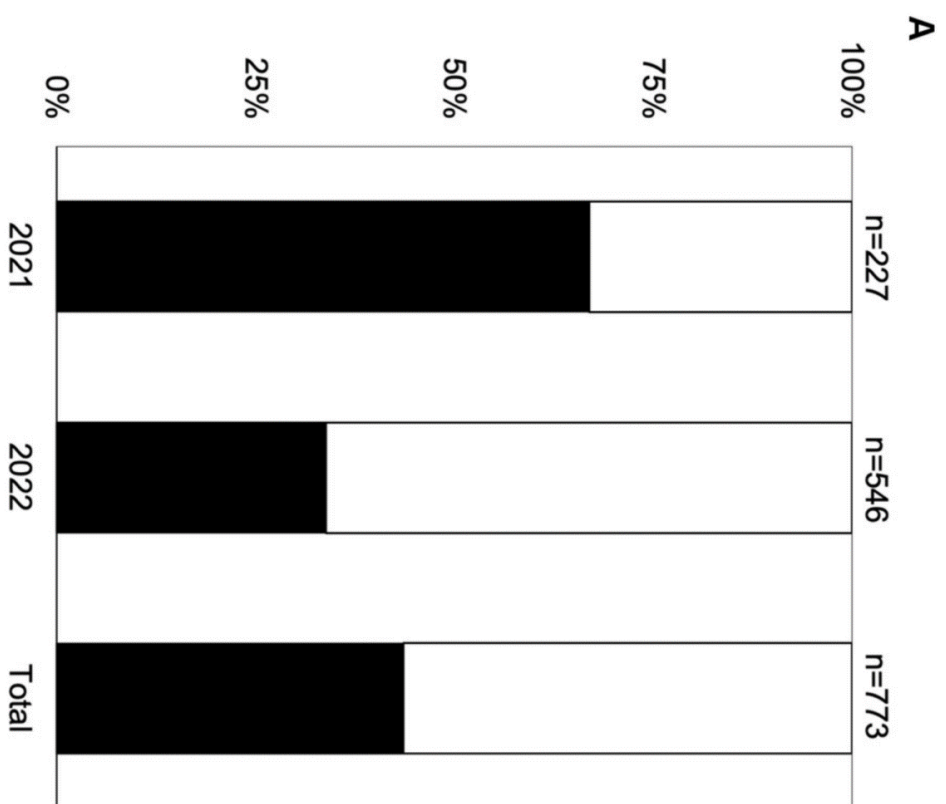


Figure 2.1. Moist feces support survival of *Salmonella* better than dry feces. (A) The distribution of moist (black) and dry (white) feces per year and in total. **(B)** Proportion of *Salmonella*-positive samples in both culture positive (left) and PCR-positive samples (right) and the number of positive samples is indicated below each pie chart. Moist feces are shown in black and dry feces shown in white.

Table 2.2. *Salmonella* prevalence increases with inclusion of molecular detection.

	Fecal Samples	Viable <i>Salmonella</i>	Prevalence (%)	Additional PCR Positive	Prevalence (%)	Total	Prevalence (%)
Year 1	227	15	6.6	35	15.4	50	22
Year 2	546	1	0.2	24	4.4	25	4.6
Total	773	16	2.1	59	7.6	75	9.7

negative samples from 2021 being PCR-positive, while 4.4% (24/545) of culture-negative samples from 2022 were PCR-positive. The proportion of PCR positive samples in moist and dry feces matched the culture data, with three quarters (74.6%, 44/59) of the PCR-positive fecal samples being moist, compared to a quarter from dry feces (25.4%, 15/59) (Figure 2.1B). Overall, *Salmonella* was significantly more likely to be detected in moist samples than dried samples [$\chi^2(1, n = 773) = 6.55, p < 0.05$].

Given the positive association between *Salmonella* presence (by culture and by PCR) and moist feces, we used a binomial generalized linear mixed model (GLMM) to explore weather factors that could influence fecal moisture. We identified four top models (i.e., $\Delta\text{AICc} < 2$) (Supplementary Table S1). Precipitation the day before sample collection was included in three of the top models and was positively associated with moist feces. During most sampling months in 2022, monthly cumulative precipitation was lower than in 2021 (Supplementary Figure S2.1), which may explain the reduced *Salmonella* detection in 2022. Humidity was also included in three models and had negative correlations with moist feces. Although temperature did not appear in our models, we expect that high temperatures would contribute to drying the feces. During May–July, the average temperatures were hotter in 2022 than in 2021 in all sampled regions, which, in combination with reduced precipitation may also contribute to the reduced *Salmonella* detection in feces in the second year of sampling.

We assessed total microbial diversity in each fecal sample ($n = 773$) by 16S rRNA sequencing, with 720 samples passing quality control. Weather variables (precipitation, temperature, humidity, and wind) did not have a strong positive or negative (± 0.30) impact to alpha diversity (data not shown). *Salmonella* was not found to affect species

richness when comparing culture-only positives or PCR positives to the *Salmonella* negative group (data not shown). Because the number of *Salmonella* culture positive samples was low, we presented the microbiome data stratified into six different groups, based on *Salmonella* status and fecal moisture. The group containing *Salmonella* culture positives from dry feces was removed from the groups, as the low number of samples ($n = 4$) reduced significant findings. Within moist feces, the Enterobacteriaceae family was significantly enriched in samples containing culturable *Salmonella* compared to samples containing only molecularly detectable *Salmonella* or no detectable *Salmonella* (Figure 2.2). This included a significant increase in the Escherichia-Shigella genera (these cannot be separated using 16S) in the *Salmonella*-culture group, rather than *Salmonella* (data not shown) (Wilcoxon rank sum test adjusted p value <0.05).

Whole genome sequencing was completed on 19 isolates (JSBird1-JSBird19) (Supplementary Table S2), and eight serovars were subsequently identified: Hadar (5 isolates), Give (4), Newport (4), Saintpaul (2), Kentucky (1), Mississippi (1), Muenchen (1), and Typhimurium (1) (Table 2.3). Using Enterobase, we next searched for related isolates. Four serovar Hadar isolates were closely related to each other (JSBird3-JSBird5 and JSBird10) and to isolates collected from ground turkey meat (within the same HeirC2 cgMLST cluster) (Supplementary Figure S2A,B). The fifth serovar Hadar isolate (JSBird11) was more closely related to an isolate from chicken meat (within the same HeirC5 cgMLST cluster) than to the other serovar Hadar isolates we isolated (Supplementary Figure S2.2C). Serovar Typhimurium and Kentucky were both isolated from the same fecal sample (F26) and both isolates were most closely related to isolates from chicken (each was within the same HeirC5 cluster of a chicken isolate). Serovar

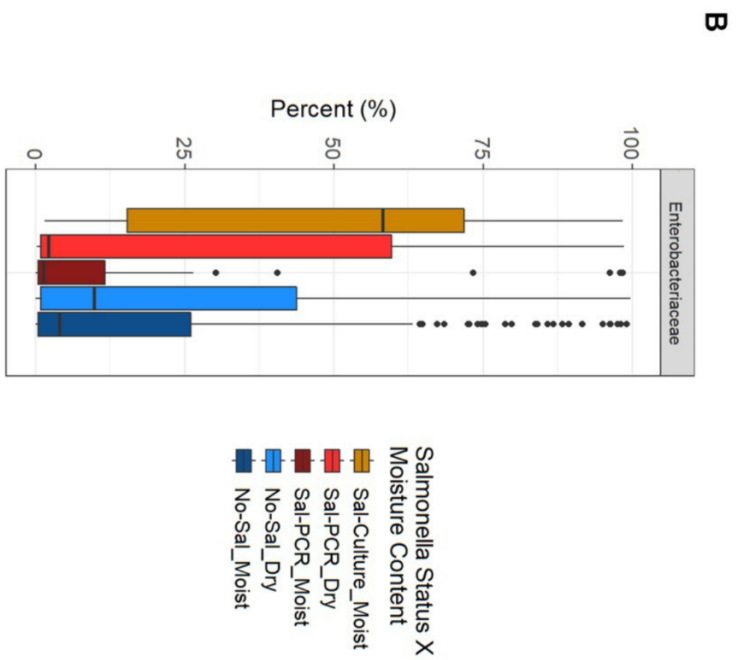
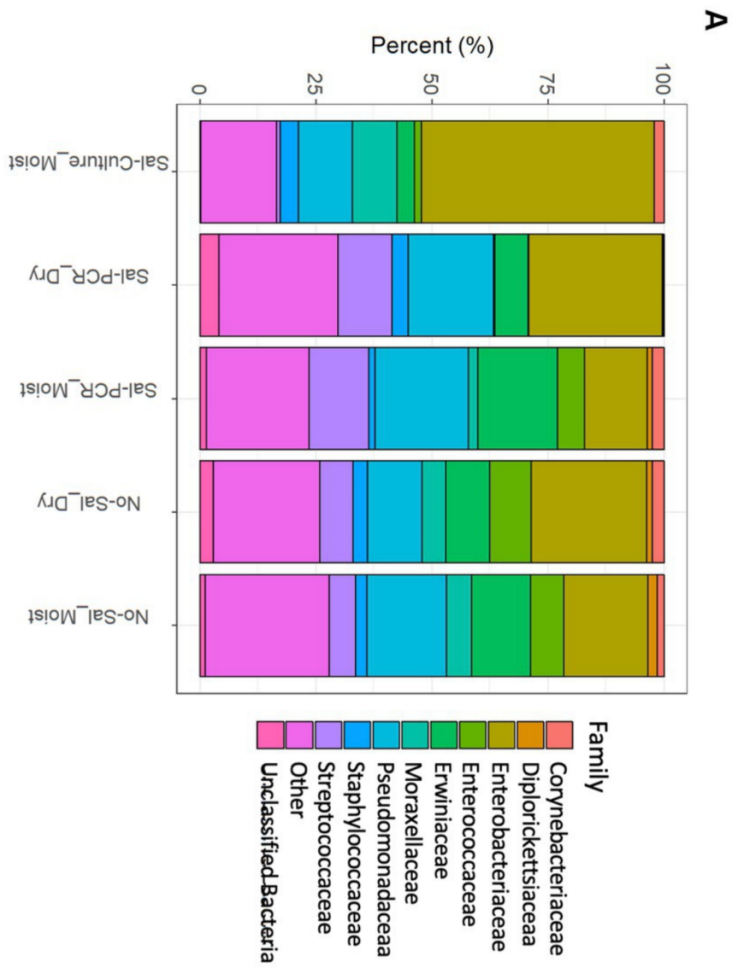


Figure 2.2. 16S sequencing of bird feces shows microbial community differences in *Salmonella* culture positive samples. **(A)** 100% bar graph of mean abundances of the 10 most prominent families identified across the entire dataset are displayed when summarized by *Salmonella* group [*Salmonella* culture positive from moist feces (Sal-Culture_Moist), *Salmonella* PCR positive from dry feces (Sal-PCR_Dry), *Salmonella* PCR positive from moist feces (Sal-PCR_Moist), no *Salmonella* from dry feces (No-Sal_Dry), and no *Salmonella* from moist feces (No-Sal_Moist)]. All taxa outside the top 10 taxa are classified as “Other.” **(B)** Differential relative abundance boxplots of prominent Enterobacteriaceae are displayed with significantly (Wilcoxon Rank Sum test, adjusted value of $p < 0.05$) different pairwise relationships displayed.

Table 2.3. Bird *Salmonella* isolates are related to isolates from a variety of sources.

Sample ID	Serovar	Farm Collected	Most closely related source type (genomic distance)	Isolate Reference
JSBird1	Typhimurium	F5	Chicken (3)	SRR10883419
JSBird2	Kentucky	F5	Chicken (3)	SRR21413100
JSBird3	Hadar	F1	Turkey (2)	SRR3664900
JSBird4	Hadar	F2	Turkey (2)	SRR3664900
JSBird5	Hadar	F2	Turkey (2)	SRR3664900
JSBird6	Give	F8	River Water (24)	SRR2050944
JSBird7	Give	F18	River Water (24)	SRR2050944
JSBird8	Give	F18	River Water (23)	SRR2050944
JSBird9	Give	F18	River Water (23)	SRR2050944
JSBird10	Hadar	F9	Turkey (2)	SRR3664900
JSBird11	Hadar	F26	Chicken (5)	SRR1122614
JSBird12	Muenchen	F7	No Similarity	N/A
JSBird13	Newport	F7	Human (8)	SRR1646204
JSBird14	Saintpaul	F7	Human (29)	SRR6231044
JSBird15	Newport	F7	Human (7)	SRR1646204
JSBird16	Mississippi	F23	Human (46)	SRR9640338
JSBird17	Newport	F7	Human (5)	SRR16925338
JSBird18	Newport	F7	Human (7)	SRR1646204
JSBird19	Saintpaul	F7	Human (30)	SRR6231044

Newport was identified four times, including two different Newport isolates from the same fecal sample (F7-5). Interestingly, while the closest whole genome match to these isolates was a single human isolate, they were also closely related to a number of serovar Newport isolates collected from surface waters in Georgia in 2011 (Supplementary Figure S2.3). For isolates belonging to serovars Give, Mississippi, Saintpaul, and Muenchen there were no other isolates in Enterobase that aligned closely, which limits assessment of potential sources for these isolates.

Deep serotyping by CRISPR-SeroSeq was performed on 14 samples. Two libraries failed to produce enough sequence reads, despite two attempts, and these both came from dry fecal samples. In total, 13 different serovars were identified (Figure 2.3). In these samples, 71% (10/14) had *Salmonella* populations consisting of multiple serovars, with an average of 2.6 serovars per sample (range, 1–7 serovars per sample). Serovars included Saintpaul (n = 6), Hadar (n = 5), Newport (n = 4), Kentucky (n = 4), Enteritidis (n = 4), Braenderup (n = 4), Give (n = 3), Rubislaw (n = 2), Heidelberg (n = 1), Infantis (n = 1), Muenchen (n = 1), Typhimurium (n = 1), and Mississippi (n = 1). Importantly, serovars Enteritidis, Infantis, and Braenderup, which were in the top 10 serovars found to cause human illness between 2019–2021 (Centers for Disease Control and Prevention, 2022), were always outnumbered by other serovars when they were present (outnumbered serovars have thinner connecting lines in Figure 2.3), and unsurprisingly, we did not isolate these by culture. In congruence with our whole genome sequence analyses, samples collected from the same sites on the same days often contained similar *Salmonella* populations. For example, two of the three fecal samples collected from farm 18 (F18-2,3) had nearly identical *Salmonella* profiles (serovars

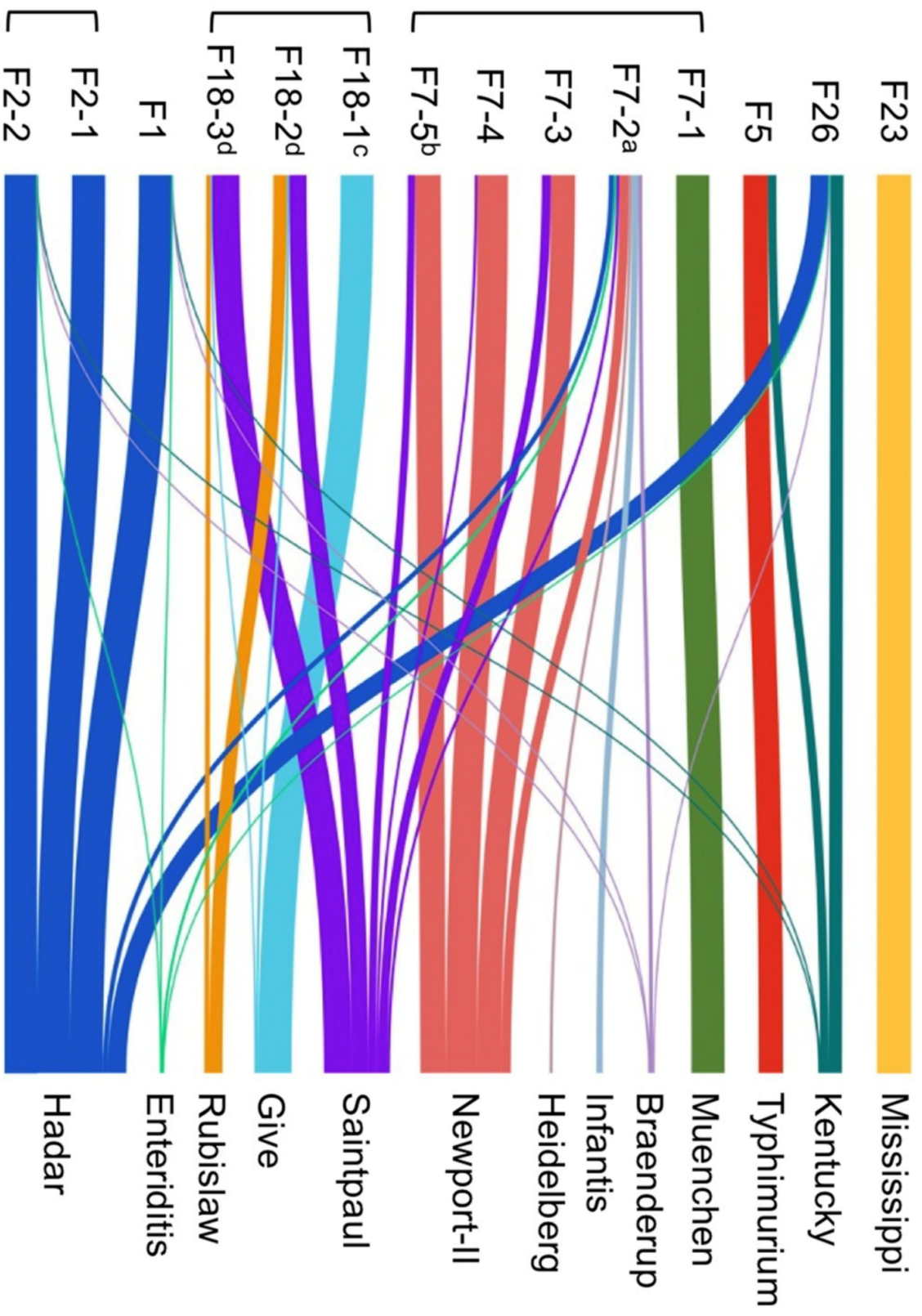


Figure 2.3. Multiserovar *Salmonella* populations exist in wild bird feces. A Sankey plot showing the sample (left nodes, indicated by the farm where the sample was collected) and the *Salmonella* serovar population within each sample. The colored bars represent different serovars (right nodes) and the thickness of the bars represent the relative abundance of each serovar within a population. Brackets around samples indicate that samples were collected from the same farm on the same day. For samples with a superscript alphabet, we were able to determine the bird species: ^achipping sparrow, ^bhouse sparrow, ^ccattle egret, and ^dfish crow.

Saintpaul, Rubislaw, and Give) with respect to the serovars that were present and their relative frequency within each sample.

Point counts were performed at each farm visit and identified 1,123 individuals. This included 51 species, with the most prevalent being the northern cardinal (*Cardinalis cardinalis*) (n = 48 visits where species was observed), the northern mockingbird (*Mimus polyglottos*) (n = 46), and the barn swallow (*Hirundo rustica*) (n = 45) (Table 2.4). A total of 31 species were observed in-field, with the most common being the song sparrow (*Melospiza melodia*) (n = 19), eastern phoebe (*Sayornis phoebe*) (n = 14), northern cardinal (n = 13), chipping sparrow (*Spizella passerine*) (n = 13), and mourning dove (*Zenaida macroura*) (n = 11). Of these, *Salmonella* was detected in four species, including three times from chipping sparrows. Off-field species included the barn swallow (n = 39), northern mockingbird (n = 38), and the Carolina wren (*Thryothorus ludovicianus*) (n = 38). Notably, some species were not often identified, but when present, were found in large numbers. For example, the rock pigeon (*Columba livia*) was only observed during four visits, but 147 individuals were recorded (Supplementary Figure S2.4). Rock pigeons were not in the top ten most frequently observed bird species across this study; however, they were the first and second highest in terms of total individuals off-field and in-field, respectively. Similarly, the common grackle (*Quiscalus quiscula*) was also observed off-field during four visits, but 83 individuals were recorded. Molecular species identification was done via PCR and sequencing of the COI gene was completed on 161 (20.8%) samples. This identified 24 species with the most common being the eastern bluebird (*Sialia sialis*) (n = 36) and the northern mockingbird (n = 19) (Table 2.4). The individuals that were culture-positive for *Salmonella* were a chipping

sparrow, an eastern bluebird, a cattle egret (*Bubulcus ibis*), a house sparrow (*Passer domesticus*), and two fish crows (*Corvus ossifragus*). Because we only identified the bird species in 13 *Salmonella*-positive fecal samples, conclusions based on the *Salmonella* status of specific bird species are limited.

2.5 Discussion

This study investigated the impact of wild birds on food safety by surveying *Salmonella* in wild bird feces deposited on foliage on produce farms over a two-year period in the Southeast. Our study demonstrated that the overall prevalence of culturable *Salmonella* in the Southeast was 2.1%, but this differed greatly between 2021 (6.6%) and 2022 (0.2%). Studies have been completed in other regions include the west coast where *Salmonella* prevalence ranged from 0.5% in cultured fecal samples (Gorski et al., 2011; Franklin and VerCauteren, 2016; Navarro-Gonzalez et al., 2020; Smith et al., 2020) to 2.5% in cultured bird gastrointestinal tracts (Kirk et al., 2002), and the Southwest where one study found a 1.9% prevalence in bird feces (Rivadeneira et al., 2016). Other studies outside of the US have included Europe (Palmgren et al., 2006; Lawson et al., 2011), South America (Cardoso et al., 2021), and the Middle East (Cohen et al., 2021). The overall prevalence identified in the current study aligns with this body of literature. Unlike most studies that sampled fresh feces (i.e., collected directly from a bird), this study offered us the opportunity to evaluate whether *Salmonella* is likely to be recovered from defecated material on foliage. *Salmonella* was isolated by culture and also detected by PCR three times more frequently in moist feces (presumably deposited within a few hours of collection) compared to dry feces. This suggests that *Salmonella* survival in feces is dynamic and the population reduces as the feces dry. While prior work has shown

Table 2.4. Bird species identified by bird counts and by molecular analysis.

Species	In field		Off field		Total species observations	Molecular observations (COI)	Salmonella culture	Salmonella PCR	Total Salmonella positive
	Number of species observations	Total number of individuals	Number of species observations	Total number of individuals					
Song sparrow (<i>Melospiza melodia</i>)	19	33	6	8	25	0	0	0	0
Eastern phoebe (<i>Sayornis phoebe</i>)	14	16	21	23	35	5	0	1	1
Northern cardinal (<i>Cardinalis cardinalis</i>)	13	25	35	41	48	14	0	1	1
Chipping sparrow (<i>Spizella passerina</i>)	13	16	22	32	35	9	1	2	3
Mourning dove (<i>Zenaidura macroura</i>)	11	67	26	44	37	9	0	1	1
Northern mockingbird (<i>Mimus polyglottos</i>)	8	13	38	47	46	19	0	0	0
Eastern bluebird (<i>Sialia sialis</i>)	7	11	15	25	22	36	1	0	1
Barn swallow (<i>Hirundo rustica</i>)	6	18	39	59	45	0	0	0	0
Killdeer (<i>Charadrius vociferus</i>)	6	6	2	4	8	0	0	0	0
Red-winged blackbird (<i>Agelaius phoeniceus</i>)	5	5	4	7	9	7	0	0	0
House finch (<i>Haemorhous mexicanus</i>)	4	8	19	56	23	11	0	1	1
Eastern meadowlark (<i>Sturnella magna</i>)	4	4	2	3	6	0	0	0	0
Indigo bunting (<i>Passerina cyanea</i>)	4	4	0	0	4	10	0	1	1
American Crow (<i>Corvus brachyrhynchos</i>)	3	8	16	29	19	12	0	0	0
European collared dove (<i>Streptopelia decaocto</i>)	3	4	2	2	5	0	0	0	0
American goldfinch (<i>Spinus tristis</i>)	3	3	4	4	7	0	0	0	0
Rock pigeon (<i>Columba livia</i>)	2	45	2	102	4	0	0	0	0
Cattle egret (<i>Bubulcus ibis</i>)	2	13	2	5	4	1	1	0	1
Carolina wren (<i>Thryothorus ludovicianus</i>)	2	3	38	46	40	0	0	0	0
Field sparrow (<i>Spizella pusilla</i>)	2	3	6	6	8	1	0	0	0
Eastern kingbird (<i>Tyrannus tyrannus</i>)	2	3	1	1	3	2	0	0	0
Blue jay (<i>Cyanocitta cristata</i>)	2	2	20	25	22	0	0	0	0
Tufted titmouse (<i>Baeolophus bicolor</i>)	2	2	9	9	11	0	0	0	0
Ruby-throated hummingbird (<i>Archilochus colubris</i>)	2	2	0	0	2	0	0	0	0
Eurasian collared dove (<i>Streptopelia decaocto</i>)	1	6	0	1	1	0	0	0	0
European starling (<i>Sturnus vulgaris</i>)	1	3	4	16	5	3	0	0	0
Common ground dove (<i>Columbina passerina</i>)	1	2	1	2	2	0	0	0	0
White-eyed vireo (<i>Vireo griseus</i>)	1	1	10	10	11	0	0	0	0
Blue grosbeak (<i>Passerina caerulea</i>)	1	1	7	7	8	7	0	0	0
Brown thrasher (<i>Toxostoma rufum</i>)	1	1	5	5	6	0	0	0	0
Downy woodpecker (<i>Picoides pubescens</i>)	1	1	5	5	6	0	0	0	0
Eastern towhee (<i>Pipilo erythrophthalmus</i>)	0	0	11	13	11	0	0	0	0
Red-bellied woodpecker (<i>Melanerpes carolinus</i>)	0	0	7	9	7	0	0	0	0
Carolina chickadee (<i>Parus carolinensis</i>)	0	0	5	12	5	0	0	0	0
Chimney swift (<i>Chaetura pelagica</i>)	0	0	5	11	5	1	0	0	0
Common grackle (<i>Quiscalus quiscula</i>)	0	0	4	83	4	0	0	0	0
Red-shouldered hawk (<i>Buteo lineatus</i>)	0	0	4	6	4	0	0	0	0
Pine warbler (<i>Setophaga pinus</i>)	0	0	3	3	3	0	0	0	0
Black vulture (<i>Coragyps atratus</i>)	0	0	2	9	2	0	0	0	0
American robin (<i>Turdus migratorius</i>)	0	0	2	6	2	0	0	0	0
American kestrel (<i>Falco sparverius</i>)	0	0	2	2	2	0	0	0	0
Brown-headed nuthatch (<i>Sitta pusilla</i>)	0	0	2	2	2	0	0	0	0
Painted bunting (<i>Passerina ciris</i>)	0	0	2	2	2	1	0	0	0
Yellow-throated vireo (<i>Vireo flavifrons</i>)	0	0	2	2	2	0	0	0	0
Northern rough-winged swallow (<i>Stelgidopteryx serripennis</i>)	0	0	1	4	1	0	0	0	0
House sparrow (<i>Passer domesticus</i>)	0	0	1	1	1	5	1	0	1
Northern parula (<i>Setophaga americana</i>)	0	0	1	1	1	0	0	0	0
Pileated woodpecker (<i>Dryocopus pileatus</i>)	0	0	1	1	1	0	0	0	0
Red-tailed hawk (<i>Buteo jamaicensis</i>)	0	0	1	1	1	0	0	0	0
Tree swallow (<i>Iridoprocne bicolor</i>)	0	0	1	1	1	0	0	0	0
White-breasted nuthatch (<i>Sitta carolinensis</i>)	0	0	1	1	1	0	0	0	0
Blue-gray gnatcatcher (<i>Poliophtila caerulea</i>)	0	0	0	0	0	1	0	0	0
Brown-headed cowbird (<i>Molothrus ater</i>)	0	0	0	0	0	2	0	0	0
Eastern wood pewee (<i>Contopus virens</i>)	0	0	0	0	0	1	0	0	0
Fish crow (<i>Corvus ossifragus</i>)	0	0	0	0	0	2	2	0	2
Great crested flycatcher (<i>Myiarchus cinerascens</i>)	0	0	0	0	0	1	0	0	0
Summer tanager (<i>Piranga rubra</i>)	0	0	0	0	0	1	0	0	0
TOTAL	146	329	419	794	565	161	6	7	13

Salmonella can survive in feces up to 291 days (Topalcengiz et al., 2020) and can have improved survival in low moisture environments (Oni et al., 2015), these studies were performed in controlled laboratory experiments and do not necessarily reflect conditions in a produce field. The fecal samples we collected had a much larger surface area to volume ratio and, therefore, are likely to dry out faster than homogenized laboratory samples.

Our statistical models suggest that precipitation the day before sampling positively influences the moisture of wild bird fecal samples, which is expected. Comparison of precipitation during sampling months in both years supports this relationship, with lower precipitation in 2022 than in 2021 likely accounting for decreased moisture and therefore a reduced *Salmonella* recovery. One model included a negative correlation between increased wind and moist feces, which is also expected as increased wind would dry the feces more rapidly. Alternatively, humidity showed a negative influence on fecal moisture in three different models. This seems counterintuitive; however one study of *Salmonella* survival in a controlled environment also saw a negative association between humidity and pathogen recovery from turkey feces (Oni et al., 2015).

Most studies of *Salmonella* in wild birds have involved capturing birds and collecting fresh feces or swabbing the cloaca (Gorski et al., 2011; Hernandez et al., 2016; Navarro-Gonzalez et al., 2020; Murray et al., 2021) while others have applied molecular techniques (i.e., PCR) to identify *Salmonella* in bird feces (Rivadeneira et al., 2016; Smith et al., 2020; Zhao et al., 2020; Olimpi et al., 2022). PCR is a very sensitive method for pathogen detection, and we detected nearly five times as many *Salmonella*-positive

fecal samples when we used PCR compared to culture. *Salmonella* has been shown to be detectable by PCR up to 10 days after inoculation; however, significant reduction occurs after four days (Lopez-Velasco et al., 2015). There are three possibilities that could explain the discrepancy between the culture results and the PCR results: (i) PCR can detect dead *Salmonella*; (ii) PCR can detect viable but non-culturable (VBNC) *Salmonella*; and (iii) because PCR is more sensitive than culture, it is possible that where the amount of *Salmonella* in the feces was very low, we were not able to recover it from culture but could detect it by PCR. We note that the background microflora was not particularly high in the selective enrichment broths nor on the XLT-4 plates, so we do not suspect that this contributed to not being able to detect *Salmonella* via our culture methods. We did attempt to serotype the PCR-detected *Salmonella* using the ISR method (Guard et al., 2022) to determine whether there were any serovar associations with PCR versus culture, but we were unsuccessful. The *Salmonella* detected by culture may have been present in higher loads, which allowed us to isolate it more easily, although we did not quantify *Salmonella*. Whether the PCR-only positive samples represent VBNC cells and pose a food safety risk should be a focus of future studies, especially as PCR-based diagnostic assays are more commonly being used to screen food products.

In addition to completing a surveillance study, this work also assessed the need for and efficacy of no-harvest buffer zones around feces in a production environment (Hamilton et al., 2015). While produce directly contacting feces cannot be harvested, the Produce Safety Rule does not require the establishment of no-harvest buffer zones, nor does it recommend suggested distances surrounding contaminated produce to exclude from harvest. Depending on the recommended buffer zone radius and the impacted

commodity type, buffer recommendations could have a substantial economic impact on growers and could be excluding produce that is safe for consumption. *Salmonella* was not isolated from additional plant samples below foliage with fecal contamination nor from neighboring plants downwind. However, depending on the weather or other climate factors, the rate at which feces dry on the plant surface may vary; this may be important to consider since our data shows that culturable *Salmonella* is primarily present in moist feces. The low incidence of *Salmonella* in bird feces and the lack of evidence supporting spread to adjacent plants in this study may be useful data for growers as they establish procedures for managing bird feces before and during harvest.

Alongside determining *Salmonella* prevalence in bird fecal samples, a deeper analysis was conducted into individual *Salmonella* isolates and serovar populations. Previous work has shown a high level of diversity within bird feces, including identifying as many as three serovars of *Salmonella* from a single sample (Antilles et al., 2021). Our culture-based analysis supported this high diversity by identifying eight serovars among 19 isolates. High-resolution analysis by deep serotyping revealed even great serovar diversity, by detecting 13 serovars across 14 samples. Further, we showed that 62.5% of culture positive samples contain multiple serovars, which included one fecal sample that contained seven different serovars (F7-2). Six serovars identified here (serovars Enteritidis, Newport, Typhimurium, Infantis, Saintpaul, and Muenchen) were determined by the Centers for Disease Control (CDC) to be among the top 10 serovars associated with human illness between 2019–2021 (Centers for Disease Control and Prevention, 2022). Additionally, serovars Hadar, Heidelberg, and Braenderup have all been linked to human outbreaks in produce or animal products in the past ten years (CDC, 2023) and

were also found in our samples. Importantly, serovars Enteritidis and Braenderup were each found in four different fecal samples and in each of these, they were significantly outnumbered by other serovars that are not known to be associated with human foodborne illness. For example, in one sample (F2-2), serovar Hadar constituted 95.9% of the total *Salmonella*, and serovar Enteritidis was only 0.1%. As our results demonstrated, using traditional culture-based *Salmonella* isolation, serovars Enteritidis and Braenderup were never detected, indicating that these important serovars were overlooked. From the five serovar Hadar isolates identified, four were closely related (within the same hierCC 2 cluster on Enterobase), to isolates from commercial turkeys (Supplementary Figure S2.1B). It should be noted that there is no commercial turkey production within at least 200 miles of the location of these farms and the turkey isolates were from 2012–2016. The fifth serovar Hadar isolate was related (within the same hierCC 5 cluster) to a chicken isolate, though that isolate was collected in 2015 from Oregon. While chicken production in the southeast is well established, further research is needed to determine whether and how wild birds acquire *Salmonella* from commercial poultry operations (e.g., from foraging on poultry farms, or from encountering contaminated poultry manure on produce farms). Interestingly, the four serovar Newport isolates most closely matched to human isolates; however, they were also closely related to isolates collected from fresh water sources in Georgia (Supplementary Figure S2.2). Two different Newport isolates came from the same fecal sample (F7-5), where one was identified from each selective enrichment broth, indicating there is also strain diversity within single fecal samples. Deep serotyping showed that samples collected from the same farm often had similar *Salmonella* serovar populations in addition to closely related isolates, suggesting that

similar sources of *Salmonella* may occur in the environment that contribute to contamination in wild birds, or that a single bird was defecating multiple times in the same field. Alternatively, for birds that flock together (e.g., crows), this similarity may reflect transmission within a flock, for example at common feeding or watering locations. Overall, our findings indicate that wild birds have the potential to obtain and transmit *Salmonella* from a wide range of sources over large geographic areas.

Bird species were identified in this study using both physical and molecular methods. Other studies have used a more direct collection approach where birds are caught using nets or traps followed by the collection of feces or swabbing the cloaca (Gruszynski et al., 2014; Fuentes-Castillo et al., 2019; Navarro-Gonzalez et al., 2020). In these instances, bird species can be identified quite easily, and the sample is fresher. Alternatively, the collection method used in this study resulted in lower molecular characterization of bird species (20.8% identified). However, it was non-invasive and provided an opportunity to investigate bird species actively defecating on the field, not just those primarily foraging in adjacent habitats. Data collected in this study identified 51 species of bird from point counts and 24 species from COI, for a total of 57 species. Molecular detection from bird feces allowed for the identification of six additional species, including the fish crow, which was identified in two *Salmonella* culture positive samples but not identified during point counts. This demonstrates the importance of the two complementary methods for bird identification.

We categorized birds from our point counts as in-field or off-field. The off-field category included species that are often associated with agricultural structures (e.g., barns, packing houses, fences) or other structures (e.g., powerlines adjacent to the farm),

such as the barn swallow (n = 39 species observations off-field), house finch (*Haemorrhous mexicanus*) (n = 19), European starling (*Sturnus vulgaris*) (n = 4), and rock pigeon (n = 2). Of these, only a single house finch fecal sample tested positive for *Salmonella*. Although rock pigeons were observed twice off-field, the total number of individuals was 102, suggesting that flock size may also be relevant with respect to understanding the risk posed by different species. The off-field category also included birds found away from the farm premises (e.g., in tree line, neighboring pasture) and included the Carolina wren (n = 26), woodpecker (n = 12), white-eyed vireo (*Vireo griseus*) (n = 8), and eastern towhee (*Pipilo erythrophthalmus*) (n = 3). This latter group poses the lowest risk of pathogen transmission because they are infrequently observed interacting with produce. Although we were not able to identify the bird species for the majority of *Salmonella*-positive fecal samples, of the ones we were able to identify, none belonged to this category. Conversely, birds on agricultural structures and in-field pose a higher food safety risk because of their interactions with farm livestock and produce, so deterrents targeting these species would be more effective. Three-quarters (10/13) of *Salmonella*-positive fecal samples were from birds that were also observed as in-field during point counts. For the three that were not observed, one was a cattle egret and the other two were fish crows. Interestingly, fish crows are associated with water and both had *Salmonella* serovars Give and Rubislaw (F18-2, -3), which are associated with surface water (Haley et al., 2009; Gorski et al., 2011; McEgan et al., 2014; Maurer et al., 2015; Callahan et al., 2019; Deaven et al., 2021).

Mitigating risks associated with wild birds in produce fields remains a complicated issue that will require a One Health approach to fully understand how the

interaction of animals (including wildlife, such as birds, and food animals), the environment, and human activity contribute to *Salmonella* ecology. In this work, we found a low prevalence of *Salmonella*, however; serovars associated with human illness were often identified when *Salmonella* was present. Moreover, the prevalence increased from 2% to over 9% when molecular detection was included, suggesting that different methods of detection can influence the establishment of risk due to this environmental source of *Salmonella*. The complexity of this problem is highlighted by our whole genome analysis showing that *Salmonella* isolates recovered in this study were related within 10 pairwise allelic differences (PADs) to isolates from a range of sources including humans, animal agriculture, and the environment, as well as some without any links to these sources. The freshness of the wild bird feces was shown to impact viability of *Salmonella*; however, more work will need to be completed to show how risk of feces changes with time and if certain serovars are better adapted to this environment. While factors affecting the prevalence of *Salmonella* within wild birds and *Salmonella* survival within feces are not fully understood, the findings presented here contribute to our understanding of these complex food safety systems.

2.6 Data availability statement

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found in the article.

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2.9 Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

2.10 Author Contributions

JCS: Data curation, Formal analysis, Investigation, Writing – original draft, Writing – review & editing, Visualization. SV: Data curation, Formal analysis, Investigation, Writing – review & editing. KR: Resources, Writing – review & editing. ZS: Resources, Writing – review & editing. JLD: Resources, Writing – review & editing. JS: Resources, Writing – review & editing. LLD: Conceptualization, Funding acquisition, Investigation, Methodology, Project administration, Resources, Supervision, Writing – review & editing. WES: Conceptualization, Funding acquisition, Investigation, Methodology, Project administration, Supervision, Writing – review & editing. NWS: Conceptualization, Data curation, Formal analysis, Funding acquisition, Investigation, Methodology, Project administration, Supervision, Writing – original draft, Writing – review & editing.

2.11 References

Alegbeleye, O. O., Singleton, I., and Sant’Ana, A. S. (2018). Sources and contamination routes of microbial pathogens to fresh produce during field cultivation: A review. *Food Microbiol.* 73, 177–208. <https://doi.org/10.1016/j.fm.2018.01.003>

- Antilles, N., García-Bocanegra, I., Alba-Casals, A., López-Soria, S., Pérez-Méndez, N., Saco, M., González-Solís, J., and Cerdà-Cuéllar, M. (2021). Occurrence and antimicrobial resistance of zoonotic enteropathogens in gulls from southern Europe. *Sci. Total Environ.* 763, 143018.
<https://doi.org/10.1016/j.scitotenv.2020.143018>
- Bankevich, A., Nurk, S., Antipov, D., Gurevich, A. A., Dvorkin, M., Kulikov, A. S., Lesin, V. M., Nikolenko, S. I., Pham, S., Prjibelski, A. D., Pyshkin, A. V., Sirotkin, A. V., Vyahhi, N., Tesler, G., Alekseyev, M. A., and Pevzner, P. A. (2012). SPAdes: A New Genome Assembly Algorithm and Its Applications to Single-Cell Sequencing. *J. Comput. Biol.* 19(5), 455–477.
<https://doi.org/10.1089/cmb.2012.0021>
- Bardsley, C. A., Weller, D. L., Ingram, D. T., Chen, Y., Oryang, D., Rideout, S. L., and Strawn, L. K. (2021). Strain, Soil-Type, Irrigation Regimen, and Poultry Litter Influence *Salmonella* Survival and Die-off in Agricultural Soils. *Front. Microbiol.* 12. <https://doi.org/10.3389/fmicb.2021.590303>
- Bolyen, E., Rideout, J. R., Dillon, M. R., Bokulich, N. A., Abnet, C. C., Al-Ghalith, G. A., Alexander, H., Alm, E. J., Arumugam, M., Asnicar, F., Bai, Y., Bisanz, J. E., Bittinger, K., Brejnrod, A., Brislawn, C. J., Brown, C. T., Callahan, B. J., Caraballo-Rodríguez, A. M., Chase, J., ... Caporaso, J. G. (2019). Reproducible, interactive, scalable and extensible microbiome data science using QIIME 2. *Nat. Biotechnol.* 37(8), 852–857. <https://doi.org/10.1038/s41587-019-0209-9>
- Brooks, M., Bolker, B., Kristensen, K., Maechler, M., Skaug, H., Nielson, A., Berg, C., and Van Benthem, K. (2023). Generalized Linear Mixed Models using Template

- Model Builder (1.1.7). (Package ‘glmmTMB’). <https://cran.r-project.org/web/packages/glmmTMB/index.html> [Accessed March 1, 2023]
- Bruce B, Hartley C, Rose EB, Tierney R, Bazaco M, Parker CC, Waldrop C, Cote A, Golden N, Iverson SA, Griffin PM, Tauxe RV, Batz M, Harris S, Hale KR, and Kufel JZ. (2022). Foodborne illness source attribution estimates for 2020 for *Salmonella*, *Escherichia coli* O157, and *Listeria monocytogenes* using multi-year outbreak surveillance data, United States. <https://www.cdc.gov/foodsafety/ifsac/annual-reports.html> [Accessed March 12, 2023]
- Burnham, K. P., and Anderson, D. R. (2002). Model selection and multimodel inferences: A practical information-theoretic approach (2nd ed.). Springer-Verlag.
- Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., and Holmes, S. P. (2016). DADA2: High-resolution sample inference from Illumina amplicon data. *Nat. Methods*. 13(7), 581–583. <https://doi.org/10.1038/nmeth.3869>
- Callahan, M. T., Van Kessel, J. A., and Micallef, S. A. (2019). *Salmonella enterica* recovery from river waters of the Maryland Eastern Shore reveals high serotype diversity and some multidrug resistance. *Environ. Res.* 168, 7–13. <https://doi.org/10.1016/j.envres.2018.09.012>
- Cardoso, M. D., Santos, A. F. D. M., Rodrigues, M. D. S., Pribul, B. R., Grael, A. S., Pedroso, V. M., Pires, J. R., Travassos, C. E. P. F., Domit, C., Vieira-Da-Motta, O., Rodrigues, D. D. P., and Siciliano, S. (2021). *Salmonella* spp. profiles isolated from seabird samples from the Brazilian coast. *Prev. Vet. Med.* 193, 105413. <https://doi.org/10.1016/j.prevetmed.2021.105413>

- Cason J., Cox N., Buhr R., Bourassa D., and Richardson L. (2011). Probability of identifying different *Salmonella* serotypes in poultry samples. *Int. Poult. Sci. Forum.* 75–76.
- CDC. (2023). *Reports of Selected Salmonella Outbreak Investigations*.
<https://www.cdc.gov/Salmonella/outbreaks.html> [Accessed March 9, 2023]
- Centers for Disease Control and Prevention. (2022). *FoodNet Fast*. Pathogen Surveillance. <https://wwwn.cdc.gov/foodnetfast/> [Accessed March 9, 2023]
- Cheng, R. A., Eade, C. R., and Wiedmann, M. (2019). Embracing Diversity: Differences in Virulence Mechanisms, Disease Severity, and Host Adaptations Contribute to the Success of Nontyphoidal *Salmonella* as a Foodborne Pathogen. *Front. Microbiol.* 10, 1368. <https://doi.org/10.3389/fmicb.2019.01368>
- Cohen, E., Azriel, S., Auster, O., Gal, A., Zitronblat, C., Mikhlin, S., Scharte, F., Hensel, M., Rahav, G., and Gal-Mor, O. (2021). Pathoadaptation of the passerine-associated *Salmonella enterica* serovar Typhimurium lineage to the avian host. *PLOS Pathog.* 17(3), e1009451. <https://doi.org/10.1371/journal.ppat.1009451>
- Critzer, F. J., and Doyle, M. P. (2010). Microbial ecology of foodborne pathogens associated with produce. *Curr. Opin. Biotechnol.* 21(2), 125–130.
<https://doi.org/10.1016/j.copbio.2010.01.006>
- CroplandCROS. (2022). National Agricultural Statistics Service.
<https://croplandcros.scinet.usda.gov/> [Accessed March 16, 2023]
- Deaven, A. M., Ferreira, C. M., Reed, E. A., Chen See, J. R., Lee, N. A., Almaraz, E., Rios, P. C., Marogi, J. G., Lamendella, R., Zheng, J., Bell, R. L., and Shariat, N. W. (2021). *Salmonella* Genomics and Population Analyses Reveal High Inter-

- and Intraserovar Diversity in Freshwater. *Appl. Environ. Microbiol.* 87(6).
<https://doi.org/10.1128/AEM.02594-20>
- Devarajan, N., McGarvey, J. A., Scow, K., Jones, M. S., Lee, S., Samaddar, S., Schmidt, R., Tran, T. D., and Karp, D. S. (2021). Cascading effects of composts and cover crops on soil chemistry, bacterial communities and the survival of foodborne pathogens. *J. Appl. Microbiol.* 131(4), 1564–1577.
<https://doi.org/10.1111/jam.15054>
- Devarajan, N., Weller, D. L., Jones, M., Adell, A. D., Adhikari, A., Allende, A., Arnold, N. L., Baur, P., Beno, S. M., Clements, D., Olimpi, E. M., Critzer, F., Green, H., Gorski, L., Ferelli Gruber, A., Kovac, J., McGarvey, J., Murphy, C. M., Murphy, S. I., ... Karp, D. S. (2023). Evidence for the efficacy of pre-harvest agricultural practices in mitigating food-safety risks to fresh produce in North America. *Front. Sustain. Food Syst.* 7, 1101435. <https://doi.org/10.3389/fsufs.2023.1101435>
- Elsohaby, I., Samy, A., Elmoslemany, A., Alorabi, M., Alkafafy, M., Aldoweriej, A., Al-Marri, T., Elbehiry, A., and Fayez, M. (2021). Migratory Wild Birds as a Potential Disseminator of Antimicrobial-Resistant Bacteria around Al-Asfar Lake, Eastern Saudi Arabia. *Antibiotics*, 10(3), 260. <https://doi.org/10.3390/antibiotics10030260>
- Franklin, A. B., and VerCauteren, K. C. (2016). Keeping Wildlife Out of Your Food: Mitigation and Control Strategies to Reduce the Transmission Risk of Food-Borne Pathogens. Nebraska: Springer International Publishing
- Fuentes-Castillo, D., Farfán-López, M., Esposito, F., Moura, Q., Fernandes, M. R., Lopes, R., Cardoso, B., Muñoz, M. E., Cerdeira, L., Najle, I., Muñoz, P. M., Catão-Dias, J. L., González-Acuña, D., and Lincopan, N. (2019). Wild owls

colonized by international clones of extended-spectrum β -lactamase (CTX-M)-producing *Escherichia coli* and *Salmonella* Infantis in the Southern Cone of America. *Sci. Total Environ.* 674, 554–562.

<https://doi.org/10.1016/j.scitotenv.2019.04.149>

Gorski, L., Parker, C. T., Liang, A., Cooley, M. B., Jay-Russell, M. T., Gordus, A. G., Atwill, E. R., and Mandrell, R. E. (2011). Prevalence, Distribution, and Diversity of *Salmonella enterica* in a Major Produce Region of California. *Appl. Environ. Microbiol.* 77(8), 2734–2748. <https://doi.org/10.1128/AEM.02321-10>

Grimont, P. A. D., and Weill, F.-X. (n.d.). *WHO Collaborating Centre for Reference and Research on Salmonella Antigenic Formulae of the Salmonella Serovars 2007 9th edition.*

Gruszynski, K., Pao, S., Kim, C., Toney, D., Wright, K., Ross, P. G., Colon, A., and Levine, S. (2014). Evaluating Wildlife as a Potential Source of *Salmonella* serotype Newport (JJPX01.0061) Contamination for Tomatoes on the Eastern Shore of Virginia. *Zoonoses Public Health*, 61(3), 202–207.
<https://doi.org/10.1111/zph.12061>

Guard, J., Jones, D. R., Gast, R. K., Garcia, J. S., and Rothrock, M. J. (2022). Serotype Screening of *Salmonella enterica* Subspecies I by Intergenic Sequence Ribotyping (ISR): Critical Updates. *Microorganisms*, 11(1), 97.
<https://doi.org/10.3390/microorganisms11010097>

Haley, B. J., Cole, D. J., and Lipp, E. K. (2009). Distribution, Diversity, and Seasonality of Waterborne *Salmonellae* in a Rural Watershed. *Appl. Environ. Microbiol.* 75(5), 1248–1255. <https://doi.org/10.1128/AEM.01648-08>

- Hamilton, K. E., Umber, J., Hultberg, A., Tong, C., Schermann, M., Diez-Gonzalez, F., and Bender, J. B. (2015). Validation of Good Agricultural Practices (GAP) on Minnesota Vegetable Farms. *Foodborne Pathogens and Disease*, 12(2), 145–150. <https://doi.org/10.1089/fpd.2014.1817>
- Hartig, F., and Lohse, L. (2022). *DHARMa*: Residual Diagnostics for Hierarchical (Multi-Level / Mixed) Regression Models (0.4.6). <http://florianhartig.github.io/DHARMa/>
- Hebert, P. D. N., Cywinska, A., Ball, S. L., and deWaard, J. R. (2003). Biological identifications through DNA barcodes. *Proc. R. Soc. Lond. Series B: Biol. Sci.* 270(1512), 313–321. <https://doi.org/10.1098/rspb.2002.2218>
- Hernandez, S. M., Keel, K., Sanchez, S., Trees, E., Gerner-Smidt, P., Adams, J. K., Cheng, Y., Ray, A., Martin, G., Presotto, A., Ruder, M. G., Brown, J., Blehert, D. S., Cottrell, W., and Maurer, J. J. (2012). Epidemiology of a *Salmonella enterica* subsp. *enterica* Serovar Typhimurium Strain Associated with a Songbird Outbreak. *Appl. Environ. Microbiol.* 78(20), 7290–7298. <https://doi.org/10.1128/AEM.01408-12>
- Hernandez, S. M., Welch, C. N., Peters, V. E., Lipp, E. K., Curry, S., Yabsley, M. J., Sanchez, S., Presotto, A., Gerner-Smidt, P., Hise, K. B., Hammond, E., Kistler, W. M., Madden, M., Conway, A. L., Kwan, T., and Maurer, J. J. (2016). Urbanized White Ibises (*Eudocimus albus*) as Carriers of *Salmonella enterica* of Significance to Public Health and Wildlife. *PLoS ONE*, 11(10), e0164402. <https://doi.org/10.1371/journal.pone.0164402>

- Issenhuth-Jeanjean, S., Roggentin, P., Mikoleit, M., Guibourdenche, M., de Pinna, E., Nair, S., Fields, P. I., and Weill, F.-X. (2014). Supplement 2008–2010 (no. 48) to the White–Kauffmann–Le Minor scheme. *Res. Microbiol.* 165(7), 526–530. <https://doi.org/10.1016/j.resmic.2014.07.004>
- Ivanova, N. V., Zemlak, T. S., Hanner, R. H., and Hebert, P. D. N. (2007). Universal primer cocktails for fish DNA barcoding. *Mol. Ecol. Notes.* 7(4), 544–548. <https://doi.org/10.1111/j.1471-8286.2007.01748.x>
- Joo, S., and Park, S. (2012). Identification of bird species and their prey using DNA barcode on feces from Korean traditional village groves and forests (*maeulsoop*). *Anim. Cells Syst.* 16(6), 488–497. <https://doi.org/10.1080/19768354.2012.720939>
- Karp, D. S., Mendenhall, C. D., Sandí, R. F., Chaumont, N., Ehrlich, P. R., Hadly, E. A., and Daily, G. C. (2013). Forest bolsters bird abundance, pest control and coffee yield. *Ecol. Lett.* 16(11), 1339–1347. <https://doi.org/10.1111/ele.12173>
- Katoh, K., and Standley, D. M. (2013). MAFFT Multiple Sequence Alignment Software Version 7: Improvements in Performance and Usability. *Mol. Biol. Evol.* 30(4), 772–780. <https://doi.org/10.1093/molbev/mst010>
- Kerr, K. C. R., Stoeckle, M. Y., Dove, C. J., Weigh, L. A., Francis, C. M., and Hebert, P. D. N. (2007). Comprehensive DNA barcode coverage of North American birds. *Mol. Ecol. Notes.* 7(4), 535–543. <https://doi.org/10.1111/j.1471-8286.2007.01670.x>
- Kirk, J. H., Holmberg, C. A., and Jeffrey, J. S. (2002). Prevalence of *Salmonella* spp in selected birds captured on California dairies. *J. Am. Vet. Med. Assoc.* 220(3), 359–362. <https://doi.org/10.2460/javma.2002.220.359>

- Lawson, B., Hughes, L. A., Peters, T., de Pinna, E., John, S. K., Macgregor, S. K., and Cunningham, A. A. (2011). Pulsed-Field Gel Electrophoresis Supports the Presence of Host-Adapted *Salmonella enterica subsp. enterica* Serovar Typhimurium Strains in the British Garden Bird Population. *Appl. Environ. Microbiol* 77(22), 8139–8144. <https://doi.org/10.1128/AEM.00131-11>
- Lopez-Velasco, G., Tomas-Callejas, A., Sbodio, A. O., Pham, X., Wei, P., Diribsa, D., and Suslow, T. V. (2015). Factors affecting cell population density during enrichment and subsequent molecular detection of *Salmonella enterica* and *Escherichia coli* O157:H7 on lettuce contaminated during field production. *Food Control*. 54, 165–175. <https://doi.org/10.1016/j.foodcont.2015.01.041>
- Lüdecke, D., Ben-Shachar, M., Patil, I., Waggoner, P., and Makowski, D. (2021). performance: An R Package for Assessment, Comparison and Testing of Statistical Models. *J. Open Source Softw.* 6(60), 3139. <https://doi.org/10.21105/joss.03139>
- Maurer, J. J., Martin, G., Hernandez, S., Cheng, Y., Gerner-Smidt, P., Hise, K. B., Tobin D’Angelo, M., Cole, D., Sanchez, S., Madden, M., Valeika, S., Presotto, A., and Lipp, E. K. (2015). Diversity and Persistence of *Salmonella enterica* Strains in Rural Landscapes in the Southeastern United States. *PLoS ONE*, 10(7), e0128937. <https://doi.org/10.1371/journal.pone.0128937>
- McEgan, R., Chandler, J. C., Goodridge, L. D., and Danyluk, M. D. (2014). Diversity of *Salmonella* Isolates from Central Florida Surface Waters. *Appl. Environ. Microbiol* 80(21), 6819–6827. <https://doi.org/10.1128/AEM.02191-14>

- Meiklejohn, K. A., Damaso, N., and Robertson, J. M. (2019). Assessment of BOLD and GenBank – Their accuracy and reliability for the identification of biological materials. *PLoS ONE*, *14*(6), e0217084.
<https://doi.org/10.1371/journal.pone.0217084>
- Murray, M. H., Hernandez, S. M., Rozier, R. S., Kidd, A. D., Hepinstall-Cymerman, J., Curry, S. E., Yabsley, M. J., Adams, H., Ellison, T., Welch, C. N., and Lipp, E. K. (2021). Site Fidelity is Associated with Food Provisioning and *Salmonella* in an Urban Wading Bird. *EcoHealth*, *18*(3), 345–358. <https://doi.org/10.1007/s10393-021-01543-x>
- Navarro-Gonzalez, N., Wright, S., Aminabadi, P., Gwinn, A., Suslow, T. V., and Jay-Russell, M. T. (2020). Carriage and Subtypes of Foodborne Pathogens Identified in Wild Birds Residing near Agricultural Lands in California: a Repeated Cross-Sectional Study. *Appl. Environ. Microbiol* *86*(3).
<https://doi.org/10.1128/AEM.01678-19>
- Obe, T., Siceloff, A. T., Crowe, M. G., Scott, H. M., and Shariat, N. W. (2023). Combined Quantification and Deep Serotyping for *Salmonella* Risk Profiling in Broiler Flocks. *Appl. Environ. Microbiol.* *89*(4).
<https://doi.org/10.1128/aem.02035-22>
- Olimpi, E. M., Garcia, K., Gonthier, D. J., Kremen, C., Snyder, W. E., Wilson-Rankin, E. E., and Karp, D. S. (2022). Semi-natural habitat surrounding farms promotes multifunctionality in avian ecosystem services. *J. Appl. Ecol.* *59*(4), 898–908.
<https://doi.org/10.1111/1365-2664.14124>

- Oni, R. A., Sharma, M., and Buchanan, R. L. (2015). Survival of *Salmonella enterica* in Dried Turkey Manure and Persistence on Spinach Leaves. *J. Food Prot.* 78(10), 1791–1799. <https://doi.org/10.4315/0362-028X.JFP-15-047>
- Palmgren, H., Aspan, A., Broman, T., Bengtsson, K., Blomquist, L., Bergström, S., Sellin, M., Wollin, R., and Olsen, B. (2006). *Salmonella* in Black-headed gulls (*Larus ridibundus*); prevalence, genotypes and influence on *Salmonella* epidemiology. *Epidemiol. Infect.* 134(3), 635–644. <https://doi.org/10.1017/S0950268805005261>
- Patel, S., Waugh, J., Millar, C. D., and Lambert, D. M. (2010). Conserved primers for DNA barcoding historical and modern samples from New Zealand and Antarctic birds. *Mol. Ecol. Resour.* 10(3), 431–438. <https://doi.org/10.1111/j.1755-0998.2009.02793.x>
- Price, M. N., Dehal, P. S., and Arkin, A. P. (2010). FastTree 2 – Approximately Maximum-Likelihood Trees for Large Alignments. *PLoS ONE*, 5(3), e9490. <https://doi.org/10.1371/journal.pone.0009490>
- Prosser, D. J., Cui, P., Takekawa, J. Y., Tang, M., Hou, Y., Collins, B. M., Yan, B., Hill, N. J., Li, T., Li, Y., Lei, F., Guo, S., Xing, Z., He, Y., Zhou, Y., Douglas, D. C., Perry, W. M., and Newman, S. H. (2011). Wild Bird Migration across the Qinghai-Tibetan Plateau: A Transmission Route for Highly Pathogenic H5N1. *PLoS ONE*, 6(3), e17622. <https://doi.org/10.1371/journal.pone.0017622>
- Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., Peplies, J., and Glöckner, F. O. (2012). The SILVA ribosomal RNA gene database project:

- improved data processing and web-based tools. *Nucleic Acids Res.* 41(D1), D590–D596. <https://doi.org/10.1093/nar/gks1219>
- Rahn, K., De Grandis, S. A., Clarke, R. C., McEwen, S. A., Galán, J. E., Ginocchio, C., Curtiss, R., and Gyles, C. L. (1992). Amplification of an *invA* gene sequence of *Salmonella* Typhimurium by polymerase chain reaction as a specific method of detection of *Salmonella*. *Mol. Cell. Probes.* 6(4), 271–279. [https://doi.org/10.1016/0890-8508\(92\)90002-F](https://doi.org/10.1016/0890-8508(92)90002-F)
- Reddy, S. P., Wang, H., Adams, J. K., and Feng, P. C. H. (2016). Prevalence and Characteristics of *Salmonella* Serotypes Isolated from Fresh Produce Marketed in the United States. *J. Food Prot.* 79(1), 6–16. <https://doi.org/10.4315/0362-028X.JFP-15-274>
- Rivadeneira, P., Hilson, C., Justice-Allen, A., and Jay-Russell, M. (2016). Pathogen Risks Related to the Movement of Birds Frequenting Livestock and Fresh Produce Growing Areas in the Southwestern U.S. *Vertebr. Pest Conf.* 27, 258–263. <https://doi.org/10.5070/V427110583>
- Rivadeneira, P., Kross, S., Navarro-Gonzalez, N., and Jay-Russell, M. (2018). A Review of Bird Deterrents Used in Agriculture. *Vertebr. Pest Conf.* 28, 218–223. <https://doi.org/10.5070/V42811040>
- Rodrigues, C., da Silva, A. L. B. R., and Dunn, L. L. (2019). Factors Impacting the Prevalence of Foodborne Pathogens in Agricultural Water Sources in the Southeastern United States. *Water*, 12(1), 51. <https://doi.org/10.3390/w12010051>

- Rosenstraus M, Wang Z, Chang S-Y, Debonville D, Spadaro JP. 1998. An internal control for routine diagnostic PCR: design, properties, and effect on clinical performance. *J Clin Microbiol* 36:191–197.
- Scallan, E., Hoekstra, R. M., Angulo, F. J., Tauxe, R. V., Widdowson, M.-A., Roy, S. L., Jones, J. L., and Griffin, P. M. (2011). Foodborne Illness Acquired in the United States—Major Pathogens. *Emerg. Infect. Dis.* 17(1), 7–15.
<https://doi.org/10.3201/eid1701.P11101>
- Segata, N., Izard, J., Waldron, L., Gevers, D., Miropolsky, L., Garrett, W. S., and Huttenhower, C. (2011). Metagenomic biomarker discovery and explanation. *Genome Biol.* 12(6), R60. <https://doi.org/10.1186/gb-2011-12-6-r60>
- Siceloff, A. T., Ohta, N., Norman, K. N., Loneragan, G. H., Norby, B., Scott, H. M., and Shariat, N. W. (2021). Antimicrobial Resistance Hidden within Multiseroovar *Salmonella* Populations. *Antimicrob. Agents Chemother.* 65(6), e00048-21.
<https://doi.org/10.1128/AAC.00048-21>
- Siceloff, A. T., Waltman, D., and Shariat, N. W. (2022). Regional *Salmonella* Differences in United States Broiler Production from 2016 to 2020 and the Contribution of Multiseroovar Populations to *Salmonella* Surveillance. *Appl. Environ. Microbiol.* 88(8). <https://doi.org/10.1128/aem.00204-22>
- Smith, O. M., Edworthy, A., Taylor, J. M., Jones, M. S., Tormanen, A., Kennedy, C. M., Fu, Z., Latimer, C. E., Cornell, K. A., Michelotti, L. A., Sato, C., Northfield, T., Snyder, W. E., and Owen, J. P. (2020). Agricultural intensification heightens food safety risks posed by wild birds. *J. Appl. Ecol.* 57(11).
<https://doi.org/10.1111/1365-2664.13723>

- Strawn, L. K., Gröhn, Y. T., Warchocki, S., Worobo, R. W., Bihn, E. A., and Wiedmann, M. (2013). Risk Factors Associated with *Salmonella* and *Listeria monocytogenes* Contamination of Produce Fields. *Appl. Environ. Microbiol.* 79(24), 7618–7627. <https://doi.org/10.1128/AEM.02831-13>
- Tack, D. M., Marder, E. P., Griffin, P. M., Cieslak, P. R., Dunn, J., Hurd, S., Scallan, E., Lathrop, S., Muse, A., Ryan, P., Smith, K., Tobin-D’Angelo, M., Vugia, D. J., Holt, K. G., Wolpert, B. J., Tauxe, R., and Geissler, A. L. (2019). Preliminary incidence and trends of infections with pathogens transmitted commonly through food — Foodborne Diseases Active Surveillance Network, 10 U.S. sites, 2015–2018. *Am. J. Transplant.* 19(6), 1859–1863. <https://doi.org/10.1111/ajt.15412>
- Thompson, C. P., Doak, A. N., Amirani, N., Schroeder, E. A., Wright, J., Kariyawasam, S., Lamendella, R., and Shariat, N. W. (2018). High-Resolution Identification of Multiple *Salmonella* Serovars in a Single Sample by Using CRISPR-SeroSeq. *Appl. Environ. Microbiol.* 84(21). <https://doi.org/10.1128/AEM.01859-18>
- Tizard, I. (2004). Salmonellosis in wild birds. *Semin. avian exot. pet med.* 13(2), 50–66. <https://doi.org/10.1053/j.saep.2004.01.008>
- Topalcengiz, Z., Spanninger, P. M., Jearnsripong, S., Persad, A. K., Buchanan, R. L., Saha, J., LeJeune, J., Jay-Russell, M. T., Kniel, K. E., and Danyluk, M. D. (2020). Survival of *Salmonella* in Various Wild Animal Feces That May Contaminate Produce. *J. Food Prot.* 83(4), 651–660. <https://doi.org/10.4315/0362-028X.JFP-19-302>
- Produce Safety Rule, 74354 (2015) (testimony of U.S. Food and Drug Administration). <https://www.federalregister.gov/documents/2015/11/27/2015-28159/standards->

[for-the-growing-harvesting-packing-and-holding-of-produce-for-human-consumption](#) [Accessed March 10, 2023]

- Uzzau, S., Brown, D. J., Wallis, T., Rubino, S., Leori, G., Bernard, S., Casadesus, J., Platt, D. J., and Olsen, J. E. (2000). Host adapted serotypes of *Salmonella enterica*. *Epidemiol. Infect.* 125(2), 229–255.
<https://doi.org/10.1017/S0950268899004379>
- Walters, W., Hyde, E. R., Berg-Lyons, D., Ackermann, G., Humphrey, G., Parada, A., Gilbert, J. A., Jansson, J. K., Caporaso, J. G., Fuhrman, J. A., Apprill, A., and Knight, R. (2016). Improved Bacterial 16S rRNA Gene (V4 and V4-5) and Fungal Internal Transcribed Spacer Marker Gene Primers for Microbial Community Surveys. *MSystems*, 1(1). <https://doi.org/10.1128/mSystems.00009-15>
- Zhang, S., den Bakker, H. C., Li, S., Chen, J., Dinsmore, B. A., Lane, C., Lauer, A. C., Fields, P. I., and Deng, X. (2019). SeqSero2: Rapid and Improved *Salmonella* Serotype Determination Using Whole-Genome Sequencing Data. *Appl. Environ. Microbiol.* 85(23). <https://doi.org/10.1128/AEM.01746-19>
- Zhao, H., Sun, R., Yu, P., and Alvarez, P. J. J. (2020). High levels of antibiotic resistance genes and opportunistic pathogenic bacteria indicators in urban wild bird feces. *Environ. Pollut.* 266, 115200. <https://doi.org/10.1016/j.envpol.2020.115200>
- Zhou, Z., Alikhan, N.-F., Mohamed, K., Fan, Y., and Achtman, M. (2020). The Enterobase user's guide, with case studies on *Salmonella* transmissions, *Yersinia pestis* phylogeny, and *Escherichia* core genomic diversity. *Genome Res.* 30(1), 138–152. <https://doi.org/10.1101/gr.251678.11>

CHAPTER 3

A NEED FOR INCREASED ENVIRONMENTAL SURVEILLANCE TO CLOSE THE
ONE HEALTH LOOP¹

¹Jared C. Smith, Amy T. Siceloff, Sherwin M. Shirazi, Rebecca L. Bell, Nikki W. Shariat. 2025. To be submitted to a peer-reviewed journal. JCS, ATS, and NWS conceived and planned the study.

3.1 Abstract

Salmonella enterica is a leading cause of bacterial foodborne illness, often transmitted through contaminated food and water. Considerable reductions of *Salmonella* contamination in meat and poultry products do not wholly contribute to decreased foodborne illness, highlighting the need to define alternative reservoirs and transmission pathways. In this study, we collected samples from four distinct creeks over 24 months to characterize *Salmonella* serovar diversity and utilized phylogenetic approaches, along with proximal land use analyses, to identify relationships between environmental reservoirs and hosts. Across 19 sites, including animal agriculture, suburban, and forested areas, 10L water samples were collected using modified Moore swabs (n = 456), and cultured for *Salmonella*, followed by whole genome sequencing of isolates and deep serotyping of multiserovar populations. Overall prevalence was 69% (314/456), and generalized linear mixed models showed that seasonal weather patterns, including precipitation and humidity, significantly influenced recovery and complexity in comparison to surrounding land use. Antimicrobial resistance was detected in 11% (33/314) of isolates, with 21% (7/33) classified as multidrug resistant. CRISPR-SeroSeq identified 37 serovars, and multiserovar populations were detected in 89% (229/258) of positive samples with sequencing data, averaging 3.7 serovars per sample (range: 1-13). Comparison with national food animal production monitoring showed limited serovar overlap, with serovar Rubislaw dominating water samples but absent in agricultural datasets. Collectively, these results demonstrate extensive serovar diversity within *Salmonella* populations in freshwater systems, including clinically relevant serovars, and

emphasize the need to develop a robust surveillance platform for source attribution and, ultimately, prevention of future outbreaks.

3.2 Significance statement

Contaminated surface water significantly contributes to global *Salmonella* illnesses, marking a critical need to assess serovars present and determine environmental variables affecting the population dynamics in this reservoir. We found that complex multiserovar populations, often including pathogenic serovars, occur in surface water regardless of proximal land use. Notably, many aquatic serovars are not detected in animal agriculture monitoring. However, limited serotyping data is available for alternative reservoirs of foodborne illness, namely wildlife, which hinders source attribution. Phylogenetic analysis revealed aquatic antimicrobial-resistant Infantis and Typhimurium isolates were more closely related to clinical than animal-source isolates. This study highlights a significant gap in understanding environmental *Salmonella* transmission and underscores the importance of a One Health surveillance approach to protect public health.

3.3 Introduction

The WHO reports that 1.4 million deaths worldwide could be prevented by improving water safety measures and reducing the load of contaminating bacteria (WHO, 2023). Notably, *Salmonella enterica* subsp. *enterica* is one of four key global causes of diarrheal diseases, and serovar Typhi alone contributes an estimated 9 million illnesses and 110,000 deaths through the consumption of water contaminated with fecal material (<https://www.who.int/news-room/fact-sheets/detail/typhoid>). While typhoidal *Salmonella* illnesses are uncommon in the United States (CDC, 2024), there are an estimated 1.35

million non-typhoidal illnesses each year, resulting in 26,500 hospitalizations, 420 deaths (CDC, *Salmonella*, 2024), and a total cost of \$4.1 billion (U.S. Department of Agriculture (USDA), Economic Research Service (ERS). Cost Estimates of Foodborne Illnesses). Non-typhoidal *Salmonella enterica* subsp. *enterica*, hereafter *Salmonella*, is also a contributing microorganism to the global threat of antibiotic resistance, as several resistant serovars have been affecting the food chain (WHO). However, 60-80% of salmonellosis cases are not connected to a known outbreak, which underscores a critical lack of knowledge needed to prevent future illnesses (WHO). Final products from the four main food animals in the United States (beef, chicken, pork, turkey) are routinely tested by the United States Department of Agriculture – Food Safety and Inspection Service (USDA-FSIS). In response to an ever-increasing number of related foodborne outbreaks, produce is also screened for pathogens in the U.S. Notably, an estimated 43.9% of *Salmonella* illnesses in the United States can be attributed to the consumption of contaminated produce, while 42.7% can be attributed to meat and poultry (IFSAC, 2024). These surveillance programs are necessary to protect consumers but also widen the knowledge gap of *Salmonella* transmission as wildlife and other environmental contributors to contamination in food production are not monitored. For example, contaminated water is believed to be a significant contributor to disseminating foodborne pathogens in produce (Bell et al., 2021). Most recently in cucumbers, serovars Braenderup and Africana infected 551 individuals, where both serovars were found in the agricultural water at the farms where the outbreaks originated (FDA, 2024). It is likely that contaminated irrigation water contributes to many produce-related outbreaks but these traceback investigations do not often result in exact sources due to the consumption

or spoilage of products, along with the clearing of fields following harvesting.

Additionally, produce contamination may occur both in the field and in packinghouses, further increasing the difficulty of attributing sources of *Salmonella*. Because meat, poultry, and egg surveillance, along with clinical isolates, comprise the majority of publicly available data, any large-scale genomic attribution studies are skewed towards animal agriculture. Subsequently, the risk is not properly assessed for serovars arising from alternative sources, such as wildlife or contaminated water.

There are over 2,600 *Salmonella* serovars, characterized by their combination of somatic (O) and flagellar (H) antigens (Grimont and Weill, 2007; Issenhuth-Jeanjean et al., 2014). Significant phenotypic and genotypic plasticity can be found across serovars, including host restriction (Uzzau et al., 2000), antimicrobial resistance (Shah et al., 2016), virulence factors (Cheng et al., 2019), and stress response (Gorski and Noriega, 2023). Most serovars are host generalists or host-adapted, which can lead to patterns of serovar isolation from the same sources repeatedly. While some wildlife reservoirs are often found with host-adapted serovars or strains, such as serovar Cholerasuis in wild boars and some subtypes of serovar Typhimurium circulating in wild birds (Uelze et al. 2021; Patel et al. 2023; Hernandez et al. 2016; Chiu et al., 2002; Cohen et al., 2021), serovar association from sources such as deer (Renter et al., 2006; Salas-Rosas et al. 2020; Topalcengiz et al. 2020), reptiles (Doden et al. 2021), raccoons (Very et al. 2016; Maurer et al. 2015), and other wildlife (Gorski et al. 2011; Millan et al. 2004; Maurer et al., 2015) are largely understudied. Additionally, wild animals are recognized as carriers of antimicrobial resistance (AMR), so identifying points of introduction and transmission

within the environment is critical to understanding the public health risk posed by wildlife activity. (Carroll et al., 2015; Ramey et al., 2020; Laborda et al., 2022).

In addition to its presence in a variety of animal hosts and surface water, *Salmonella* is often found in dust and soil as well (Whyte et al., 2003; Pal et al., 2021; Jacobsen and Bech, 2012; Jechalk et al., 2019; Bell et al., 2015; Deaven et al., 2021; Gorski et al., 2022; Haley et al., 2009). As an enteric pathogen, *Salmonella* likely enters these abiotic environments via fecal contamination from humans, wildlife, and animal agriculture (Liu et al., 2018; Santo Domingo et al. 2000; Abulreesh, 2012) and may be maintained following continual re-introduction or via persistence (Byappanahall et al., 2009; Sha et al., 2013; Gaertner et al., 2011; Brandl et al., 2005; Bell et al., 2015; Li et al., 2015). As *Salmonella* often exists as mixed serovar populations (Thompson et al., 2018; Obe et al., 2023; Siceloff et al., 2022), surface water may serve as a conduit for highly complex populations following contamination from multiple hosts (Deaven et al., 2021; Gorski et al., 2011; Maurer et al., 2015b; Micallef et al., 2012). *Salmonella* has been shown to survive at least 300 days in a freshwater source (Topalcengiz et al., 2019), indicating that, while growth may be limited, survival for extended periods is possible. Previous research (Haley et al., 2009; McEgan et al. 2013; Murphy et al., 2022; Weller et al., 2020; Truitt et al., 2018; Deaven et al., 2021) has indicated that meteorological variables, including temperature and precipitation, can influence *Salmonella* recovery from surface water.

Conventional *Salmonella* isolation by culture-based methods can be quite sensitive; however, previous work has demonstrated that the use of selective enrichment media, is necessary for *Salmonella* recovery but leads to skewed population proportions

(Gorski et al., 2024). While media bias can result in discordance between true and observed proportions, traditional culturing has the advantage of recovering isolates that can then be subtyped. Molecular approaches such as rapid detection through qPCR can identify positive samples quickly; however, no strain analysis for outbreak traceback is possible using this method. As an alternative approach, deep serotyping using CRISPR-SeroSeq (serotyping based on the sequences of native *Salmonella* CRISPRs) is an amplicon-based next-generation sequencing approach that profiles the relative frequency of multiple *Salmonella* serovars in a single sample (Thompson et al. 2018). Applying this method to *Salmonella*-positive surface water samples has shown that rivers harbor complex *Salmonella* populations (Deaven et al., 2021). At the time, this was attributed to the large size of the watershed (Susquehanna R. watershed, Pennsylvania), and that the land use in that region was diverse. The current study was designed to apply this methodology to gain an improved representation of the *Salmonella* population within a watershed, such that smaller creeks were sampled, and to compare serovar differences between land that was selected for one primary use (i.e., human activity, animal agriculture, or national forest).

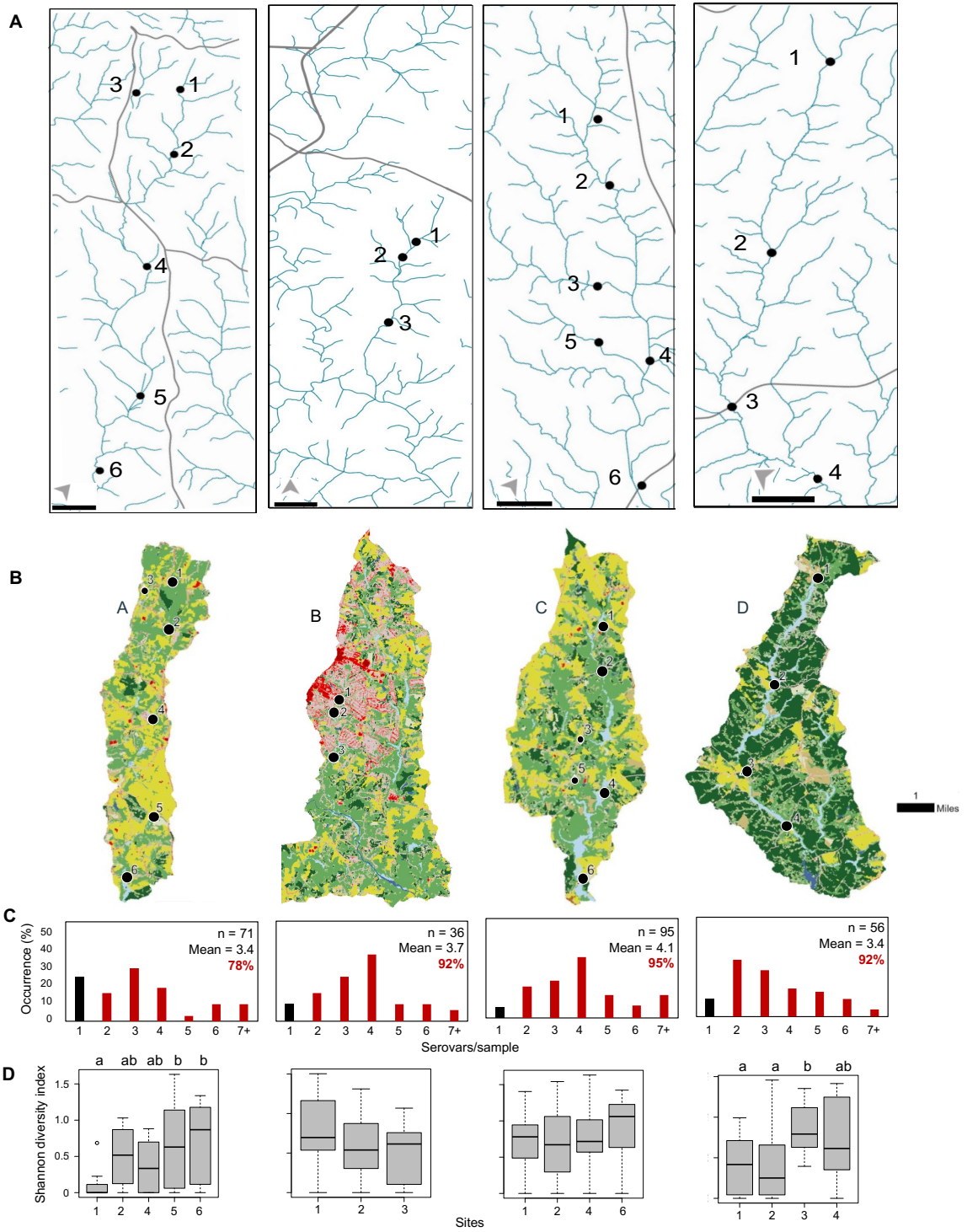
In this study, we sought to reveal the prevalence, population complexity, and antimicrobial resistance of non-typhoidal *Salmonella* within diverse freshwater aquatic environments representing four distinct creeks in the southeastern United States. Additionally, we investigated the influence of proximal land-attribution and weather on *Salmonella* in these systems. We utilized a total of 19 collection sites across the four creeks, where 10L water samples were collected from each site monthly over two years (November 2021 – October 2023; n = 456). From each sample, *Salmonella* was isolated

by traditional culturing techniques followed by a deep sequencing method of serotyping to characterize serovar populations. Additionally, antimicrobial resistance testing was performed. Generalized linear mixed models (GLMMs) were developed based on *Salmonella* prevalence and population complexity, weather patterns, and land-attribution data to determine significant contributors to *Salmonella* incidence and complexity as a risk assessment tool.

3.4 Results

Over 24 months (November 2021-October 2023), water samples were collected at 19 sites across four creeks using modified Moore swabs (Figure S3.1A). Each creek was selected for a different prominent land use, including animal agriculture (Creek A and C), a suburban community (B), and a national forest (D) (Figure 3.1A). In total, 69% (314/456) of samples were *Salmonella*-positive (Figure S3.1B). Prevalence differed across the creeks ($p < 0.05$, Chi-squared Test) and was highest in Creek C (78%; 112/144), and lowest in Creek A (63%; 91/114); both systems were adjacent to animal agriculture. Prevalence also differed by season ($p < 0.001$, Chi-squared Test), being highest in Spring (93%; 106/114) and lowest in Summer (47%; 54/114). Differences in *Salmonella* recovery were observed across culturing methods and fluctuated based on the season as well (Figure S3.1B).

A generalized linear mixed model (GLMM) was used to collectively analyze weather and land attribution data as well as the non-linear and interaction effects between and among variables to determine the likelihood of *Salmonella* detection and population complexity (Tables S3.1, S3.2). The models used fall as the baseline season (intercept), with the other seasons reported in relation to fall, and resulted in significance values and



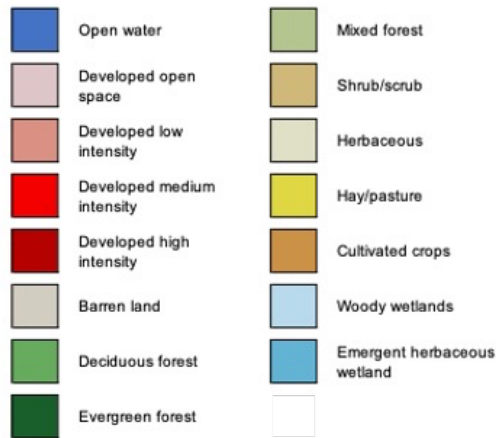


Figure 3.1. Sampling sites, multiserovar prevalence, and diversity of creeks. A) Each creek is shown with its corresponding land attribution from the national land cover database. Individual sites are shown as black circles with numbers representing each site. “Branching” sites off the main creek stem are shown as smaller black circles. B) Results of deep serotyping reveal the occurrence of multiserovar populations (red bar) compared to single-serovar populations (black bar). The number of positive samples, average serovars/sample, and percent of multiserovar populations for each creek are included. C) Shannon diversity indices of sites along the main stem of the creek, where lowercase letters show significant differences between sites using a Kruskal-Wallis and Dunn post-hoc test.

odds ratios (OR) to describe interactions between variables tested and *Salmonella* recovery. The likelihood of *Salmonella* detection was similar in winter and fall ($p > 0.05$), then higher in spring ($p < 0.05$, OR = 2.09, CI 1.26-3.48) and summer ($p < 0.05$, OR = 0.06, CI 0.04-0.01). Wind speed ($p < 0.05$, OR = 1.21, CI 1.1-1.32) and precipitation ($p < 0.05$, OR = 1.21, CI 1.13-1.29) corresponded with increased likelihood of detection, along with the presence of hay pasture ($p < 0.05$, OR = 1.03, CI 1.02-1.05) and mixed forest ($p < 0.05$, OR = 1.12, CI 1.08-1.17) as surrounding land. Similar results were obtained when modeling for serovar complexity, except summer ($p > 0.05$, OR = 0.60, CI 0.3-1.21) was not a significant predictor while winter was ($p < 0.05$, OR = 0.48, CI 0.26-0.87). Additionally, humidity ($p < 0.05$, OR = 1.06, CI 1.03-1.08) was found to increase the likelihood of complex *Salmonella* populations. Spring, as defined by sampling month, was the greatest predictor of *Salmonella* prevalence and complexity. The relationship between weather variables was explored as well using the Pearson correlation coefficient and revealed a strong association between maximum temperature and minimum temperature ($r = 0.89$), along with radiation ($r = 0.75$) (Figure S3.2).

From the antimicrobial susceptibility testing, 11% (33/314) of *Salmonella* isolates displayed AMR phenotypes, and 21% (7/33) of these were classified as multidrug resistant (MDR) (Figure S3.3). The most common resistance was to streptomycin (29/33, 88%), and one serovar Saintpaul isolate was resistant to seven antibiotics. These AMR isolates represented 15 serovars, as determined by whole genome sequencing, with serovar Anatum as the most predominant (6/33 isolates), comprising the majority of the MDR isolates (5/7). Out of the 15 serovars, seven are of human clinical importance as

denoted by their presence in the top 15 serovars listed on the BEAM Dashboard (CDC, 2024), including serovar Saintpaul which was MDR.

CRISPR-SeroSeq was applied to all *Salmonella*-positive samples. A subset of these failed to amplify ($n = 56/314$), though the majority of these samples ($45/56$, 80%) were only positive in non-selective enrichment media (BPW). Multiserovar populations were identified in 89% ($229/258$) of samples, with an average of 3.7 serovars per sample (range 1-13 serovars) and a total of 37 serovars detected (Figure S3.4). On average, Creek C had the highest number of serovars per sample (4.1) and the most samples containing multiple serovars ($90/95$, 95%), while Creek A had the lowest (3.4; $55/71$, 77%) (Figure 3.1B). Concordant with *Salmonella* prevalence, the greatest number of multiserovar samples were detected in spring ($81/86$, 94%), with significant differences in complexity when comparing based on creek (C to A), and season (Spring to Fall and Winter) ($p < 0.05$, Kruskal-Wallis test with Dunn post-hoc). Additionally, the Shannon diversity indices of *Salmonella* populations across sites following the main stem of each creek revealed that serovar complexity increased in downstream collection sites within creeks A and D (Figure 3.1C).

When comparing all serovars isolated from at least five samples across the whole dataset, serovars Give I, Muenchen I, Rubislaw, and Typhimurium were shared amongst all creeks (Figure S3.5A). Alternatively, some serovars were unique to a creek, such as serovars Agbeni and Oranienburg in Creek B and Anatum in Creek C. All serovars identified in more than five samples from Creek D were also found in other creeks. Creek A contained the greatest number of serovars ($n = 30$), while Creek D contained the least ($n = 18$). Additionally, differences in populations between sampling events were

analyzed by comparing the populations with a principal coordinates analysis based on Jaccard distance in each sample. This analysis showed that *Salmonella* populations were most consistent in Creek A, identified by measuring the average distance to the centroid point within each creek (Figure S3.5B). The most common serovar in creeks A, B, C, and D were Give I (n = 32), Rubislaw (n = 23), Montevideo II (n = 72), and Rubislaw (n = 46), respectively. The relative abundances of these serovars differ over time and within creeks, as one serovar may be highly abundant in one site and then lower in the next site (Figure 3.2). For example, serovar Infantis was almost exclusively isolated from sites 4 – 6 in Creek A. In Creek C, serovar Agbeni was only found at site 2 in the first six months of *Salmonella*-positive samples. Of the serovars identified 10 or more times across the creeks, Infantis, when present, comprised the largest average relative frequency (45.1%) within a sample. This was followed by Montevideo II and Braenderup at 42.1% and 40.7%, respectively. Serovar Rubislaw, although present in the most samples, had an average proportion of 32.6%.

USDA-FSIS conducts regular surveillance of animal agriculture systems, including poultry (turkey and chicken), pork, and beef. Utilizing current and historical data going back to 2016, with the exception of pork which only included datasets from 2019 to present day, the top 10 serovars by prevalence were identified for each commodity group and compared to water samples (Figure 3.3). The reported serovars for the former dataset do not include polyphyletic designations. The most commonly identified serovar for turkey, chicken, pork, and cattle was Reading (24%), Kentucky (33%), Anatum (18%), and Montevideo (23%), respectively. In contrast, serovar Rubislaw was most often found within the creek samples (22%), which contained

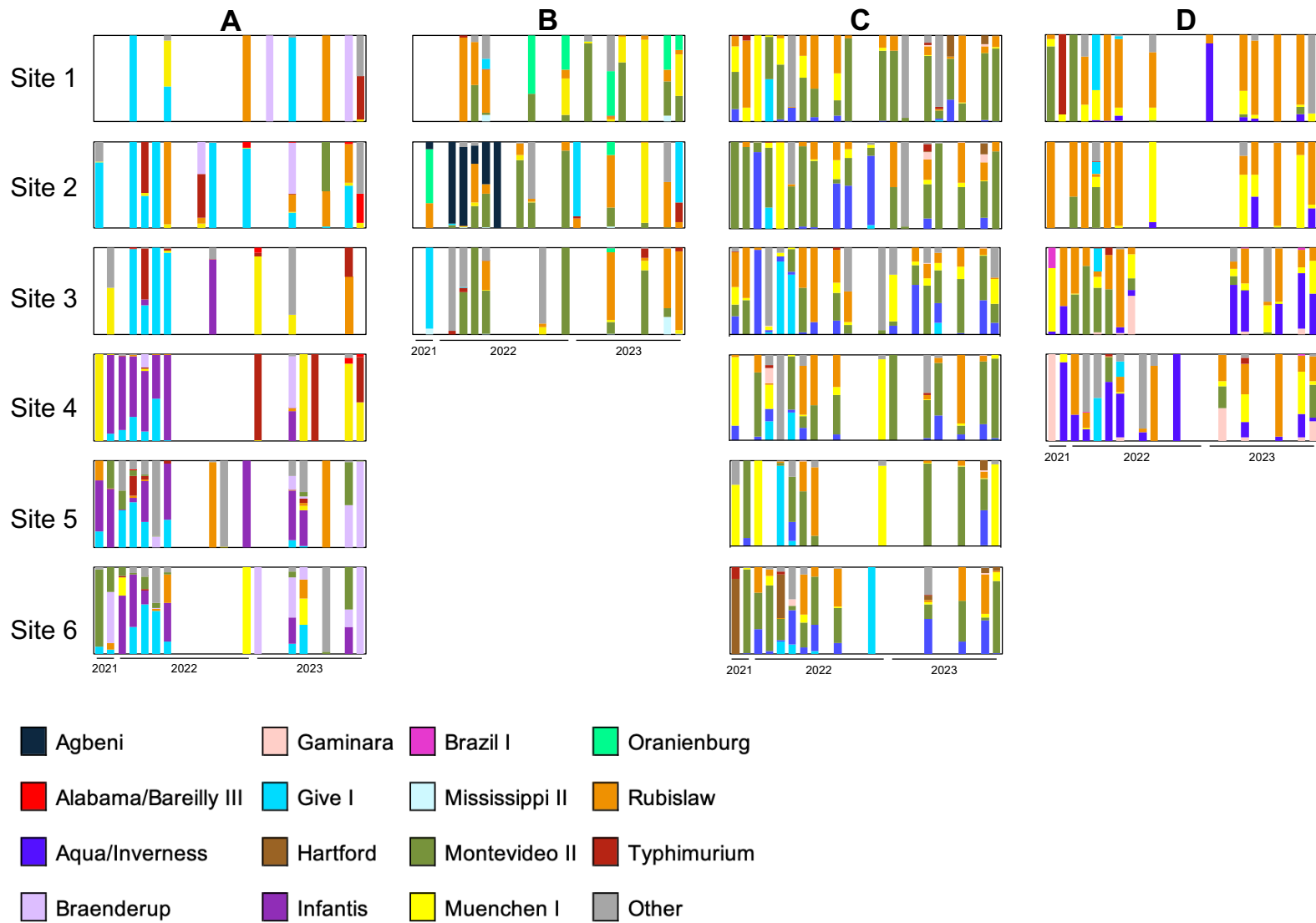


Figure 3.2. Top eight serovars present in each creek. Deep serotyping results are organized by site and creek. Each box represents the 24 months across a site, where colored bars show the proportion of each serovar present in that sample. Months with no bar represent a negative sample or a sample where deep serotyping was unsuccessful. Serovars outside of the top eight are classified as “other” and represented by a gray bar. The suffixes (-I, -II, -III) for some serovars refer to polyphyletic lineages.

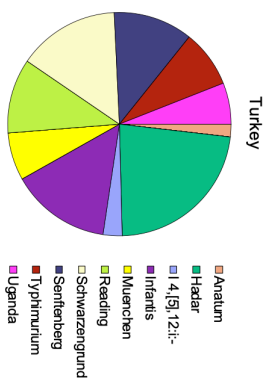
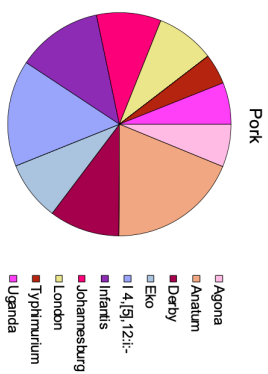
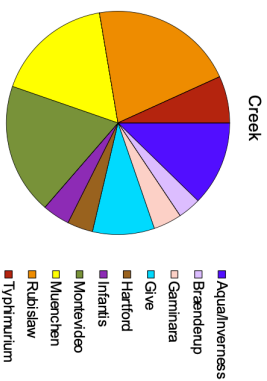
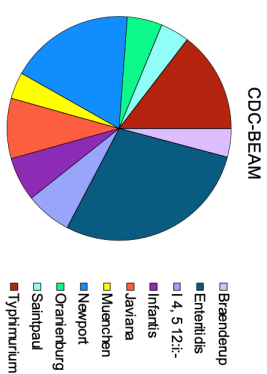
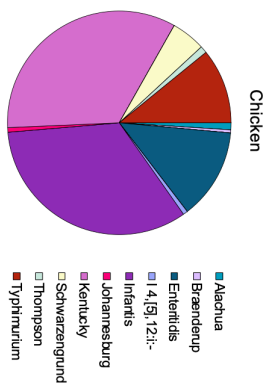
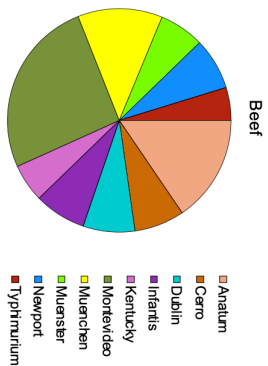
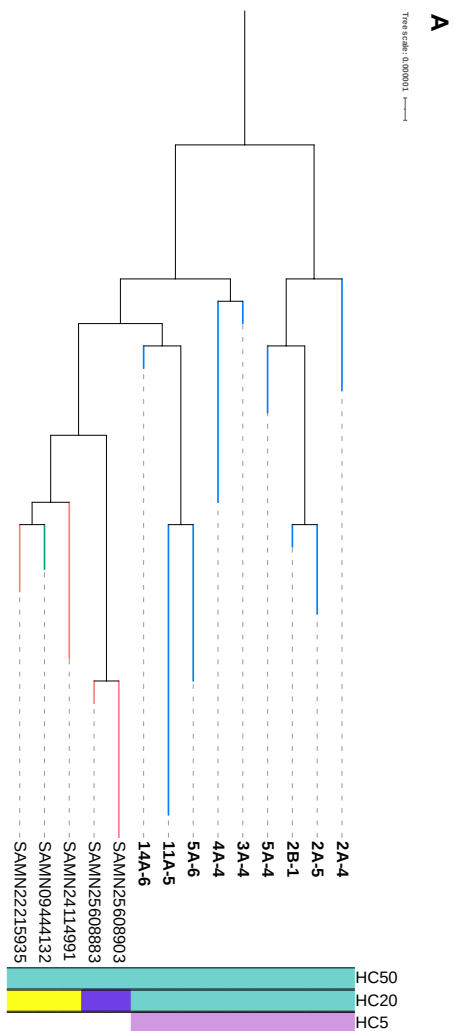


Figure 3.3. *Salmonella* serovars differ between animal agriculture, surface water, and clinical cases. The top 10 most common serovars for each food animal monitored by USDA-FSIS, clinical cases reported to CDC, and creek samples from this study are displayed. Serovar data for the commodity groups was accessed from the publicly available USDA-FSIS sampling repository (USDA-FSIS, 2024) and included samples that were collected between November 2021 – October 2023. Clinical isolates were accessed through the CDC BEAM Dashboard (CDC, 2024) and included cases for the entirety of 2021 - 2023. Polyphyletic serovars are not delineated by USDA-FSIS or CDC, so the suffixes were removed from serovars identified by deep serotyping and both results from both Montevideo lineages were combined for the creek pie chart. Serovar Typhimurium and its monophasic variant, I 4, [5], 12:i:-, cannot be distinguished by the CRISPR arrays.

additional serovars which were not present in the animal agriculture data, including Aqua/Inverness, Gaminara, Give I, Hartford, and Mississippi II.

Serovars Infantis and Typhimurium were the only serovars collected from this study that were also routinely identified in all four primary domestic food animals. Therefore, we completed a phylogenetic comparison of the study isolates belonging to these two serovars against publicly available genomes on NCBI-Pathogen Detection, which hosts human clinical and USDA-FSIS whole genome sequences, among others (Figures 3.4-3.5, S3.6-3.7, Tables S3.4-3.9). This revealed that the study isolates were most closely related to each other (they all fell within the same HC5 group) than to other isolates from human clinical, environmental, or food animal sources. For serovar Infantis, four of the five most closely related genomes were from human clinical isolates and the other was from a market swine in the midwest (Table S3.4). Importantly, the phylogenetic analyses (Figure 3.4A) showed that these five most closely related genomes did not share an HC20 type with the creek genomes, indicating that the creek isolates are genetically quite distinct from those on NCBI, which limits source attribution in this instance. Additional analysis using the CFSAN SNP Pipeline revealed that these creek isolates are between 53 – 104 SNPs apart from the NCBI isolates (Figure 3.4B). Similarly, for serovar Typhimurium, 91/95 of the most closely related isolates were clinical, and the remaining four were environmental, collected from almonds, white-tailed deer and emu feces, and a dairy cow from the northeast (Figure 3.5A, Table S3.7). There was more variation in the SNP differences between the serovar Typhimurium study isolates and NCBI isolates (Figure 3.5B), but only a select number of the human clinical



B

2A-4	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20	21	22	23	24	25	26	27	28	29	30	31	32	33
2A-5																																	
2B-1																																	
5A-4																																	
3A-4																																	
4A-4																																	
5A-6																																	
11A-5																																	
14A-6																																	
SAMN25608903																																	
SAMN25608883																																	
SAMN24114991																																	
SAMN09444132																																	
SAMN22215935																																	

Figure 3.4. Serovar Infantis isolates collected in this study are not closely related to other isolates from mixed sources. Study isolates are bolded, with the label format of month-creek-site, and NCBI isolates are listed by sample ID with branch color corresponding to isolation type (human clinical: red, environmental/other: green, creek: blue). Isolation type was provided by NCBI metadata, and the HierCC schemes were determined through Enterobase (Zhao, 2021). Schemes were only reported if they are shared between two or more isolates. The phylogenies are rooted at the midpoint, and include all isolates within the two SNP clusters most closely related to study isolates. HC20 and HC5 values were added to reflect that the study isolates were greater than 20 and 5 allelic differences from the NCBI isolates, respectively. A) Core genome phylogeny of study isolates and NCBI isolates. B) SNP matrix comparing the study isolates and NCBI isolates. Darker shading indicates a greater SNP distance.

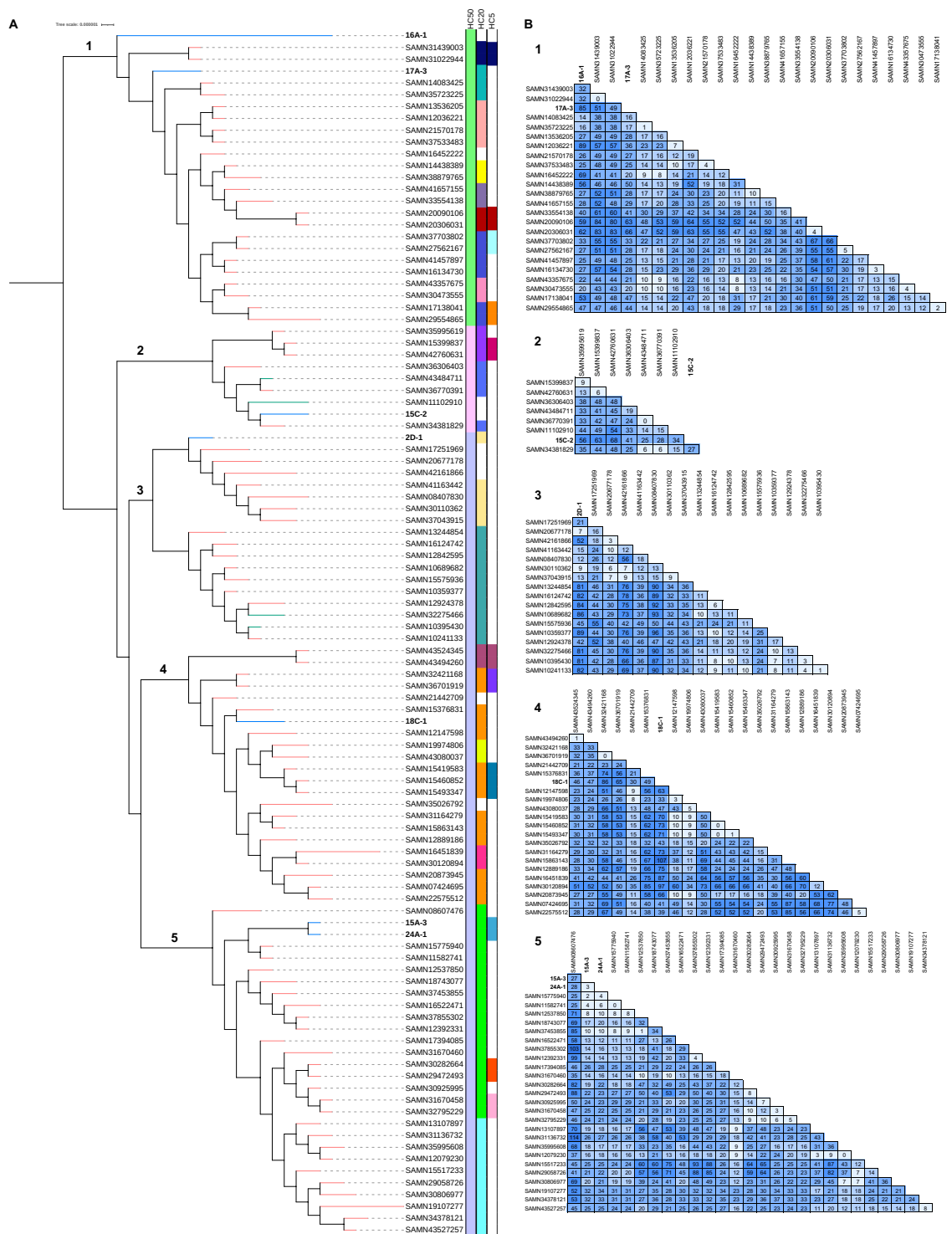


Figure 3.5. Serovar Typhimurium isolates collected in this study are more closely related to human clinical isolates than animal sources. Study isolates are bolded, with the label format of month-creek-site, and NCBI isolates are listed by sample ID with branch color corresponding to isolation type (human clinical: red, environmental/other: green, creek: blue). Isolation type was provided by NCBI metadata, and the HierCC schemes were determined through Enterobase (Zhao, 2021). Schemes were only reported if they are shared between two or more isolates. The phylogenies are rooted at the midpoint, and include all isolates within the two SNP clusters most closely related to study isolates. HC20 and HC5 values were added to reflect that the study isolates were greater than 20 and 5 allelic differences from the NCBI isolates, respectively. A) Core genome phylogeny of study isolates and NCBI isolates. B) SNP matrices comparing the study isolates and NCBI isolates based on corresponding phylogenetic clades (1-5). Darker shading indicates a greater SNP distance.

isolates ($n = 6$) were within 10 SNPs of the study isolates, and these were shared between two SNP clusters, as defined by NCBI. Further, hierarchical clustering based on cgMLST values determined that there were more than 20 and 50 allelic differences between serovar Infantis and Typhimurium study isolates and the most closely related isolates, respectively (Figs. 3.4A, 3.5A).

3.5 Discussion

Each day, 198 billion gallons of surface water are used in the US, where 60.9 billion gallons of that are used for irrigation alone, accounting for the largest proportion of this source (Dieter et al., 2018). For this reason, it is pertinent to investigate *Salmonella* in surface water. While significant progress has been made in global surveillance studies (McConn et al., 2024), much is still unknown about what factors influence survival and persistence in this environment.

Our data show that the *Salmonella* strains we recovered do not match to food animal isolates, suggesting alternative reservoirs. In particular, more research is required to understand the contribution of wildlife to *Salmonella* populations within freshwater, since these hosts may be introducing pathogenic serovars and contributing to the AMR identified in the environment (Hernandez et al., 2021; Maurer et al., 2015b). This is further supported by the identification of AMR and MDR isolates in this study. Applying a One Health approach will be necessary to mitigate the impacts of *Salmonella* found in the environment, since there is opportunity for cyclical transmission between animals and humans via shared environmental reservoirs, such as surface water. To this end, a large, multiyear study was recently completed as part of a collaboration between federal, academic, industry, and local groups in southwest Arizona to map introduction,

transmission, and persistence of foodborne pathogens

(<https://www.fda.gov/food/environmental-studies/southwest-agricultural-region-environmental-microbiology-study-2019-2024>). This study collected a wide variety of samples from irrigation waters, soil, sediment, air/dust, animal fecal material, wildlife scat, and other sources, and the resulting data will help inform management decisions and create a foundation for future surveillance studies.

As surface water is commonly used for agricultural water (i.e., irrigation), the Food Safety Modernization Act – Produce Safety Rule (FSMA-PSR) details its impact on food safety risks and provides guidance on best practices (FSMA Final Rule on Produce Safety, FDA 2024). In-field, antimicrobial treatments of irrigation water have limitations, since not all growers have access to the necessary resources and size of farm can greatly impact the efficacy of these treatments (Karp et al., 2015; Baur et al., 2017; Adalja and Lichtenberg, 2018; Devarajan et al. 2023). The extensive application of surface water and variability of sources also complicates risk assessment as *Salmonella* testing can be lengthy and costly, and there is likely not a one-size-fits-all control strategy. Many previous studies have sought to find an indicator organism that is easier to identify in the field and suggests the presence of *Salmonella* contamination. While *E. coli* has been considered an indicator of *Salmonella*, due to its association with fecal contamination and niche as enteric bacteria, some work has highlighted inconsistencies in the applicability of this organism (McEgan et al., 2013; Weller et al., 2020; Chung et al., 2023).

Salmonella prevalence throughout this study was variable, resulting in an overall prevalence of 69% (314/456), which is comparable to other studies in this region (Haley et al., 2009; McEgan et al., 2013; Cho et al., 2020) For example, a 3-month period during

2023 exhibited incidence of 100% (19/19) of sites to 0% (0/19), and back to 100% (19/19).

Salmonella prevalence was lower in the summer months. This could have been due to reduced rain and increased temperature that could have contributed to reduced nutrient availability and *Salmonella* influx to the environment. Previous work has shown a decrease in *Salmonella* prevalence in summer months (Deaven et al., 2021), while others have shown an increase during the same period (Liu et al., 2018; Haley et al., 2009), indicating high inter-creek variability. Similarly, precipitation has been associated with *Salmonella* in some studies (Strawn et al. 2012; Weller et al. 2020) while others found no relationship (Murphy et al. 2022; McEgan et al. 2013).

These findings were further supported by the GLMM output which suggested that *Salmonella* was more likely to be detected during spring, along with increased wind and precipitation. Increased proximity of hay pastures and mixed forests also increased the odds of detection, indicating that surrounding land attributes can contribute to *Salmonella* status. Alternatively, serovar complexity was associated with spring and winter, along with humidity and previously mentioned land use variables identified in the GLMM for prevalence. These findings can be used to support informed management decisions, as it could be more cost-effective to screen surface water for bacterial contamination after major precipitation events. However, more work is required to measure the contribution of different attribution sources to *Salmonella* surface water contamination, especially the wholly understudied role of wildlife.

Developing predictive models for likelihood of *Salmonella* detection in the environment, similar to work completed in this project, is a complex problem due to the

abundance of external interactions, along with regional differences in weather and creek composition, including nutrient availability and bacterial populations. Further, a model is only as good as the input data so cultural limitations of *Salmonella* identification and characterization, as well as variance in sample collection methods, may impact model development. One recent study demonstrated the importance of detection methods as both *Salmonella* prevalence and complexity differed in water samples that were processed with separate approaches (Murphy et al., 2024). Our study highlights the need for robust enrichment techniques to allow for the recovery of *Salmonella* while also trying to maintain original population dynamics, and high-resolution approaches to identify all serovars present within a sample.

When this study began, two different selective enrichment media, Rappaport-Vassiliadis (RV) and tetrathionate (TT) broths, were utilized and enrichments were plated onto xylose lysine tergitol-4 (XLT-4) agar. While the overall prevalence in the study was relatively high, there were several months when the prevalence was lower, particularly during the summer, and the total number of observed colonies present per plate during these times decreased as well. In the first 12 months, when *Salmonella* was not recovered following TT or RV enrichment, the pre-enrichment (buffered peptone water, BPW) culture was plated directly onto XLT-4 and, in some cases, yielded culturable isolates. It is possible that this is indicative of the stress upon cells in an aquatic environment, from which the non-selective enrichment for 24 hours prior was not able to allow for full recovery and thus, *Salmonella* were not able to proliferate in either RV or TT. It was noted that the inability to recover *Salmonella* in some of these samples was not due to overwhelming background as the plates were generally free of all growth. In the second

year of the study, a semisolid RV agar (MSRV) was included in addition to TT and RV as an alternate selective enrichment method, since the media components prevent growth of gram-positive bacteria. Importantly, MSRV plates serve as a motility test, such that *Salmonella* can use flagella to “swim” to the edge of the plate following culture inoculation in the middle and this provides a cleaner isolation (i.e., limited background bacteria) (Gorski et al., 2011). A previous study demonstrated that the parallel use of MSRV and a selective enrichment broth increased *Salmonella* recovery from water samples (Gorski, 2022), demonstrating the robustness of this approach and its applicability for *Salmonella* isolation from environmental samples.

Surface water harbors complex mixed-serovar *Salmonella* populations (Deaven et al., 2021), with an average of three serovars per sample identified in a large river (range: 1-10). Findings here show even higher complexity in small creeks (average of 3.7 serovars per sample, range 1-13). This is significant, demonstrating that these complex populations can form early in a creeks and are relevant to food safety. A previous study utilizing 16S analyses showed higher microbial community complexity upstream, with populations becoming more stable and less complex further downstream in creeks (Teachey et al., 2019; Liu et al., 2023). Focusing on a single species, the *Salmonella* data presented here is inverse; in two creeks, serovar populations increased in complexity as they flowed downstream, while the remaining two did not show complexity differences between up- and downstream sites (Figure 3.1C). A potential explanation for increasing complexity in some downstream sites is that the salmonellae are moving in the creek stem, and those populations are augmented by additional salmonellae that enter at later points, collectively accumulating downstream. Additionally, *Salmonella* can invade and

live in amoebas, ciliates, and biofilms which promote long-term survival in aquatic environments (Brandl et al., 2005; Byappanahalli et al., 2009; Gaertner et al., 2011; Sha et al., 2013).

The *Salmonella* populations identified here in surface water are more complex than those observed in food animals or in wild birds (Figure 3.3) (Siceloff et al. 2022; Obe et al., 2023; Siceloff et al., 2021; Cason et al., 2024; Smith et al. 2023). This is likely due to surface water acting as a catch-all for salmonellae from multiple different sources. Importantly, many serovars identified here are not commonly found in food animals or human isolates, suggesting that sources not traditionally surveyed may contribute to these populations. Additional research is needed to identify the factors that influence *Salmonella* complexity within all three components of a One Health scheme: animals, humans, and the environment. Meanwhile, surface water surveillance may serve as an indicator of salmonellae that are circulating within different ecosystems. Further, the expansion of environmental surveillance would help to develop a robust database that could begin to identify potential sources of sporadic illnesses not attributed to outbreaks. Subsequently, any generated whole genome sequences could provide new genomic insights on *Salmonella* transmission, including virulence factors needed for host colonization, environmental persistence, or stress response (Lipman et al., 2024). Previous work has been able to identify outbreak strains that had not previously been reported by using wastewater to test for serovar presence (M'ikanatha et al., 2024; Diemert et al., 2019; Vincent et al., 2007). Similarly to the concern that surface water acts as a reservoir for contamination by humans, agriculture, and wildlife, wastewater can potentially be used to measure burden of *Salmonella* illness or identify cases that would

otherwise go unreported (Berge et al., 2006; M'ikanatha et al., 2024). Since salmonellosis is often self-limiting within 7-10 days, it is significantly underrepresented in clinical cases, with an estimated 30 cases not reported for each lab-confirmed case, thus making wastewater surveillance a potential way to identify the presence and persistence of clinically relevant *Salmonella* serovars in human populations.

Human salmonellosis is often attributed, directly, or indirectly, to food animal production via consumption or improper storage of contaminated meat. While some serovars that are often found in food animals were also found in our study (e.g., Typhimurium, Infantis, Braenderup, Muenchen, Montevideo), the creeks were dominated by serovars that are not typical of these production systems. Serovars Give (Creek A) and Rubislaw (Creeks B-D) have been commonly associated with surface water environments, where reptiles are considered a reservoir for serovar Rubislaw and have been responsible for some human salmonellosis outbreaks (Waltenburg et al. 2022. Cho et al., 2020; Haley et al., 2009; Gorski et al., 2022; Pees et al., 2023; Lima Rocha et al., 2022; Rajabi et al., 2011). Notably, serovar Rubislaw contributes to more illnesses in Georgia than the national average, accounting for almost one-tenth of the illnesses caused by Enteritidis; the leading serovar in human cases (BEAM, 2025). Of these cases, nearly 13% come from blood or urine, suggesting this serovar is more likely to contribute to more serious illness compared to other serovars. Serovar Aqua/Inverness was frequently detected in Creeks C and D, and this also is not associated with food animals, based on data from the USDA-FSIS *Salmonella* verification program (USDA-FSIS, 2023). The current study did not include source attribution of fecal contamination within these creeks, but the Infantis isolates from this study are not related to the clonal lineage

responsible for the recent *Salmonella* outbreak in raw chicken products as determined by phylogenetic analyses and lack of pESI megaplasmid within the sequenced isolates (McMillan et al., 2020).

Serovar Agbeni was detected over six consecutive months at Site 2 in Creek B, and found once in Creek D. Interestingly, after these six months in Creek B, this serovar was never detected again for the remainder of the study, despite the depth provided by deep serotyping. While in a suburban area, this site is immediately downstream of a wooded area separating sites 1 and 2 with high incidence of wildlife (e.g., deer, raccoons, wild birds), and a sewage pipe and multiple manholes are within 500 ft of the sample collection. As it is unlikely that this serovar was shed from a human source for such a length of time, it is most likely linked to a wildlife reservoir(s) that repeatedly introduces it to the water (Langholz & Jay-Russell, 2013). However, previous work has yet to identify a food animal or wildlife source of serovar Agbeni. This illustrates the importance of further investigation into wildlife sources of clinically relevant *Salmonella* serovars. *Salmonella* surveillance data is dominated by that made available by federal agencies such as the U.S. Food and Drug Administration, the Centers for Disease Control and Prevention and USDA-FSIS. Under a One Health umbrella, understanding *Salmonella* in the environment and in wildlife is considerably understudied and there is a critical need to close this loop. While more work is needed to attribute serovars within these highly variable aquatic environments to likely hosts, the surveillance framework developed in this study can be helpful in identifying the salmonellae that are being circulated in respective environments.

The prevalence of AMR in this study was 11%. This is slightly greater than a recent study performed in Georgia (Cho et al., 2020), where isolates collected from water in the Middle Oconee River contained AMR phenotypes in 4.4% of samples. Our results indicate that antimicrobial resistance genes are being circulated even when human impact is limited but wildlife may be abundant, such as in the case of samples collected in a national forest (Creeek D). The presence of MDR in these isolates raises concern over the dissemination of antimicrobial resistance genes (ARGs) via various pathways, including direct and indirect wildlife interactions, horizontal gene transfer from bacterial communities, or human activity. As surveillance of wildlife has remained limited, an accurate assessment of the prevalence of both foodborne pathogens and ARGs is currently not within reach but this information can be supplemented with routine creek sampling. Otherwise, this limitation will continue to be a barrier to effective risk assessment as wildlife interaction in agricultural water and produce fields is a multifaceted problem with complex solutions (Gruszynski et al., 2014b).

While this work resulted in GLMMs to predict factors that influence the prevalence and complexity of *Salmonella* in creeks, the implications are limited to the effect of local climate and may not be applicable to other regions. The data used in these models can serve as a guide for research performed elsewhere to measure the variable effect of weather and land use on *Salmonella* in surface water and contribute to the growing body of work on *Salmonella* dynamics in the environment. Due to direct access restriction at some sites, water quality metrics, such as turbidity, dissolved oxygen, pH, conductivity, and salinity, were not collected and therefore not considered in modeling. However, previous work has demonstrated that the relationship between turbidity levels

and bacterial populations in water is dynamic (McEgan et al., 2013; Kim et al., 2023; Murphy et al., 2022), which further emphasizes that there is a complex network of biotic and abiotic factors that contributes to *Salmonella* abundance in water and there is not likely one main driver. Further, the sampling sites do not exist within a vacuum; for example, there may be human activity affecting *Salmonella* populations in the creek located within a national forest, and this is one of the limitations in modeling. *Salmonella* isolation from environmental samples can be difficult if there is an overwhelming amount of background bacteria competing for the same resources during growth and if the cells are unable to recover following any environmental stresses. As such, we had to modify our methods throughout the study as needed to increase *Salmonella* recovery (i.e., isolation directly from primary non-selective enrichments and inclusion of a primary mobility enrichment); this, in turn, limited next-generation sequencing results due to a decreased abundance of the target organism in the non-selectively enriched cultures. While we recorded the media for each *Salmonella* isolate (Figure S3.1B) and we continued to use traditional methods alongside alternative approaches, we acknowledge the variability added into our dataset through the use of multiple enrichment media and accounted for that by normalizing the relative serovar abundances across each media type for each sample. Our results demonstrate that most samples contained multiseroovar populations. This continues to demonstrate that selecting multiple colonies is necessary to reflect the *Salmonella* resistome within each sample (Cason J. et al., 2011).

In conclusion, this study highlights the complexity and variability of *Salmonella* populations within creeks influenced by diverse land uses and climatic factors. Our findings demonstrate that multiseroovar populations are prevalent within surface water and

suggest that environmental factors such as precipitation and humidity significantly influence *Salmonella* prevalence and complexity, although it is a multifaceted relationship. Population analyses showed increasing complexity further downstream, as well as higher serovar complexity in the spring compared to fall and winter. The identification of 37 serovars, including those of human concern with antimicrobial resistance, highlights the need for enhanced surveillance strategies that integrate environmental and wildlife monitoring. In particular, our serotyping data and subsequent whole genome sequence analysis revealed that there is a significant discordance between environmental, clinical, and animal agriculture isolates. These insights emphasize the critical importance of a One Health approach to address the interconnected risks posed by *Salmonella* in the environment, wildlife, and human systems. Future work should prioritize understanding the transmission pathways between these reservoirs and improving predictive modeling to inform targeted interventions and public health policies.

3.6 Materials and methods

Sample site selection

Four distinct creeks in the Southeastern United States were selected based on their surrounding land attribution, including animal agriculture, national forest, and suburban communities. The number of sites within each creek varied, based on size of the creek and accessibility of sites from the road. For each creek, at least one sample was collected from within 0.5 miles of the headwater. All sampling locations were within first or second order channels and were upstream of any wastewater or drinking water treatment facilities.

***Salmonella* population analysis**

To identify the populations of *Salmonella* within water samples, enrichments resulting in *Salmonella* culture positive samples were processed individually by centrifuging 1 mL of each positive selective enrichment at 18,000 rcf for three minutes. Total genomic DNA was isolated from the resulting pellet using a Promega Genome Wizard kit (Madison, WI) according to the manufacturer's instructions and resuspended in 200uL of molecular-grade water. A total of 2 µL of this template was used in the PCR for CRISPR-SeroSeq with primers targeting the conserved direct repeat sequences within *Salmonella* CRISPR arrays (Thompson et al., 2018; Siceloff et al., 2022). Primers also included index sequences, which facilitated multiplexed, high-throughput sequencing. PCR products were purified using the AMPure system (Beckman Coulter, Indianapolis, IN) and pooled in approximate equimolar ratios. Pooled libraries were sequenced using the Illumina NextSeq 550 platform (Illumina, California, USA) mid-output 150 cycle v2.5 kit with single-end reads. A water control and a positive control containing *Salmonella* serovar Enteritidis genomic DNA with a known CRISPR profile was included in the library. Sequence reads were parsed and matched in a local BLAST search to a lab-curated database of over 160 serovars (Siceloff et al., 2022). Serovars were called only if they contained multiple CRISPR spacers that were unique to that serovar. Where there were sufficient *Salmonella* sequence reads (>1,000 reads) for both the TT and RV enrichments, the relative frequency of each serovar was normalized across both enrichments to provide a single serovar profile. Within this dataset, serovars Alabama and Bareilly III, Aqua and Inverness, Johannesburg, and Urbana are not able to be distinguished based on CRISPR arrays, respectively.

Calculating land attribution for each sampling site

Land cover data was obtained from USGS national land cover database (NLCD) (<https://www.usgs.gov/centers/eros/science/national-land-cover-database>). Creek data was obtained from U.S. Geological Survey (USGS) national hydrography database (<https://www.usgs.gov/national-hydrography/access-national-hydrography-products>). Inverse-distance weights (IDW) as described in King et al. (2005) were used to characterize land cover within creeks. The IDW proportions for each land cover class were calculated for the following distance intervals (m): 0-100, 100-250, 250-500, 500-1,000, 1,000-2,000, and 2,000-5,000 upstream of the sampling sites. Land cover data for each creek was extracted using ArcGIS Pro version 3.1 (<https://www.mrlc.gov/data/legends/national-land-cover-database-class-legend-and-description>).

Statistical testing

All analyses were performed in R version 4.3.3. PcoA analysis was performed using the Vegan package in R. The deep serotyping results were normalized using the DESeq2 package (<http://www.bioconductor.org/packages/release/bioc/html/DESeq2.html>) to adjust the read counts per sample based on the size factors present.

Weather data collection

Meteorological data was collected using nearby weather stations. Measurements included maximum and minimum air temperature, relative humidity, soil temperature at two, four, and eight inches, average wind speed, total solar radiation and total

precipitation (Tables S10-12). These were measured between 12:15 am to 11:59pm each day. Creeks A and B had unique weather stations, while Creeks C and D shared the same weather station, as this was closer in proximity than any other station.

Predictive modeling for *Salmonella* prevalence and complexity

Preliminary exploratory data analysis (EDA) was performed to determine the modeling input. Analysis of variance inflation factors (VIFs) was used to assess the multicollinearity among the weather and land attribution variables, with a cutoff VIF value of 4. Additionally, all land attributes variables which comprised less than 5% of the total land cover per creek were removed. Generalized linear mixed models (GLMMs) were developed using lme4 package in R to investigate how the weather, land attribution, and temporal factors affected the likelihood of *Salmonella* detection and population complexity. In these models, all variables were considered fixed effects, while the creek was included as a random effect to account for pseudo-replication. The final GLMM was chosen following forward selection to determine if any quadratic or interaction effects should be added, along with backwards selection to remove any insignificant variables from the model.

Phylogenetic analyses

Due to the overlap of serovars Infantis and Typhimurium in both water and food animal sources, we compared the study isolates against publicly available genomes. For these phylogenetic analyses, we used iterative steps to reduce computational demands and to increase the resolution of relatedness in the final tree. Metadata, including accession numbers, for all isolates with computed (sero)type matching serovar Infantis or Typhimurium were downloaded from NCBI Pathogen Detection on October 10th, 2024

(Table S6-7). The following steps were completed in parallel for both serovars of interest. To reduce selection bias, the first isolate with an assembly was chosen as the representative for each single nucleotide polymorphism (SNP) cluster, as organized on Pathogen Detection. Isolates within a SNP cluster are considered highly related (NCBI, 2024). Some SNP clusters were excluded from further analyses due to isolates lacking a GenBank assembly accession, for a total of 650 and 3,057 representative genomes/SNP clusters of serovars Infantis and Typhimurium, respectively. For each serovar, a core genome alignment was generated with the NCBI and creek genomes using roary (Page et al., 2015). The resulting alignment, consisting of 3,124 and 3,511 genes for serovars Infantis and Typhimurium, respectively, was used to generate a phylogenetic tree with Very Fast Tree (Piñeiro et al., 2020). All trees were visualized with iTOL (Letunic & Bork, 2021). To increase the phylogenetic resolution, the clades containing the study isolates and all representative isolates within two most recent common ancestors were selected to generate a core genome alignment (yellow highlight, Table S3.5 and S3.8, Figure S3.6-7). As the decreased number of isolates allowed for more computationally intensive algorithms, RAXML-NG (Kozlov et al., 2019) (1,000 bootstrap replicates, GTR+G model) was used to infer a maximum-likelihood phylogeny for the subset clades. Based on the resulting tree, we determined which two SNP clusters contained the most closely related for each creek isolate (Infantis: n = 2 SNP clusters; Typhimurium: n = 14 SNP clusters) (blue highlight, Table S3.6-S3.9, Figure S3.6-7) and analyzed all annotated assemblies within those clusters (Infantis: n = 5 isolates; Typhimurium: n = 95 isolates) to create a final phylogeny of the most closely related isolates on NCBI to the study isolates, using roary, RAXML-NG, and iTol as described above. To determine the

hierarchical clustering scheme (HierCC) for all isolates included in the final phylogeny, the corresponding raw reads were transferred from NCBI Short Read Archive to Enterobase (Zhou et al., 2018, 2020a, 2021a). The HierCC 50, 20, and 5 categories were included to demonstrate the genetic distance between study isolates and NCBI isolates, based on the clusters calculated with the distance between core genome multilocus sequence types (cgMLST). Additionally, the CFSAN SNP Pipeline was used to generate a pairwise comparison matrix of each isolate included in the final phylogenies (Davis et al., 2015).

3.7 Data availability statement

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found in the article.

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3.10 Competing Interest Statement

NWS has a licensing agreement with Ancera Inc.

3.11 Author Contributions

JCS and ATS performed the experiments, completed the analysis and wrote the first draft of the paper. SMS contributed to statistical analysis. RLB contributed to the design of the study and WGS of isolates. All authors reviewed, edited, and approved the paper.

3.12 Reference

Burden of Disease Attributable to Unsafe Drinking-Water, Sanitation and Hygiene.

<https://www.who.int/publications/i/item/9789240075610>. Accessed

November 3rd, 2024.

Centers for Disease Control and Prevention. National Typhoid and Paratyphoid Fever

Surveillance Annual Report, 2020. Atlanta, Georgia: US Department of Health

and Human Services, CDC, 2023. (2020). <https://www.cdc.gov/typhoid->

[fever/php/surveillance/index.html](https://www.cdc.gov/typhoid-fever/php/surveillance/index.html). Accessed December 14th, 2024.

Centers for Disease Control and Prevention. *Salmonella*: Questions and Answers. (2023).

<https://www.cdc.gov/salmonella/general/index.html>. Accessed November

20th, 2024.

United States Department of Agriculture (USDA), E. R. S. (ERS). Cost of foodborne

Illnesses estimates for *Salmonella* (non-typhoidal). (2021).

<https://www.ers.usda.gov/data-products/cost-estimates-of-foodborne->

[illnesses](https://www.ers.usda.gov/data-products/cost-estimates-of-foodborne-illnesses). Accessed December 13th, 2024.

World Health Organization (WHO). *Salmonella* (non-typhoidal). (2018).

[https://www.who.int/news-room/fact-sheets/detail/salmonella-\(non-](https://www.who.int/news-room/fact-sheets/detail/salmonella-(non-typhoidal))

[typhoidal\)](https://www.who.int/news-room/fact-sheets/detail/salmonella-(non-typhoidal)). Accessed November 18th, 2024.

Interagency Food Safety Analytics Collaboration. Foodborne illness source attribution

estimates for 2022 for *Salmonella*, *Escherichia coli* O157, and *Listeria*

monocytogenes using multi-year outbreak surveillance data, United States. GA

and D.C.: U.S. Department of Health and Human Services, Centers for Disease

Control and Prevention, Food and Drug Administration, U.S. Department of

Agriculture's Food Safety and Inspection Service (2024).

<https://www.cdc.gov/ifsac/php/data-research/annual-report-2022.html>.

Accessed January 7th, 2025.

Grimont, P. A. D. & Weill, F.-X. WHO Collaborating Centre for Reference and Research on *Salmonella* Antigenic Formulae of The *Salmonella* Serovars 2007 9th Edition (2007).

Issenhuth-Jeanjean, S. *et al.* Supplement 2008–2010 (no. 48) to the White–Kauffmann–Le Minor scheme. *Res Microbiol* **165**, 526–530 (2014).

Uzzau, S. *et al.* Host adapted serotypes of *Salmonella enterica*. *Epidemiol Infect* **125**, 229–255 (2000).

Shah, D. H., Paul, N. C., Sischo, W. C., Crespo, R. & Guard, J. Population dynamics and antimicrobial resistance of the most prevalent poultry-associated *Salmonella* serotypes. *Poult Sci* **96**, 687–702 (2017).

Cheng, R. A., Eade, C. R. & Wiedmann, M. Embracing Diversity: Differences in Virulence Mechanisms, Disease Severity, and Host Adaptations Contribute to the Success of Nontyphoidal *Salmonella* as a Foodborne Pathogen. *Front Microbiol* **10**, 1368 (2019).

Gorski, L. & Noriega, A. A. Comparison of Phenotype Nutritional Profiles and Phosphate Metabolism Genes in Four Serovars of *Salmonella enterica* from Water Sources. *Microorganisms* **11**, 2109 (2023).

Uelze, L. *et al.* What WGS Reveals about *Salmonella enterica* subsp. *enterica* in Wildlife in Germany. *Microorganisms* **9**, 1911 (2021).

- Patel, K. *et al.* Human Salmonellosis Outbreak Linked to *Salmonella* Typhimurium Epidemic in Wild Songbirds, United States, 2020–2021. *Emerg Infect Dis* **29**, (2023).
- Hernandez, S. M. *et al.* Urbanized White Ibises (*Eudocimus albus*) as Carriers of *Salmonella enterica* of Significance to Public Health and Wildlife. *PLoS One* **11**, e0164402 (2016).
- Renter, D. G., Gnad, D. P., Sargeant, J. M. & Hygnstrom, S. E. Prevalence and Serovars of *Salmonella* in the Feces of Free-Ranging White-Tailed Deer (*Odocoileus virginianus*) in Nebraska. *J Wildl Dis* **42**, 699–703 (2006).
- Salas-Rosas, L. M. *et al.* Prevalence of *Salmonella* spp. and *Escherichia coli* O157 in a red deer herd (*Cervus elaphus scoticus*) in central Mexico. *Applied Animal Science* **36**, 622–629 (2020).
- Topalcengiz, Z. *et al.* Survival of *Salmonella* in Various Wild Animal Feces That May Contaminate Produce. *J Food Prot* **83**, 651–660 (2020).
- Doden, G. *et al.* Prevalence and antimicrobial resistance patterns of *Salmonella* spp. In two free-ranging populations of eastern box turtles (*terrapene carolina carolina*). *Journal of Zoo and Wildlife Medicine* **52**, (2021).
- Very, K. J. *et al.* Prevalence and Spatial Distribution of *Salmonella* Infections in the Pennsylvania Raccoon (*Procyon lotor*). *Zoonoses Public Health* **63**, 223–233 (2016).
- Maurer, J. J. *et al.* Diversity and Persistence of *Salmonella enterica* Strains in Rural Landscapes in the Southeastern United States. *PLoS One* **10**, e0128937 (2015).

- Gorski, L. *et al.* Prevalence, Distribution, and Diversity of *Salmonella enterica* in a Major Produce Region of California. *Appl Environ Microbiol* **77**, 2734–2748 (2011).
- Millan, J., Aduriz, G., Moreno, B., Juste, R. A. & Barral, M. *Salmonella* isolates from wild birds and mammals in the Basque Country (Spain). *Revue Scientifique et Technique de l'OIE* **23**, 905–911 (2004).
- Carroll, D., Wang, J., Fanning, S. & McMahon, B. J. Antimicrobial Resistance in Wildlife: Implications for Public Health. *Zoonoses Public Health* **62**, 534–542 (2015).
- Ramey, A. M. & Ahlstrom, C. A. Antibiotic resistant bacteria in wildlife: perspectives on trends, acquisition and dissemination, data gaps, and future directions. *J Wildl Dis* **56**, 1 (2020).
- Laborda, P. *et al.* Wildlife and Antibiotic Resistance. *Front Cell Infect Microbiol* **12**, (2022).
- Whyte, P., McGill, K. & Collins, J. D. A survey of the prevalence of *Salmonella* and other enteric pathogens in a commercial poultry feed mill. *J Food Saf* **23**, 13–24 (2003).
- Pal, A. *et al.* Impact of poultry litter *Salmonella* levels and moisture on transfer of *Salmonella* through associated in vitro generated dust. *Poult Sci* **100**, 101236 (2021).
- Jacobsen, C. S. & Bech, T. B. Soil survival of *Salmonella* and transfer to freshwater and fresh produce. *Food Research International* **45**, 557–566 (2012).
- Jechalke, S. *et al.* *Salmonella* Establishment in Agricultural Soil and Colonization of Crop Plants Depend on Soil Type and Plant Species. *Front Microbiol* **10**, (2019).

- Liu, H., Whitehouse, C. A. & Li, B. Presence and Persistence of *Salmonella* in Water: The Impact on Microbial Quality of Water and Food Safety. *Front Public Health* **6**, (2018).
- Santo Domingo, J. W., Harmon, S. & Bennett, J. Survival of *Salmonella* Species in River Water. *Curr Microbiol* **40**, 409–417 (2000).
- Abulreesh H. Salmonellae in the Environment. in *Salmonella - Distribution, Adaptation, Control Measures and Molecular Technologies* (InTech, 2012).
doi:10.5772/28201.
- Byappanahalli, M. N. *et al.* Seasonal stability of Cladophora-associated *Salmonella* in Lake Michigan watersheds. *Water Res* **43**, 806–814 (2009).
- Sha, Q., Vattem, D. A., Forstner, M. R. J. & Hahn, D. Quantifying *Salmonella* Population Dynamics in Water and Biofilms. *Microb Ecol* **65**, 60–67 (2013).
- Gaertner, J. P., Mendoza, J. A., Forstner, M. R. J. & Hahn, D. Recovery of *Salmonella* from biofilms in a headwater spring ecosystem. *J Water Health* **9**, 458–466 (2011).
- Brandl, M. T., Rosenthal, B. M., Haxo, A. F. & Berk, S. G. Enhanced Survival of *Salmonella enterica* in Vesicles Released by a Soilborne *Tetrahymena* Species. *Appl Environ Microbiol* **71**, 1562–1569 (2005).
- Cox, N. A. *et al.* Population Analyses Reveal Preenrichment Method and Selective Enrichment Media Affect *Salmonella* Serovars Detected on Broiler Carcasses. *J Food Prot* **82**, 1688–1696 (2019).
- Thompson, C. P. *et al.* High-Resolution Identification of Multiple *Salmonella* Serovars in a Single Sample by Using CRISPR-SeroSeq. *Appl Environ Microbiol* **84**, (2018).

- Obe, T., Siceloff, A. T., Crowe, M. G., Scott, H. M. & Shariat, N. W. Combined Quantification and Deep Serotyping for *Salmonella* Risk Profiling in Broiler Flocks. *Appl Environ Microbiol* **89**, (2023).
- Siceloff, A. T., Waltman, D. & Shariat, N. W. Regional *Salmonella* Differences in United States Broiler Production from 2016 to 2020 and the Contribution of Multiserovar Populations to *Salmonella* Surveillance. *Appl Environ Microbiol* **88**, (2022).
- Topalcengiz, Z., McEgan, R. & Danyluk, M. D. Fate of *Salmonella* in Central Florida Surface Waters and Evaluation of EPA Worst Case Water as a Standard Medium. *J Food Prot* **82**, 916–925 (2019).
- Haley, B. J., Cole, D. J. & Lipp, E. K. Distribution, Diversity, and Seasonality of Waterborne Salmonellae in a Rural Watershed. *Appl Environ Microbiol* **75**, 1248–1255 (2009).
- McEgan, R., Mootian, G., Goodridge, L. D., Schaffner, D. W. & Danyluk, M. D. Predicting *Salmonella* Populations from Biological, Chemical, and Physical Indicators in Florida Surface Waters. *Appl Environ Microbiol* **79**, 4094–4105 (2013).
- Murphy, C. M. *et al.* Factors Associated With *E. coli* Levels in and *Salmonella* Contamination of Agricultural Water Differed Between North and South Florida Waterways. *Frontiers in Water* **3**, (2022).
- Weller, D. *et al.* Complex Interactions Between Weather, and Microbial and Physicochemical Water Quality Impact the Likelihood of Detecting Foodborne Pathogens in Agricultural Water. *Front Microbiol* **11**, (2020).

Truitt, L. N. *et al.* Microbial Quality of Agricultural Water Used in Produce Preharvest Production on the Eastern Shore of Virginia. *J Food Prot* **81**, 1661–1672 (2018).

United States Food and Drug Administration. Outbreak Investigation of *Salmonella*: Cucumbers (June 2024). <https://www.fda.gov/food/outbreaks-foodborne-illness/outbreak-investigation-salmonella-cucumbers-june-2024>. Accessed January 2nd, 2024

Deaven, A. M. *et al.* *Salmonella* Genomics and Population Analyses Reveal High Inter- and Intraseroovar Diversity in Freshwater. *Appl Environ Microbiol* **87**, (2021).

Gorski, L. Selective Enrichment Media Bias the Types of *Salmonella enterica* Strains Isolated from Mixed Strain Cultures and Complex Enrichment Broths. *PLoS One* **7**, e34722 (2012).

Gorski, L. *et al.* Growth assessment of *Salmonella enterica* multi-serovar populations in poultry rinsates with commonly used enrichment and plating media. *Food Microbiol* **119**, 104431 (2024).

CDC. BEAM Dashboard. <https://www.cdc.gov/ncezid/dfwed/BEAM-dashboard.html> (2024). Accessed January 8th, 2025

McConn, B. R. *et al.* An analysis of culture-based methods used for the detection and isolation of *Salmonella* spp., *Escherichia coli*, and *Enterococcus* spp. from surface water: A systematic review. *Science of The Total Environment* **927**, 172190 (2024).

Hernandez, S. M., Maurer, J. J., Yabsley, M. J., Peters, V. E., Presotto, A., Murray, M. H., Curry, S., Sanchez, S., Gerner-Smidt, P., Hise, K., Huang, J., Johnson, K.,

- Kwan, T., & Lipp, E. K. Free-Living Aquatic Turtles as Sentinels of *Salmonella* spp. for Water Bodies. *Frontiers in veterinary science*, **8**, 674973 (2021).
- US Food and Drug Administration (FDA). FSMA Final Rule on Produce Safety. (2024).
<https://www.fda.gov/food/food-safety-modernization-act-fsma/fsma-final-rule-produce-safety>. Accessed January 8th, 2025
- Karp, D. S. *et al.* Comanaging fresh produce for nature conservation and food safety. *Proceedings of the National Academy of Sciences* **112**, 11126–11131 (2015).
- Baur, P., Getz, C. & Sowerwine, J. Contradictions, consequences and the human toll of food safety culture. *Agric Human Values* **34**, 713–728 (2017).
- Adalja, A. & Lichtenberg, E. Produce growers' cost of complying with the Food Safety Modernization Act. *Food Policy* **74**, 23–38 (2018).
- Devarajan, N. *et al.* Evidence for the efficacy of pre-harvest agricultural practices in mitigating food-safety risks to fresh produce in North America. *Front Sustain Food Syst* **7**, 1101435 (2023).
- Chung, T., Weller, D. L. & Kovac, J. The Composition of Microbial Communities in Six Streams, and Its Association With Environmental Conditions, and Foodborne Pathogen Isolation. *Front Microbiol* **11**, (2020).
- Cho, S., Jackson, C. R. & Frye, J. G. The prevalence and antimicrobial resistance phenotypes of *Salmonella*, *Escherichia coli* and *Enterococcus* sp. in surface water. *Lett Appl Microbiol* **71**, 3–25 (2020).
- Murphy, C. M., Weller, D. L. & Strawn, L. K. Scale and detection method impacted *Salmonella* prevalence and diversity in ponds. *Science of The Total Environment* **907**, 167812 (2024).

- Gorski, L. *et al.* *Salmonella enterica* Serovar Diversity, Distribution, and Prevalence in Public-Access Waters from a Central California Coastal Leafy Green-Growing Region from 2011 to 2016. *Appl Environ Microbiol* **88**, (2022).
- Teachey, M. E., McDonald, J. M. & Ottesen, E. A. Rapid and Stable Microbial Community Assembly in the Headwaters of a Third-Order Stream. *Appl Environ Microbiol* **85**, (2019).
- Liu, B. *et al.* The Variations of Microbial Diversity and Community Structure Along Different Stream Orders in Wuyi Mountains. *Microb Ecol* **86**, 2330–2343 (2023).
- Siceloff, A. T. *et al.* Antimicrobial Resistance Hidden within Multiserovar *Salmonella* Populations. *Antimicrob Agents Chemother* **65**, e00048-21 (2021).
- Cason, E. E., Carlson, A. V., Siemens, A. L. & Shariat, N. W. High-resolution Serotyping Reveals *Salmonella* Surveillance Challenges in the Turkey Industry. *J Food Prot* **87**, 100319 (2024).
- M’ikanatha, N. M. *et al.* Outbreak-associated *Salmonella* Baildon found in wastewater demonstrates how sewage monitoring can supplement traditional disease surveillance. *J Clin Microbiol* **62**, (2024).
- Diemert, S. & Yan, T. Clinically Unreported Salmonellosis Outbreak Detected via Comparative Genomic Analysis of Municipal Wastewater *Salmonella* Isolates. *Appl Environ Microbiol* **85**, (2019).
- Vincent, V., Scott, H. M., Harvey, R. B., Alali, W. Q. & Hume, M. E. Novel Surveillance of *Salmonella enterica* Serotype Heidelberg Epidemics in a Closed Community. *Foodborne Pathog Dis* **4**, 375–385 (2007).

- Berge, A. C. B., Dueger, E. L. & Sischo, W. M. Comparison of *Salmonella enterica* serovar distribution and antibiotic resistance patterns in wastewater at municipal water treatment plants in two California cities. *J Appl Microbiol* **101**, 1309–1316 (2006).
- Waltenburg, M.A., Perez, A., Salah, Z., Karp, B.E., Whichard, J., Tolar, B., Gollarza, L., Koski, L., Blackstock, A., Basler, C., Nichols, M. Multistate reptile- and amphibian-associated salmonellosis outbreaks in humans, United States, 2009–2018. *Zoonosis Public Hlth* **69**, 925–937 (2022).
- Pees, M., Brockmann, M., Steiner, N. & Marschang, R. E. *Salmonella* in reptiles: a review of occurrence, interactions, shedding and risk factors for human infections. *Front Cell Dev Biol* **11**, (2023).
- Rocha, A. D. de L. *et al.* Revisiting the Biological Behavior of *Salmonella enterica* in Hydric Resources: A Meta-Analysis Study Addressing the Critical Role of Environmental Water on Food Safety and Public Health. *Front Microbiol* **13**, (2022).
- Rajabi, M., Jones, M., Hubbard, M., Rodrick, G. & Wright, A. C. Distribution and Genetic Diversity of *Salmonella enterica* in the Upper Suwannee River. *Int J Microbiol* 1–9, (2011).
- Centers for Disease Control and Prevention (CDC). BEAM (Bacteria, Enterics, Ameba, and Mycotics) Dashboard. Atlanta, Georgia: U.S. Department of Health and Human Services. www.cdc.gov/ncezid/dfwed/BEAM-dashboard.html. Accessed January 23rd, 2025.

- McMillan, E. A. *et al.* Carriage and Gene Content Variability of the pESI-Like Plasmid Associated with *Salmonella* Infantis Recently Established in United States Poultry Production. *Genes (Basel)* **11**, 1516 (2020).
- Kim, S. *et al.* Temporal Stability of *Salmonella enterica* and *Listeria monocytogenes* in Surface Waters Used for Irrigation in the Mid-Atlantic United States. *J Food Prot* **86**, 100058 (2023).
- King, R. S. *et al.* Spatial considerations for linking watershed land cover to ecological indicators in streams. *Ecological Applications* **15**, 137–153 (2005).
- NCBI. Data Processing Pipeline.
https://www.ncbi.nlm.nih.gov/pathogens/docs/data_processing/. 2024.
 Accessed December 12, 2024.
- Page, A. J. *et al.* Roary: rapid large-scale prokaryote pan genome analysis. *Bioinformatics* **31**, 3691–3693 (2015).
- Piñeiro, C., Abuín, J. M. & Pichel, J. C. Very Fast Tree: speeding up the estimation of phylogenies for large alignments through parallelization and vectorization strategies. *Bioinformatics* **36**, 4658–4659 (2020).
- Letunic, I. & Bork, P. Interactive Tree Of Life (iTOL) v5: an online tool for phylogenetic tree display and annotation. *Nucleic Acids Res* **49**, 293–296 (2021).
- Kozlov, A. M., Darriba, D., Flouri, T., Morel, B. & Stamatakis, A. RAxML-NG: a fast, scalable and user-friendly tool for maximum likelihood phylogenetic inference. *Bioinformatics* **35**, 4453–4455 (2019).
- Zhou, Z. *et al.* GrapeTree: visualization of core genomic relationships among 100,000 bacterial pathogens. *Genome Res* **28**, 1395–1404 (2018).

- Zhou, Z., Alikhan, N.-F., Mohamed, K., Fan, Y. & Achtman, M. The EnteroBase user's guide, with case studies on *Salmonella* transmissions, *Yersinia pestis* phylogeny, and *Escherichia* core genomic diversity. *Genome Res* **30**, 138–152 (2020).
- Zhou, Z., Charlesworth, J. & Achtman, M. HierCC: a multi-level clustering scheme for population assignments based on core genome MLST. *Bioinformatics* **37**, 3645–3646 (2021).
- Davis, S. *et al.* CFSAN SNP Pipeline: an automated method for constructing SNP matrices from next-generation sequence data. *PeerJ Comput Sci* **1**, e20 (2015).
- Clinical and Laboratory Standards Institute (CLSI). *Performance Standards for Antimicrobial Susceptibility Testing*. (2024).

CHAPTER 4

SALMONELLA TRANSMISSION AND PERSISTENCE IN TWO RIVERS IN
PICHINCHA, ECUADOR¹

¹Jared C. Smith, David Ayala-Velastegui, Christian Vinizio Vinueza Burgos,
Nikki W. Shariat. 2025. To be submitted to a peer-reviewed journal.

4.1 Abstract

Salmonella enterica is a leading global cause of bacterial foodborne illness, often transmitted through contaminated food and water. *Salmonella* transmission dynamics in freshwater systems remain poorly understood. In this study, we collected samples from two rivers in Pichinca province, Ecuador daily over four days to determine *Salmonella* prevalence and characterize any genetic relationship between isolates at different points downstream and between rivers as well as repeated isolation of isolates at the same location on different days. These rivers, including an urban river and an agricultural river, contained four sites each. At each site, 10L water samples were collected using modified Moore swabs (n = 32) and cultured for *Salmonella*. All recovered isolates were analyzed by whole genome sequencing (WGS). Overall prevalence was 75% (24/32), where the urban river, Río Machángara, had a higher prevalence (87.5%; 14/16) than the agricultural river, Río San Pedro (62.5%; 10/16). Serotyping with WGS identified eight serovars, four of which had a high degree of genetic relatedness based on multilocus sequence typing and suggested possible transmission, persistence, or reintroduction events across sampling sites and days. Additionally, in positive samples where two or more isolates were recovered, 71.4% (10/14) contained multiple serovars. Importantly, both rivers contained clinically relevant serovars, which underscores the need for more robust environmental surveillance efforts to identify routes of foodborne pathogen transmission and reduce *Salmonella* illnesses.

4.2 Significance statement

Surface water contamination is a significant contributor to *Salmonella* illnesses worldwide; however, little is known about the ecology of this pathogen following

introduction to a water source. Thus, a knowledge gap exists in the understanding of diversity and long-term survivability of *Salmonella* within surface water sources. This study aimed to resolve these questions by analyzing *Salmonella* prevalence and serovar diversity in surface water samples collected in two important rivers in a highly populated region of Ecuador. Importantly, whole genome sequencing identified significant serovar overlap was identified between these rivers; however, more serovars were found in the urban river including those more commonly associated with human illness, demonstrating that human health risk is unequally distributed. Additionally, close WGS relationships between isolates downstream and across multiple days of sampling suggest potential transmission or persistence of *Salmonella* in these rivers. Overall, these results underscore the need to develop a One Health environmental monitoring approach that can be used to improve *Salmonella* control.

4.3 Introduction

Salmonella enterica is a leading cause of bacterial foodborne illness worldwide. This species is categorized into over 2,600 distinct serovars, which are defined by the combination of their somatic (O) and flagellar (H) antigens, according to the White-Kauffmann-LeMinor classification scheme (Grimont & Weill, 2007; Issenhuth-Jeanjean et al., 2014). Importantly, serovar-specific differences such as host restriction (Uzzau et al., 2000), virulence factors (Cheng et al., 2019), antimicrobial resistance (D. H. Shah et al., 2017), and stress response (Gorski & Noriega, 2023) can significantly influence the clinical relevance of a serovar. Subsequently, 100 serovars within *Salmonella enterica* subspecies *enterica* account for the majority of global human infections (Hendriksen et al., 2011; McVey et al., 2022; Park et al., 2009).

Subspecies *enterica* can be divided into two categories that separate typhoidal from non-typhoidal salmonellae (Crump et al., 2004). Serovars Typhi, Sendai, and Paratyphi (A, B, and C), are considered typhoidal serovars and the causative agents of typhoid (enteric) fever and paratyphoid fever (Gal-Mor et al., 2014). In Latin America, Sub-Saharan Africa, and Southeast Asia, where enteric fever is more common, illnesses are often attributed to ingestion of food or water contaminated with fecal material (IHME, 2021b, 2021a). Non-typhoidal serovars (NTS) are the causative agent of salmonellosis, a self-limiting gastroenteritis. This bacterium causes an estimated 150 million global illnesses and 60,000 deaths annually (CDC, 2024b).

As an enteric pathogen, *Salmonella* is often found in the gastrointestinal tract of animal hosts; however, *Salmonella* is also ubiquitous in the environment, and has been identified in various abiotic samples, especially in surface water such as creeks, rivers, ponds, and canals (Deaven et al., 2021; Haley et al., 2009; Khouja et al., 2024; You et al., 2006), but also soil and dust (Bardsley et al., 2021; Fahimipour et al., 2018; Jechalke et al., 2019b; Pal et al., 2021).

Salmonella can be introduced to water by direct or indirect fecal contamination by humans or animals (Casanova et al., 2020; Li et al., 2015; Topalcengiz et al., 2019). Direct contamination includes defecation into the water, or sewage release into the water, while indirect includes run-off from the land following rain. Once introduced, *Salmonella* can survive within biofilms, free-living protozoa, or free-floating in planktonic or viable but non-culturable (VBNC) states (Brandl et al., 2005; Byappanahalli et al., 2009; Gaertner et al., 2011; Oguadinma et al., 2022). Supporting this, a study showed long-term viability of *Salmonella* in a lab experiment where it was recovered in freshwater after

over 300 days post-inoculation. There is the potential for even longer survival in an open system when nutrients are regularly cycled through, which would occur with frequent introduction of fecal material into surface water (Topalcengiz et al., 2019). This ability to survive long periods in non-host environments likely explains why *Salmonella* has been regularly isolated from various aquatic sources such as urban wastewater, runoff caused by rain, and contaminated water due to agricultural waste (Berge et al., 2006; Casanova et al., 2020; Magana-Arachchi & Wanigatunge, 2020; Murphy et al., 2022; Santiago et al., 2018; Toro et al., 2022).

Food outbreak traceback capabilities were improved drastically with the development and implementation of whole genome sequencing (WGS) technology as a highly discriminatory typing tool (Allard et al., 2012, 2016; Timme et al., 2013). While individual whole genome sequences provide important genetic information, a robust database, such as GenomeTrakr, generates the capacity to assess isolates within the wider context of almost 750,000 *Salmonella* genomes (FDA, 2025). Broadly, WGS analysis utilizes single nucleotide polymorphisms (SNPs) to compare the sequence fidelity between a set of isolates. One form of this comparison is core genome multilocus sequence typing (cgMLST), which assesses the genomic relationship of isolates based on the sequence of genes present in 95% of a given dataset (Zhou et al., 2021b).

Pichincha lies in the Northern Sierra of Ecuador and, with a population of over three million, is the second most populated province in Ecuador (City Populations, 2024). This region is home to the Ecuadorian capital of Quito; the second highest capital in the world at 2,850m. As such, the population has increased by more than 60% and urban expansion by 7.1% since 2000 (Atlas of Urban Expansion, 2016; Carrión & Erazo

Espinosa, 2012). These changes are followed by additional stress on the environment, contributing to increases in chemical and microbial intrusion to water sources (Adler Miserendino et al., 2013; Benítez et al., 2019). In addition to direct human contributions to water pollution such as sewage, indirect runoff has the potential to collect bacteria and nutrients that contribute to unsafe water conditions. Importantly, these conditions can be exacerbated when runoff occurs over land where higher concentrations of *Salmonella* are present. The common use of rivers throughout this region for livestock, agricultural, and domestic activities, including livestock interactions, crop irrigation, washing clothes, and bathing in contaminated waters can produce severe health effects for the inhabitants (Borja-Serrano et al., 2020).

The Río Machángara is considered one of the main rivers in Pichincha; its headwaters lie at the base of Atacazo volcano near the southern tip of Quito, where it flows for 40km before entering Río Guayllabamba. This river in turn is a tributary to the Río Esmeraldas that flows into the Pacific Ocean. From Quito, water from the Río Machángara crosses three different Ecuadorian provinces before reaching the ocean. The Río San Pedro is also in Pichincha province. It begins high in the volcanic region of Iliniza Sur and flows into the Río Machángara just east of Quito. Along its 73 km course the Río San Pedro flows through agricultural land, including dense dairy cattle production in the Mejía canton, which includes the towns of El Chaupi (population 1,425) and Machachi (population 32,814), which lie at 3,386m and 2,933m above sea level, respectively.

The Río Machángara acts as the destination for untreated wastewater, collecting approximately 70% of the city's waste, including effluent and contaminants (Campaña et

al., 2017). Wastewater treatment facilities present at the city exit prevent untreated waste from extending beyond the heavily-populated area; however, this does little to prevent contamination within the city. For this reason, this river has been studied in multiple projects to measure chemical and microbiological traces (Alomía Herrera & Carrera Burneo, 2017; Méndez et al., 2022; Ortega-Paredes et al., 2020; Vinueza et al., 2021; Vizcaíno et al., 2016). These studies have recently led to a monumental impact: following Ecuador becoming the first country in the world to recognize the Rights of Nature in its constitution, in July 2024, an Ecuadorian judge ruled that pollution has violated the rights of the Río Machángara, and called for the city of Quito to clean the river (Altamirano Cardenas & Cervantes Galván, 2024).

Although several investigations have been completed regarding the presence of *Salmonella* in water sources, further characterization and investigation of transmission within rivers is understudied. The comparable sizes and contrasting creeks of the Rios Machángara and San Pedro in the Pichincha province lend themselves to a useful study to compare *Salmonella* across two individual rivers situated in the same region but with very different surrounding land-use. This study aimed to determine the presence and diversity of *Salmonella* over four sites in each river that were sampled repeatedly over four days. Using WGS, we sought to assess whether *Salmonella* was being re-introduced to the river daily and whether it was being transmitted downstream based on the genetic relatedness of isolates.

4.4 Materials and methods

4.4.1 Site selection

Two distinct rivers were chosen as the subject of this study based on land attribution characteristics associated with each river and both lie in the Pichincha province (Figure 4.1A). Four sites were selected along each river based on different land use and ease of access. The four sites along the Río Machángara in Quito were chosen to reflect the river just after entering the city (site 1), two points within the city (sites 2 and 3), and a final site after the river exits the city (site 4) (Figure 4.1B).

For the Río San Pedro, the four sites were selected to cover areas including upstream (site 1), within (site 2) and downstream (site 3) of the major agriculturally concentrated areas as well as just prior to entering urban populations (site 4) (Figure 4.1C). The Río San Pedro flows into the Río Machángara; the last sampling site on the Río Machángara is upstream of where the Río San Pedro enters.

4.4.2 Sample collection and *Salmonella* isolation

A modified Moore swab and peristaltic pump were used to collect 10 L of creek water from each site (n = 32 samples) as described (Deaven et al., 2021). The swabs were stored on ice until returning to the lab, where 100 ml of buffered peptone water (BPW; BD Difco, Franklin Lakes, NJ) and novobiocin (40 mg/L; Thermo Scientific Chemicals, Waltham, MA) were added. Swabs were each hand-massaged for 1 minute, then incubated at 42°C for 20-24 hours. Following incubation, 1 and 0.1 mL of BPW were transferred into 12.2 mL of tetrathionate (TT; BD Difco Franklin Lakes, NJ) and 9.9mL Rappaport-Vassiliadis (RV; BD Difco Franklin Lakes, NJ) broth, respectively, and incubated at 37°C for 20-24 hours. Iodine-iodide solution was not readily available when

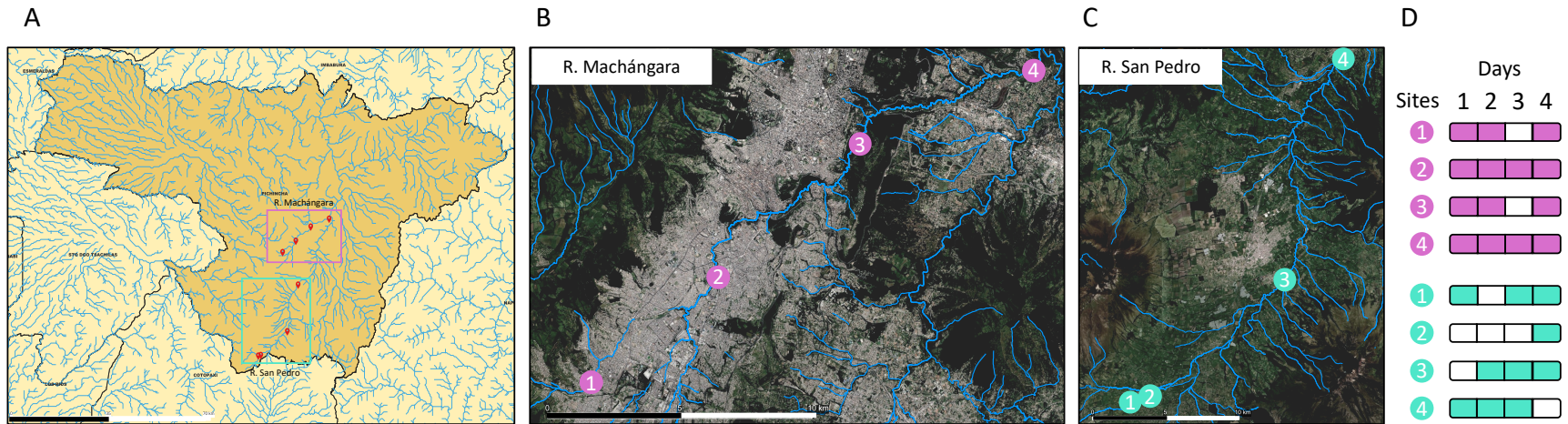


Figure 4.1. Sampling site distribution and *Salmonella* prevalence of two rivers in Pichincha, Ecuador. Regional map of Pichincha (A) indicates the relative location of both rivers sampled in this study (Río Machángara in purple and Río San Pedro in green). Sampling sites of Río Machángara (B) and Río San Pedro (C) are shown along with respective tributaries. *Salmonella* prevalence (D) is shown as colored boxes representing positive samples for Río Machángara (purple, top) and Río San Pedro (green, bottom).

we were performing this work. Therefore, prior to the study, we tested multiple commercially available iodine products as potential replacements including potassium triiodide, povidone, and an iodine tincture. In a comparison of *Salmonella* recovery in TT broth between traditional iodine-iodide solution and these alternatives, povidone demonstrated consistent and comparable recovery resulting in its use during this study (data not shown). To achieve the same concentration as is typically used, 3.2 ml of 10% povidone (1% free iodine) were added to 9 ml of TT for a final volume of 12.2 ml. Additionally, 0.1 mL of the pre-enriched BPW culture was inoculated into modified semisolid Rappaport-Vassiliadis (MSRV; BD Difco Franklin Lakes, NJ) and incubated at 37°C for 24-48 hours. *Salmonella* presence was tested by streaking the TT and RV enrichments, along with any presumptive *Salmonella* growth from the MSRV plates, onto xylose lysine deoxycholate (XLD; BD Difco Franklin Lakes, NJ) plates with novobiocin (40mg/L) followed by incubation at 37°C for 24-48 hours. Up to three colonies were selected (based on differing morphologies) from each plate and restreaked for isolation onto fresh XLD plates. After incubation at 37°C for 24-48 hours, a single colony was streaked onto Luria-Bertani (LB; Hardy Diagnostics, Santa Maria, CA) agar. These were confirmed as *Salmonella* using poly A-I serum agglutination (BD Difco, Franklin Lakes, NJ). All confirmed *Salmonella* isolates were grown overnight in LB broth and then stored at -80°C in 20% glycerol.

4.4.3 Whole genome sequencing

Stored glycerol stocks were used to propagate *Salmonella* isolates. Stocks were streaked onto LB agar and grown overnight. A single colony was picked from the LB plate and used to start an overnight growth. DNA was isolated from overnight growth and

shipped to the FDA Center for Food Safety and Applied Nutrition (FDA-CFSAN) for library preparation and sequencing. Enterobase was used to analyze the whole genome sequences for serotyping and core genome MLST analysis (Zhou et al., 2020b, 2021b). SeqSero and SISTR (as part of Enterobase) were used for serovar identification (Yoshida et al., 2016; Zhang et al., 2019).

4.5 Results

Over four consecutive days, water samples were collected at eight sites across two rivers in the Pichincha region in Ecuador ($n = 32$). The Río Machángara and Río San Pedro contained four sites each, where the rivers primarily flowed through a dense urban environment and animal agriculture, respectively (Figure 4.1). In total, 75% (24/32) of samples were positive, where the Río Machángara and Río San Pedro had an overall prevalence of 87.5% (14/16) and 62.5% (10/16), respectively. Daily incidence ranged from 75% (3/4) to 100% (4/4) in the Río Machángara and 25% (1/4) to 75% (3/4) in the Río San Pedro.

Whole genome sequencing (WGS) was completed on 54 isolates, ranging from one to six isolates per sample. A total of eight serovars were identified, where serovars Uganda, Typhimurium, I 4,[5],12:i:-, Infantis, and Anatum were identified in both rivers (Figure 4.2). Serovars Amager and Derby were unique to the Río Machángara, and serovar Manhattan was unique to the Río San Pedro. Serovar Uganda was the most frequent serovar isolated and was present in 50% (13/24) of positive samples, including 71% (10/14) and 30% (3/10) of samples from the Río Machángara and Río San Pedro, respectively. Serovar Typhimurium and its monophasic variant I 4,[5],12:I:-, were also common, with isolation from 25% (6/24) and 21% (5/24) of positive samples,

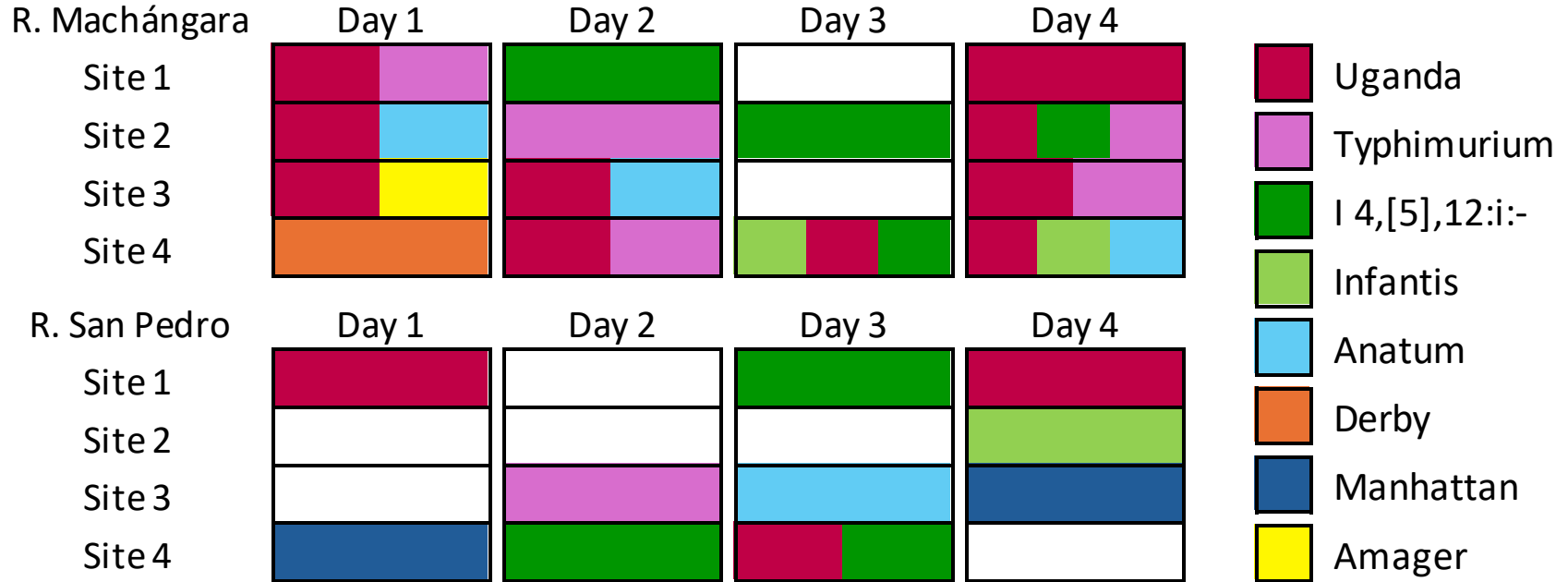


Figure 4.2. Serovar identification from whole genome sequencing of river isolates. Río Machángara (top) and Río San Pedro (Bottom) positive samples are shown as colored squares. Samples are represented as one, two, or three colors showing total number of serovars identified (right). Negative samples are shown as white squares.

respectively. In 14 positive samples, two or more isolates were recovered, resulting in the identification of multiple serovars in 71.4% (10/14) of samples, including 64% (9/14) of samples from the Río Machángara and 10% (1/10) of samples in the Río San Pedro.

Enterobase was used to designate a cgMLST to each isolate (Zhou et al., 2020b). Additionally, hierarchical clustering (HC) was used to group related isolates based on allelic differences. Isolates within two allelic differences (i.e. sharing the same HC2 cgMLST) are considered highly related (Zhou et al., 2021b). We assigned HC2 sequence types to isolates belonging to serovars that were identified in three or more samples and used this to look at relationships between different isolates in our study. Serovar Uganda isolates were categorized into three sequence types, where 54% (7/13) of samples containing this serovar belonged to HC2 472322 type (Figure 4.3). Less closely related, HC2 472380 had between 5 and 10 allelic differences from the prominent HC2 472322 type. Interestingly, the remaining serovar Uganda cluster, HC2 472372, had a distant relationship of more than 50 allelic differences from the other clusters. Isolates within this sequence type were recovered from all four sites of the Río Machángara and in three out of four days in this river. Notably, all isolates of serovars Anatum and I 4,[5],12:I:- belonged to a single sequence type, respectively. Patterns of related isolated between sampling sites or across days at the same site were marked with arrows or circles, respectively (Figure 4.3). For example, this pattern was identified for HC2 472322 was recovered in the first three three sites of the Río Machángara on day one and on site four of day two. Additionally, this same cluster was found in site one on days one and four in the Río Machángara.

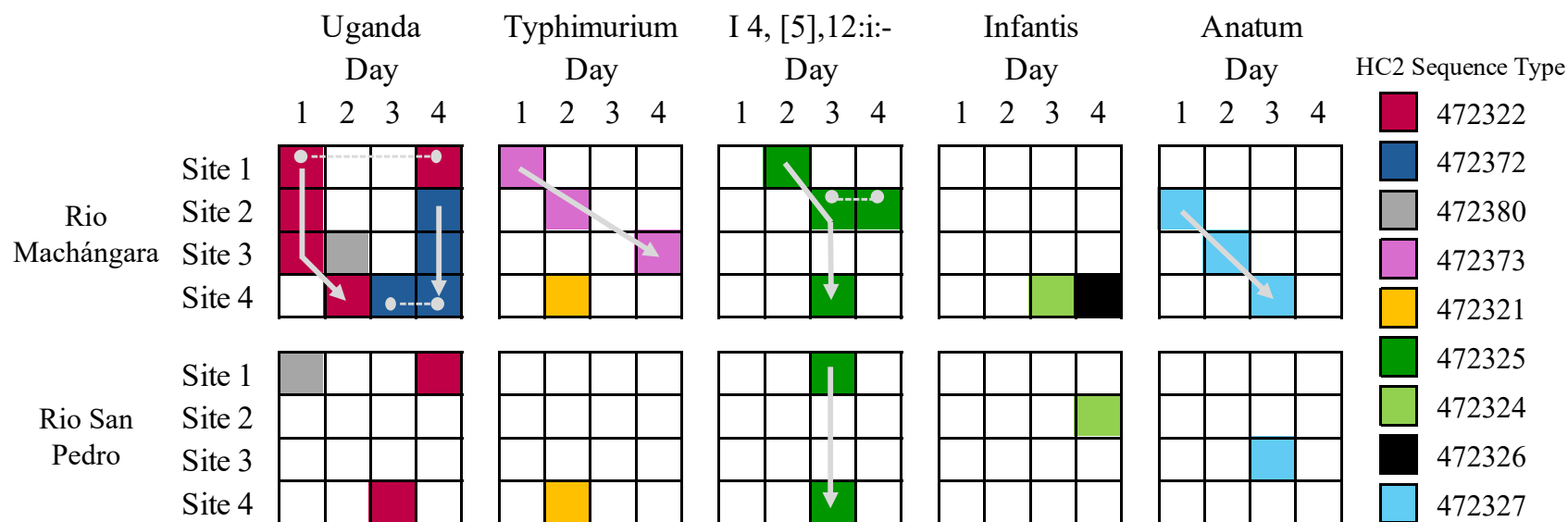


Figure 4.3. cgMLST highlights trends of *Salmonella* persistence and transmission. Study isolates sharing core genomes within two allelic differences (HC2) are shown by color-coded squares. Possible connections of transmission and persistence are indicated with solid gray arrows and dashed gray lines with circles.

4.6 Discussion

Several studies have been conducted on both Río San Pedro and Río Machángara to identify various pathogenic microorganisms, including studies that specifically looked for *Salmonella* (Borja-Serrano et al., 2020; Campaña et al., 2017; Ortega-Paredes et al., 2020; Vinueza et al., 2021). Interestingly, these studies were unable to identify the presence of *Salmonella* in their samples, however, high levels of *E. coli* and fecal coliforms were detected. Both studies utilized membrane filtration, a method consisting of collecting and filtering water from an 800 mL grab sample through a nitrocellulose membrane (Entis, 1990; Meinersmann et al., 2008). A previous study has successfully identified *Salmonella* prevalence using DNA extracted from resuspended sediment in Bangladesh following membrane filtration with no enrichment (Amin et al., 2020). The use of the modified Moore swabs in this study provided an advantage of filtering 10L of water, increasing the likelihood of detecting *Salmonella* when present at low concentrations (Sharma et al., 2020). The use of selective enrichment is advantageous for *Salmonella* recovery as the growth of nontarget organisms is suppressed, which may explain why *Salmonella* was not detected in previous studies. Therefore, this current study represents the first to recover and characterize *Salmonella* in these rivers.

Recently, a collaborative research effort, known as the Joint Institute for Food Safety and Applied Nutrition (JIFSAN), has brought together research groups throughout Latin America, the University of Maryland, and the US Food and Drug Administration. This group has carried out surveillance *Salmonella* testing in surface water in Chile, Brazil, and Mexico, utilizing similar sampling and culturing methods as those described in this study (Toro et al., 2022). One of these studies similarly compares rates of

Salmonella incidence in urban areas (29.1%) compared to agricultural areas (27.0%) (Toro et al., 2022). That study found that seasonality and water type (creek or canal versus river or pond) were the most significant contributors to *Salmonella* presence. In another integrated study conducted in Chile, Brazil, and Mexico, samples were taken from surface waters and showed variable *Salmonella* prevalence: 33%, 63%, and 68%, respectively (Chen et al., 2024). Notably, serovars Infantis and Typhimurium were frequently identified in both the integrated study and this current work. Additional work in Mexico has surveyed *Salmonella* prevalence and serovar diversity in multiple projects (González-López et al., 2022; Jiménez et al., 2014). These studies have identified 24 and 27 serovar from samples collected in 2008-2009 and 2018-2019, respectively. An additional study compared isolates of serovar Oranienburg collected in these studies and found close genomic relationships, regardless of the year of isolation (González-Torres et al., 2023).

Previous studies in Ecuador identifying *Salmonella* serovars have been limited to those in animal agriculture. These studies have reported the presence of many serovars identified in this work, including Infantis, Uganda, I 4,[5],12:i:-, Typhimurium, Manhattan, and Derby (Medina-Santana et al., 2022; Vinueza-Burgos et al., 2016, 2024). Importantly, serovar Amager was identified for the first time in Ecuador in this study; however, this serovar has been isolated from chicken carcasses in Venezuela (Boscán-Duque et al., 2007) and pigs in Colombia (Chamorro-Tobar et al., 2024). The largest *Salmonella* sequence repositories are the National Center for Biological information (NCBI) and Enterobase. While they provide a global repository for whole genome sequences, they contain significantly fewer *Salmonella* isolates from Ecuador and other

countries in Latin America, which limits possible comparative analyses between studies. This highlights a need for greater equitable pathogen surveillance across the world to improve food safety on a global level.

WGS results demonstrated that some serovars that were collected from the same river were represented by up to three different cgMLSTs. Interestingly, highly related isolates of serovars Typhimurium, Infantis, and Anatum were found in both rivers. While these rivers eventually merge, the sampling sites were upstream of this location, suggesting that a shared source of transmission may be responsible for these findings such as wildlife or other animal activities, or via humans (Kagambèga et al., 2017). Additionally, evidence of persistence and transmission were identified throughout this study. Identical cgMLSTs were found at the same sampling points on different days, suggesting the reintroduction or persistence of the same strain over multiple days. As the river flows downstream, we would expect any persistent *Salmonella* to be somewhat stationary, such as within biofilms on sediment (Gaertner et al., 2011; Sha et al., 2013). Given that sampling was conducted in moving water and not the sediment at the base, additional serovars or potentially persistent strains could be identified with alternative sampling methods. Evidence of *Salmonella* transmission was supported by finding identical or highly related cgMLSTs at different sampling locations on the same day. This could also be explained by regular contamination events from a common host.

This study demonstrates the importance of routine environmental *Salmonella* surveillance to better understand exposure to pathogenic bacteria, such as *Salmonella*. During the week we sampled, *Salmonella* prevalence was high and we also isolated multiple clinically relevant serovars. In many cases, more than one serovar was identified

from a sample, highlighting the need to more fully elucidate multiserovar *Salmonella* populations within environmental samples. While limited to four sequential sampling days, this study establishes a potential framework for more routine sampling of surface water sources. Longitudinal studies would be helpful in determining whether any seasonal or weather patterns influence *Salmonella* in these rivers. Patterns of persistence and transmission were also identified throughout this study, suggesting a need to identify routes of *Salmonella* contamination in these rivers and apply management strategies to reduce human health risks associated with contaminated surface water.

4.7 Data availability statement

Whole genome sequences are available on NCBI under the following isolate names: CFSAN137150-CFSAN137204.

4.8 Funding

This work was supported through funds from the University of Georgia to NWS. Sequencing for this project was funded through the University of Georgia Graduate School summer research grant awarded to JCS.

4.9 Acknowledgements

We thank members of the Vinueza lab for their technical assistance and Dr. Chris Grim at the FDA-CFSAN for sequencing support.

4.10 Competing interest statement

NWS has a licensing agreement with Ancera Inc.

4.11 Author Contributions

JCS, DAV, and NWS conceived and planned the study. JCS and DAV performed the experiments, completed the analysis and wrote the first draft of the paper. CVB

contributed laboratory support in Ecuador. All authors reviewed, edited, and approved the paper.

4.12 References

- Adler Miserendino, R., Bergquist, B. A., Adler, S. E., Guimarães, J. R. D., Lees, P. S. J., Niquen, W., Velasquez-López, P. C., & Veiga, M. M. (2013). Challenges to measuring, monitoring, and addressing the cumulative impacts of artisanal and small-scale gold mining in Ecuador. *Resources Policy*, 38(4), 713–722. <https://doi.org/10.1016/j.resourpol.2013.03.007>
- Allard, M. W., Luo, Y., Strain, E., Li, C., Keys, C. E., Son, I., Stones, R., Musser, S. M., & Brown, E. W. (2012). High resolution clustering of *Salmonella enterica* serovar Montevideo strains using a next-generation sequencing approach. *BMC Genomics*, 13(1), 32. <https://doi.org/10.1186/1471-2164-13-32>
- Allard, M. W., Strain, E., Melka, D., Bunning, K., Musser, S. M., Brown, E. W., & Timme, R. (2016). Practical Value of Food Pathogen Traceability through Building a Whole-Genome Sequencing Network and Database. *Journal of Clinical Microbiology*, 54(8), 1975–1983. <https://doi.org/10.1128/JCM.00081-16>
- Alomía Herrera, I., & Carrera Burneo, P. (2017). Environmental flow assessment in Andean rivers of Ecuador, case study: Chanlud and El Labrado dams in the Machángara River. *Ecohydrology & Hydrobiology*, 17(2), 103–112. <https://doi.org/10.1016/j.ecohyd.2017.01.002>
- Altamirano Cardenas, F. I., & Cervantes Galván, G. H. (2024, August). Quito Court Ruling on Rights of Machángara River 2024. (Accessed January 11, 2025)

<https://ecojurisprudence.org/initiatives/ecuador-court-case-on-the-rights-of-the-machangara-river/>.

Amin, N., Liu, P., Foster, T., Rahman, M., Miah, M. R., Ahmed, G. B., Kabir, M., Raj, S., Moe, C. L., & Willetts, J. (2020). Pathogen flows from on-site sanitation systems in low-income urban neighborhoods, Dhaka: A quantitative environmental assessment. *International Journal of Hygiene and Environmental Health*, 230, 113619. <https://doi.org/10.1016/j.ijheh.2020.113619>

Atlas of Urban Expansion. (2016). Quito: Areas and Densities. (Accessed January 15, 2025) <http://www.atlasofurbanexpansion.org/cities/view/quito>.

Benítez, M. B., Champagne, P., Ramos, A., Torres, A. F., & Ochoa-Herrera, V. (2019). Wastewater treatment for nutrient removal with Ecuadorian native microalgae. *Environmental Technology*, 40(22), 2977–2985. <https://doi.org/10.1080/09593330.2018.1459874>

Berge, A. C. B., Dueger, E. L., & Sischo, W. M. (2006). Comparison of *Salmonella enterica* serovar distribution and antibiotic resistance patterns in wastewater at municipal water treatment plants in two California cities. *Journal of Applied Microbiology*, 101(6), 1309–1316. <https://doi.org/10.1111/j.1365-2672.2006.03031.x>

Borja-Serrano, P., Ochoa-Herrera, V., Maurice, L., Morales, G., Quilumbaqui, C., Tejera, E., & Machado, A. (2020). Determination of the Microbial and Chemical Loads in Rivers from the Quito Capital Province of Ecuador (Pichincha)—A Preliminary Analysis of Microbial and Chemical Quality of the Main Rivers. *International*

- Journal of Environmental Research and Public Health, 17(14), 5048.
<https://doi.org/10.3390/ijerph17145048>
- Boscán-Duque, L. A., Arzálluz-Fisher, A. M., Ugarte, C., Sánchez, D., Wittum, T. E., & Hoet, A. E. (2007). Reduced Susceptibility to Quinolones among *Salmonella* Serotypes Isolated from Poultry at Slaughter in Venezuela. Journal of Food Protection, 70(9), 2030–2035. <https://doi.org/10.4315/0362-028X-70.9.2030>
- Brandl, M. T., Rosenthal, B. M., Haxo, A. F., & Berk, S. G. (2005). Enhanced Survival of *Salmonella enterica* in Vesicles Released by a Soilborne Tetrahymena Species. Applied and Environmental Microbiology, 71(3), 1562–1569.
<https://doi.org/10.1128/AEM.71.3.1562-1569.2005>
- Byappanahalli, M. N., Sawdey, R., Ishii, S., Shively, D. A., Ferguson, J. A., Whitman, R. L., & Sadowsky, M. J. (2009). Seasonal stability of Cladophora-associated *Salmonella* in Lake Michigan watersheds. Water Research, 43(3), 806–814.
<https://doi.org/10.1016/j.watres.2008.11.012>
- Campaña, A., Gualoto, E., & Chiluisa-Utreras, V. (2017). Evaluación físico-química y microbiológica de la calidad del agua de los ríos Machángara y Monjas de la red hídrica del distrito metropolitano de Quito. Bionatura, 2(2), 305–310.
<https://doi.org/10.21931/RB/2017.02.02.6>
- Carrión, F., & Erazo Espinosa, J. (2012). La forma urbana de Quito: una historia de centros y periferias. Bulletin de l’Institut Français d’études Andines, 41 (3), 503–522. <https://doi.org/10.4000/bifea.361>

- Casanova, L. M., Hill, V. R., & Sobsey, M. D. (2020). Antibiotic-resistant *Salmonella* in swine wastes and farm surface waters. *Letters in Applied Microbiology*, 71(1), 117–123. <https://doi.org/10.1111/lam.13242>
- CDC. (2024). *CDC Yellow Book: Health Information for International Travel*. Oxford University Press. (Accessed January 14, 2025).
<https://wwwnc.cdc.gov/travel/page/yellowbook-home>.
- Chamorro-Tobar, I. C., Pulido-Villamarín, A., Carrascal-Camacho, A. K., Barrientos-Anzola, I., Wiesner, M., Hernández-Toro, I., Alban, L., Olsen, J. E., Dalsgaard, A., & Hounmanou, Y. M. G. (2024). Phenotypic and Genotypic Characterization of Antimicrobial Resistance in *Salmonella enterica* Serovars from Colombian Pig Farms. *Applied Microbiology*, 4(4), 1729–1744.
<https://doi.org/10.3390/applmicrobiol4040116>
- Chen, Z., Moreno-Switt, A. I., Reyes-Jara, A., Delgado Suarez, E., Adell, A. D., Oliveira, C. J. B., Bonelli, R. R., Huang, X., Brown, E., Allard, M., Grim, C., Bell, R., Meng, J., & Toro, M. (2024). A multicenter genomic epidemiological investigation in Brazil, Chile, and Mexico reveals the diversity and persistence of *Salmonella* populations in surface waters. *MBio*, 15(7), e0077724.
<https://doi.org/10.1128/mbio.00777-24>
- Cheng, R. A., Eade, C. R., & Wiedmann, M. (2019). Embracing Diversity: Differences in Virulence Mechanisms, Disease Severity, and Host Adaptations Contribute to the Success of Nontyphoidal *Salmonella* as a Foodborne Pathogen. *Frontiers in Microbiology*, 10, 1368. <https://doi.org/10.3389/fmicb.2019.01368>

- City Populations. (2024). Ecuador: Localities. (Accessed January 15th, 2025)
<https://www.citypopulation.de/en/ecuador/towns/>.
- Crump, J. A., Luby, S. P., & Mintz, E. D. (2004). The global burden of typhoid fever. *Bulletin of the World Health Organization*, 82(5), 346–353.
- Deaven, A. M., Ferreira, C. M., Reed, E. A., Chen See, J. R., Lee, N. A., Almaraz, E., Rios, P. C., Marogi, J. G., Lamendella, R., Zheng, J., Bell, R. L., & Shariat, N. W. (2021). *Salmonella* Genomics and Population Analyses Reveal High Inter- and Intraserovar Diversity in Freshwater. *Applied and Environmental Microbiology*, 87(6). <https://doi.org/10.1128/AEM.02594-20>
- Entis, P. (1990). Improved Hydrophobic Grid Membrane Filter Method, Using EF-18 Agar, for Detection of *Salmonella* in Foods: Collaborative Study. *Journal of AOAC International*, 73(5), 734–742. <https://doi.org/10.1093/jaoac/73.5.734>
- FDA. (2025). GenomeTrakr Fast Facts. (Accessed January 8, 2025).
<https://www.fda.gov/food/whole-genome-sequencing-wgs-program/genometrakr-fast-facts>.
- Gaertner, J. P., Mendoza, J. A., Forstner, M. R. J., & Hahn, D. (2011). Recovery of *Salmonella* from biofilms in a headwater spring ecosystem. *Journal of Water and Health*, 9(3), 458–466. <https://doi.org/10.2166/wh.2011.173>
- Gal-Mor, O., Boyle, E. C., & Grassl, G. A. (2014). Same species, different diseases: how and why typhoidal and non-typhoidal *Salmonella enterica* serovars differ. *Frontiers in Microbiology*, 5. <https://doi.org/10.3389/fmicb.2014.00391>
- González-López, I., Medrano-Félix, J. A., Castro-del Campo, N., López-Cuevas, O., González-Gómez, J. P., Valdez-Torres, J. B., Aguirre-Sánchez, J. R., Martínez-

- Urtaza, J., Gómez-Gil, B., Lee, B. G., Quiñones, B., & Chaidez, C. (2022). Prevalence and Genomic Diversity of *Salmonella enterica* Recovered from River Water in a Major Agricultural Region in Northwestern Mexico. *Microorganisms*, 10(6), 1214. <https://doi.org/10.3390/microorganisms10061214>
- González-Torres, B., González-Gómez, J. P., Ramírez, K., Castro-del Campo, N., González-López, I., Garrido-Palazuelos, L. I., Chaidez, C., & Medrano-Félix, J. A. (2023). Population structure of the *Salmonella enterica* serotype Oranienburg reveals similar virulence, regardless of isolation years and sources. *Gene*, 851, 146966. <https://doi.org/10.1016/j.gene.2022.146966>
- Gorski, L., & Noriega, A. A. (2023). Comparison of Phenotype Nutritional Profiles and Phosphate Metabolism Genes in Four Serovars of *Salmonella enterica* from Water Sources. *Microorganisms*, 11(8), 2109. <https://doi.org/10.3390/microorganisms11082109>
- Grimont, P., & Weill, F.-X. (2007). Antigenic Formulae of the *Salmonella* serovars.
- Haley, B. J., Cole, D. J., & Lipp, E. K. (2009). Distribution, Diversity, and Seasonality of Waterborne Salmonellae in a Rural Watershed. *Applied and Environmental Microbiology*, 75(5), 1248–1255. <https://doi.org/10.1128/AEM.01648-08>
- Hendriksen, R. S., Vieira, A. R., Karlsmose, S., Lo Fo Wong, D. M. A., Jensen, A. B., Wegener, H. C., & Aarestrup, F. M. (2011). Global Monitoring of *Salmonella* Serovar Distribution from the World Health Organization Global Foodborne Infections Network Country Data Bank: Results of Quality Assured Laboratories from 2001 to 2007. *Foodborne Pathogens and Disease*, 8(8), 887–900. <https://doi.org/10.1089/fpd.2010.0787>

- IHME. (2021a). Paratyphoid fever—Level 4 cause. (Accessed January 20, 2025)
<https://www.healthdata.org/research-analysis/diseases-injuries-risks/factsheets/2021-paratyphoid-fever-level-4-disease>.
- IHME. (2021b). Typhoid fever—Level 4 cause. (Accessed January 20, 2025)
<https://www.healthdata.org/research-analysis/diseases-injuries-risks/factsheets/2021-typhoid-fever-level-4-disease>.
- Issenhuth-Jeanjean, S., Roggentin, P., Mikoleit, M., Guibourdenche, M., de Pinna, E., Nair, S., Fields, P. I., & Weill, F.-X. (2014). Supplement 2008–2010 (no. 48) to the White–Kauffmann–Le Minor scheme. *Research in Microbiology*, 165(7), 526–530. <https://doi.org/10.1016/j.resmic.2014.07.004>
- Jechalke, S., Schierstaedt, J., Becker, M., Flemer, B., Grosch, R., Smalla, K., & Schikora, A. (2019). *Salmonella* Establishment in Agricultural Soil and Colonization of Crop Plants Depend on Soil Type and Plant Species. *Frontiers in Microbiology*, 10. <https://doi.org/10.3389/fmicb.2019.00967>
- Jiménez, M., Martínez-Urtaza, J., Rodríguez-Alvarez, M. X., Leon-Felix, J., & Chaidez, C. (2014). Prevalence and genetic diversity of *Salmonella* spp. in a river in a tropical environment in Mexico. *Journal of Water and Health*, 12(4), 874–884. <https://doi.org/10.2166/wh.2014.051>
- Kagambèga, A., Bouda, S. C., Bako, E., Cissé, H., Barro, N., & Haukka, K. (2017). Diversity and antimicrobial resistance of *Salmonella* strains isolated from different sources in Burkina Faso. *African Journal of Microbiology Research*, 11(40), 1495–1504. <https://doi.org/10.5897/AJMR2017.8698>

- Khouja, B. A., Salazar, J. K., Babaria, H., Fay, M. L., & Stewart, D. S. (2024). Method of Inoculation Influences the Survival of *Salmonella enterica* on Retail and Orchard Peaches. *Journal of Food Protection*, 87(6), 100289.
<https://doi.org/10.1016/j.jfp.2024.100289>
- Li, B., Jackson, S. A., Gangiredla, J., Wang, W., Liu, H., Tall, B. D., Beaubrun, J. J.-G., Jay-Russell, M., Vellidis, G., & Elkins, C. A. (2015). Genomic Evidence Reveals Numerous *Salmonella enterica* Serovar Newport Reintroduction Events in Suwannee Watershed Irrigation Ponds. *Applied and Environmental Microbiology*, 81(24), 8243–8253. <https://doi.org/10.1128/AEM.02179-15>
- Magana-Arachchi, D. N., & Wanigatunge, R. P. (2020). Ubiquitous waterborne pathogens. In *Waterborne Pathogens* (pp. 15–42). Elsevier.
<https://doi.org/10.1016/B978-0-12-818783-8.00002-5>
- McVey, S., Kennedy, M., Chengappa, M., & Wilkes, R. (2022). *Veterinary Microbiology* (5th Ed). Willey Blackwell.
- Medina-Santana, J. L., Ortega-Paredes, D., de Janon, S., Burnett, E., Ishida, M., Sauders, B., Stevens, M., & Vinueza-Burgos, C. (2022). Investigating the dynamics of *Salmonella* contamination in integrated poultry companies using a whole genome sequencing approach. *Poultry Science*, 101(2), 101611.
<https://doi.org/10.1016/j.psj.2021.101611>
- Meinersmann, R. J., Berrang, M. E., Jackson, C. R., Fedorka-Cray, P., Ladely, S., Little, E., Frye, J. G., & Mattsson, B. (2008). *Salmonella*, *Campylobacter* and *Enterococcus* spp.: Their Antimicrobial Resistance Profiles and their Spatial

- Relationships in a Synoptic Study of the Upper Oconee River Basin. *Microbial Ecology*, 55(3), 444–452. <https://doi.org/10.1007/s00248-007-9290-6>
- Méndez, G., Vásquez, K., & Coyago, E. (2022). Bacteria Identification in Machángara River Water Capable of Metabolizing Emerging Estrone Pollutant (pp. 21–30). https://doi.org/10.1007/978-981-16-4126-8_3
- Murphy, C. M., Strawn, L. K., Chapin, T. K., McEgan, R., Gopidi, S., Friedrich, L., Goodridge, L. D., Weller, D. L., Schneider, K. R., & Danyluk, M. D. (2022). Factors Associated With *E. coli* Levels in and *Salmonella* Contamination of Agricultural Water Differed Between North and South Florida Waterways. *Frontiers in Water*, 3. <https://doi.org/10.3389/frwa.2021.750673>
- Oguadinma, I. C., Mishra, A., & Kumar, G. D. (2022). Sunlight Parameters Influence the Survival and Decline of *Salmonella* and *Escherichia coli* in Water. *Journal of Food Protection*, 85(11), 1614–1624. <https://doi.org/10.4315/JFP-22-042>
- Ortega-Paredes, D., Barba, P., Mena-López, S., Espinel, N., Crespo, V., & Zurita, J. (2020). High quantities of multidrug-resistant *Escherichia coli* are present in the Machángara urban river in Quito, Ecuador. *Journal of Water and Health*, 18(1), 67–76. <https://doi.org/10.2166/wh.2019.195>
- Pal, A., Bailey, M. A., Talorico, A. A., Krehling, J. T., Macklin, K. S., Price, S. B., Buhr, R. J., & Bourassa, D. V. (2021). Impact of poultry litter *Salmonella* levels and moisture on transfer of *Salmonella* through associated in vitro generated dust. *Poultry Science*, 100(8), 101236. <https://doi.org/10.1016/j.psj.2021.101236>
- Park, S. H., Kim, H. J., Cho, W. H., Kim, J. H., Oh, M. H., Kim, S. H., Lee, B. K., Ricke, S. C., & Kim, H. Y. (2009). Identification of *Salmonella enterica* subspecies I,

Salmonella enterica serovars Typhimurium, Enteritidis and Typhi using multiplex PCR. FEMS Microbiology Letters, 301(1), 137–146.

<https://doi.org/10.1111/j.1574-6968.2009.01809.x>

Santiago, P., Jiménez-Belenguer, A., García-Hernández, J., Estellés, R. M., Hernández Pérez, M., Castillo López, M. A., Ferrús, M. A., & Moreno, Y. (2018). High prevalence of *Salmonella* spp. in wastewater reused for irrigation assessed by molecular methods. International Journal of Hygiene and Environmental Health, 221(1), 95–101. <https://doi.org/10.1016/j.ijheh.2017.10.007>

Sha, Q., Vattam, D. A., Forstner, M. R. J., & Hahn, D. (2013). Quantifying *Salmonella* Population Dynamics in Water and Biofilms. Microbial Ecology, 65(1), 60–67. <https://doi.org/10.1007/s00248-012-0106-y>

Shah, D. H., Paul, N. C., Sischo, W. C., Crespo, R., & Guard, J. (2017). Population dynamics and antimicrobial resistance of the most prevalent poultry-associated *Salmonella* serotypes. Poultry Science, 96(3), 687–702. <https://doi.org/10.3382/ps/pew342>

Sharma, M., Handy, E. T., East, C. L., Kim, S., Jiang, C., Callahan, M. T., Allard, S. M., Micallef, S., Craighead, S., Anderson-Coughlin, B., Gartley, S., Vanore, A., Kniel, K. E., Haymaker, J., Duncan, R., Foust, D., White, C., Taabodi, M., Hashem, F., ... Sapkota, A. R. (2020). Prevalence of *Salmonella* and *Listeria monocytogenes* in non-traditional irrigation waters in the Mid-Atlantic United States is affected by water type, season, and recovery method. PLOS ONE, 15(3), e0229365. <https://doi.org/10.1371/journal.pone.0229365>

- Timme, R. E., Pettengill, J. B., Allard, M. W., Strain, E., Barrangou, R., Wehnes, C., Van Kessel, J. S., Karns, J. S., Musser, S. M., & Brown, E. W. (2013). Phylogenetic Diversity of the Enteric Pathogen *Salmonella enterica* subsp. *enterica* Inferred from Genome-Wide Reference-Free SNP Characters. *Genome Biology and Evolution*, 5(11), 2109–2123. <https://doi.org/10.1093/gbe/evt159>
- Topalcengiz, Z., McEgan, R., & Danyluk, M. D. (2019). Fate of *Salmonella* in Central Florida Surface Waters and Evaluation of EPA Worst Case Water as a Standard Medium. *Journal of Food Protection*, 82(6), 916–925. <https://doi.org/10.4315/0362-028X.JFP-18-331>
- Toro, M., Weller, D., Ramos, R., Diaz, L., Alvarez, F. P., Reyes-Jara, A., Moreno-Switt, A. I., Meng, J., & Adell, A. D. (2022). Environmental and anthropogenic factors associated with the likelihood of detecting *Salmonella* in agricultural watersheds. *Environmental Pollution*, 306, 119298. <https://doi.org/10.1016/j.envpol.2022.119298>
- Uzzau, S., Brown, D. J., Wallis, T., Rubino, S., Leori, G., Bernard, S., Casadesus, J., Platt, D. J., & Olsen, J. E. (2000). Host adapted serotypes of *Salmonella enterica*. *Epidemiology and Infection*, 125(2), 229–255. <https://doi.org/10.1017/S0950268899004379>
- Vinueza, D., Ochoa-Herrera, V., Maurice, L., Tamayo, E., Mejía, L., Tejera, E., & Machado, A. (2021). Determining the microbial and chemical contamination in Ecuador's main rivers. *Scientific Reports*, 11(1), 17640. <https://doi.org/10.1038/s41598-021-96926-z>

- Vinueza-Burgos, C., Cevallos, M., Ron-Garrido, L., Bertrand, S., & De Zutter, L. (2016). Prevalence and Diversity of *Salmonella* Serotypes in Ecuadorian Broilers at Slaughter Age. PLOS ONE, 11(7), e0159567. <https://doi.org/10.1371/journal.pone.0159567>
- Vinueza-Burgos, C., Hidalgo-Arellano, L., Gómez-Coronado, C., Medina-Santana, J. L., & Cevallos-Almeida, M. (2024). Prevalence, serovars, and risk factors associated with the presence of *Salmonella* in pork sold in public markets in Quito, Ecuador. F1000Research, 12, 1367. <https://doi.org/10.12688/f1000research.138671.3>
- Vizcaíno, I., Carrera, E., Sanromán-Junquera, M., Muñoz-Romero, S., Luis Rojo-Álvarez, J., & Cumbal, L. (2016). Spatio-Temporal Analysis of Water Quality Parameters in Machángara River with Nonuniform Interpolation Methods. Water, 8(11), 507. <https://doi.org/10.3390/w8110507>
- Yoshida, C. E., Kruczkiewicz, P., Laing, C. R., Lingohr, E. J., Gannon, V. P. J., Nash, J. H. E., & Taboada, E. N. (2016). The *Salmonella* In Silico Typing Resource (SISTR): An Open Web-Accessible Tool for Rapidly Typing and Subtyping Draft *Salmonella* Genome Assemblies. PLOS ONE, 11(1), e0147101. <https://doi.org/10.1371/journal.pone.0147101>
- You, Y., Rankin, S. C., Aceto, H. W., Benson, C. E., Toth, J. D., & Dou, Z. (2006). Survival of *Salmonella enterica* Serovar Newport in Manure and Manure-Amended Soils. Applied and Environmental Microbiology, 72(9), 5777–5783. <https://doi.org/10.1128/AEM.00791-06>
- Zhang, S., den Bakker, H. C., Li, S., Chen, J., Dinsmore, B. A., Lane, C., Lauer, A. C., Fields, P. I., & Deng, X. (2019). SeqSero2: Rapid and Improved *Salmonella*

Serotype Determination Using Whole-Genome Sequencing Data. *Applied and Environmental Microbiology*, 85(23). <https://doi.org/10.1128/AEM.01746-19>

Zhou, Z., Alikhan, N.-F., Mohamed, K., Fan, Y., & Achtman, M. (2020). The EnteroBase user's guide, with case studies on *Salmonella* transmissions, *Yersinia pestis* phylogeny, and *Escherichia* core genomic diversity. *Genome Research*, 30(1), 138–152. <https://doi.org/10.1101/gr.251678.119>

Zhou, Z., Charlesworth, J., & Achtman, M. (2021). HierCC: a multi-level clustering scheme for population assignments based on core genome MLST. *Bioinformatics*, 37(20), 3645–3646. <https://doi.org/10.1093/bioinformatics/btab234>

CHAPTER 5

CONCLUSIONS

5.1 Conclusions

Salmonella enterica is a significant human pathogen, causing an estimated 176 million global human illnesses and 290,000 deaths annually (CDC, 2024). In the United States, it is a leading bacterial foodborne pathogen, contributing to 1.35 million cases each year. In addition to being commonly identified in food animals, this pathogen is ubiquitous in the environment, including wildlife and water. Illness often occurs following consumption of contaminated food, including meat and poultry, which contribute to 43.7% of estimated illnesses from foodborne outbreaks (IFSAC, 2024). This is expected, as sources such as poultry, cattle, and swine can carry *Salmonella* in their gastrointestinal tract as part of their natural microbiomes. Surprisingly, produce accounts for 43.4% of estimated human illnesses, suggesting that transmission from contaminated animal feces occurs during the growing, harvesting, or packing stages of production. As the preparation of produce often does not include cooking, pathogenic contaminants are more likely to remain viable, leading to a greater potential for human illness (Devarajan et al., 2023). Because of this, it is critical to prevent initial contamination by reducing the risk of transmission from various environmental routes. Relative to meat and poultry, these vectors are significantly understudied. The work presented in this dissertation advances our understanding of *Salmonella* ecology within an animal reservoir (wild birds) and in the environment (surface water) by evaluating risk factors that influence

Salmonella viability, identifying complex serovar populations, and establishing foundations of environmental monitoring in these environments.

Wild birds present a unique food safety challenge as traditional wildlife mitigation, such as fences and netting, do not prevent intrusion of this potential contamination source (Rivadeneira et al., 2018). Previous work has identified *Salmonella* and other foodborne pathogens in wild birds near crop fields, suggesting a potential role in transmission to produce (Gorski et al., 2011; Kirk et al., 2002; Navarro-Gonzalez et al., 2020; Rivadeneira et al., 2016; O. M. Smith, Edworthy, et al., 2020; O. M. Smith et al., 2022). In the second chapter of this dissertation, the role of wild birds in *Salmonella* transmission to produce fields in the Southeastern United States is detailed. Prior work investigated this role on the West Coast, where leafy greens such as lettuce, kale, and broccoli are the predominant crop. Alternatively, produce in the Southeast is represented by peppers, tomatoes, and cucumbers. In the Southeast study presented in this dissertation, wild bird feces were sampled from produce fields and cultured for the presence of *Salmonella*. Prevalence was low in this sample type, identifying viable *Salmonella* in 2.1% (16/773) of samples. This suggests that wild birds likely have a limited role in pathogen transmission in this region. *Salmonella* was significantly more likely to be recovered from fresh feces, indicating that this matrix is particularly susceptible to desiccation. This is a contrast from studies showing long-term recovery of *Salmonella* in composited bird feces up to 291 days (Oni et al., 2015; Topalcengiz et al., 2020). The results of the model were further supported by GLMM models from weather data that identified decreased precipitation as a significant factor in reduced *Salmonella* prevalence, suggesting that weather may play a critical role in risk assessment and

pathogen transmission. Molecular detection of *Salmonella* by PCR of the *invA* gene identified 59 additional samples containing non-culturable *Salmonella* signatures. This demonstrates the relevance of culture-based surveillance to avoid overestimating risk attributed to wild birds, and is a current dispute in food safety (O. M. Smith, Snyder, et al., 2020). Additionally, *Salmonella* serovar populations were assessed in culture-positive samples and the majority of these samples were found to contain multiple serovars. Results from this data showed that while the presence of viable *Salmonella* is low, these populations can include clinically relevant serovars such as Enteritidis, Typhimurium, Newport, Saintpaul, Infantis, and Muenchen. Finally, this work aimed to identify wild bird species associated with higher transmission risk. To accomplish this, a combination approach of in-field point counts and molecular identification via PCR of the gene encoding the cytochrome C oxidase subunit I identified 57 bird species, including 10 species from feces containing *Salmonella*. Species associated with on-farm structures and direct crop interaction were assigned the highest risk category due to their increased likelihood of transmitting *Salmonella* to produce. This suggests that targeted mitigation towards these species would be the most effective and efficient in reducing food safety risks from wild birds. While weather factors vary between regions, bird species associated with large monoculture farms, along with those found in West Coast studies, were similar to species identified here (Flohre et al., 2011; Gonthier et al., 2019; Olimp et al., 2024; O. M. Smith, Edworthy, et al., 2020). This congruence of commonly identified bird species suggests that improved risk management could have a widespread impact on food safety.

Conclusions from this study showed that *Salmonella* transmission through defecation in produce fields is low, suggesting that risk management strategies should be targeted towards one or more alternative sources. Limitations of this study included species identification, as point counts identified species interacting in and around crop fields but not necessarily those contributing to fecal contamination on the produce. Molecular species identification demonstrated a direct assessment of species defecating in the fields; however, this method failed in a majority of samples, likely due to DNA degradation of the target sequence (Hebert et al., 2003; IVANOVA et al., 2007; Joo & Park, 2012; KERR et al., 2007). This method did provide an advantage over studies that trap and collect samples from live birds: while those studies were able to identify species, it was not necessarily indicative of the birds defecating on the produce. Future work on this topic is needed to better categorize *Salmonella* prevalence and multiseroovar populations in wild birds. By continuing to collect both sporadic fecal samples and direct collections from wild birds, the development of a robust *Salmonella* database from this source would improve risk assessment. This would facilitate more thorough source attribution investigations if wild birds are involved in future outbreaks. Additionally, further analysis is needed to evaluate human activities associated with increased *Salmonella* in wild birds, such as urbanization and agricultural intensification. While these activities have been assessed in a few studies (Hernandez et al., 2016; O. M. Smith, Edworthy, et al., 2020), regional differences may contribute to variable influence by wild birds, suggesting a need for more widespread research in this field.

Surface water has been regularly identified as a source of *Salmonella* around the world (Casanova et al., 2020; Chen et al., 2024; Gorski et al., 2022; Haley et al., 2009;

Jokinen et al., 2015; Mahagamage et al., 2020; Micallef et al., 2012; Ruiz et al., 1987; Somda et al., 2021). Additionally, a wide range of serovars and incidence of AMR have been reported, demonstrating a need to further evaluate this *Salmonella* reservoir as a food safety concern (Berge et al., 2006; Cho et al., 2020; Li et al., 2014; Micallef et al., 2012). Chapter three introduces a study that investigated *Salmonella* prevalence, serovar complexity, and antibiotic resistance in four distinct creeks in the Southeastern United States. Prevalence was high in each creek, ranging from 61% to 78%, where the lowest *Salmonella* prevalence was identified in the urban creek and the highest in a rural creek. Seasonality significantly influenced *Salmonella* prevalence, for example, the lowest prevalence was found in summer months, in contrast with prior work (Deaven et al., 2021; Haley et al., 2009). Deep serotyping identified multiserovar populations in almost 90% of positive samples; a proportion higher than that of previously studied subjects, including poultry, cattle, and wild birds (Cason et al., 2024; Obe et al., 2023; Siceloff et al., 2021, 2022; J. C. Smith et al., 2023). These populations contained high complexity, averaging almost four serovars per sample and ranging up to 13 serovars in a single sample. As this complexity has not regularly been identified from other sources, it is likely that surface water can act as a collection point for serovars being shed by multiple sources in the environment. Importantly, many clinically relevant serovars were identified, demonstrating the importance of water surveillance to identify previously understudied routes of *Salmonella* transmission. Land attribution characteristics and weather variables were modeled to identify predictive factors for *Salmonella* risk, finding that increased precipitation and humidity were significantly associated with higher prevalence and complexity. This agreed with prior studies (Deaven et al., 2021; Strawn et

al., 2013; Weller et al., 2020); however, proximal land use was less significantly associated with any *Salmonella* trends. This was surprising, considering these creeks were each enriched for a unique land use, including a national forest, a suburban population, and animal agriculture. Limited access by humans and the absence of animal agriculture would likely prevent the introduction of antibiotic resistance to this creek; however, results suggest a natural transmission of AMR genes. This finding is supported by other studies where AMR transmission was identified in natural environments (Cho et al., 2022; Hwengwere et al., 2022).

Overall, the results of this study demonstrate a high prevalence and complexity of pathogenic serovars in various surface water sources. The dissimilarity between serovars found in this source and those recovered from food animals underscores the need for a One Health approach for *Salmonella* surveillance to identify food safety risks and implement effective management strategies. Further surveillance of environmental and wildlife sources is needed to better understand the transmission of *Salmonella* between the agricultural and environmental interface. Future work is needed to identify *Salmonella* dynamics that influence survival in aquatic environments, such as adaptations to overcome limited nutrient availability or resistance to osmotic stress. These genomic adaptations that promote viability in water may allow a strain to outcompete others in the source, leading to its expansion within the niche.

As previously identified, a variety of contamination sources threaten surface water, making these sources of freshwater unsafe for human use. While many countries around the world have established water sanitization practices, the United Nations estimates that 2.2 billion people in the world do not have access to clean water (Burden

of Disease Attributable to Unsafe Drinking Water, Sanitation and Hygiene, n.d.). This results in the use of unsafe water, which leads to higher rates of typhoid fever and other diseases. To differentiate characteristics that influence water safety in agricultural and urban rivers, chapter four discusses a study that measured *Salmonella* prevalence and serovar diversity of two rivers in Pichincha province, Ecuador. In this investigation, three-quarters of samples were positive for *Salmonella*, where the urban river was found to have higher prevalence and complexity compared to the agricultural river. Whole genome sequencing of 54 isolates identified eight serovars, where nearly half of positive samples contained two or more serovars. Importantly, analysis of the core genome between isolates of the same serovar identified characteristics of transmission in multiple strains, including serovar Typhimurium and its monophasic variant, I 4,[5],12:i:-. The presence of highly related isolates consistently found in the same site for multiple days also suggests potential reintroduction events or persistence within these rivers. This suggests that persistent strains may be maintained for extended periods, contributing to increased safety concerns for these rivers.

Data from this study demonstrated, for the first time, high *Salmonella* serovar prevalence that was also reflected in many serovars being isolated in two rivers in Ecuador. This reinforces the need to protect public water sources and the need for regular surveillance of surface water for public health risks. As a result of previous studies that identified unsafe levels of chemical and microbial water quality in rivers including the Río Machángara (Borja-Serrano et al., 2020; Ortega-Paredes et al., 2020; Vinueza et al., 2021), an Ecuadorian court ruled that protection of this urban river is a legal requirement. This is an encouraging first step for making publicly accessible water sources safe. Future

work is needed to assess the efficacy of water treatment to reduce the prevalence and complexity of *Salmonella* populations in contaminated rivers. To further support this aim, more thorough surveillance of rivers, including upstream sites and points of convergence, would provide a system of tracking contamination routes in these rivers.

Regulations in the United States, including FSMA and GAPs, seek to reduce food safety concerns from various routes of transmission. To do so, risk assessment strategies for irrigation water, soil amendments, and worker hygiene are clearly described; however, wildlife management is uniquely non-specific. This is, in part, due to the variety of animal species that may contribute to risk, making an all-encompassing mitigation approach difficult (Gordus, 2011; Langholz & Jay-Russell, 2013). Wild bird exclusion strategies are unrealistic compared to other wildlife, as fences and other barriers do little to prevent avian movement into fields (Rivadeneira et al., 2018; Varriano et al., 2025). As a result, many research groups have characterized *Salmonella* prevalence within wild birds to assess food safety risks (Kirk et al., 2002; Maurer et al., 2015a; Navarro-Gonzalez et al., 2020; Olimpi et al., 2024; O. M. Smith, Edworthy, et al., 2020; O. M. Smith, Olimpi, Navarro-Gonzalez, et al., 2022). Research in this dissertation further characterized prevalence; however, deep serotyping set this work apart by identifying multiserovar populations within wild birds. This demonstrated a potential limitation in previous studies, where serovars may not have been detected. While the incidence of *Salmonella* was low in this sample type, the presence of clinically relevant serovars within positive samples highlights the need to identify the entire population rather than only the most abundant serovar. This research found an association between the freshness of feces and viable *Salmonella*, where the pathogen was significantly less

likely to be cultured from dried feces. Additionally, no incidence of movement from the feces to produce on the plant or to neighboring plants was identified. These findings suggest that the risk associated with wild bird feces in produce fields is low, and this risk is further reduced as feces become desiccated. The results of these findings could be used to support arguments for limiting no-harvest zones following fecal contamination on produce and reducing pressure to interfere with wild bird movement (Hamilton et al., 2015; Olimpi et al., 2024). Studies testing *Salmonella* viability and transmission in controlled environments will be needed to further support these findings.

Surface water is a critical component in food systems for the irrigation of produce and as collection points for runoff. As such, *Salmonella* prevalence has been previously characterized in many studies, both within the United States and internationally (Chen et al., 2024; Haley et al., 2009; Mahagamage et al., 2020; McEgan et al., 2014; Micallef et al., 2012; Murphy et al., 2022; Somda et al., 2021; Vadde et al., 2019). This dissertation highlights two surface water studies: one in the Southeastern United States and one in the Pichincha region of Ecuador. The use of deep serotyping technology advanced the understanding of *Salmonella* populations in the United States, where only one prior study had assessed these complex populations (Deaven et al., 2021). This study resolved a limitation of the previous investigation, where smaller creeks were sampled instead of a large river. This difference allowed for a more comprehensive analysis of the entire *Salmonella* population in the water source. Additionally, monthly sampling over two years allowed for more robust modeling of factors that influence *Salmonella* prevalence and complexity. Modeling suggested that season and weather conditions, including precipitation and humidity, influenced *Salmonella* more than neighboring land

attribution. This finding may be helpful for establishing surface water management efforts, where these risk factors can be implemented to determine when surface water is likely to be unsafe for agricultural use. The study in Ecuador isolated multiple serotypes in single samples and analyzed these using whole genome sequencing for the first time. While rivers in this region have been evaluated for water quality in previous work (Borja-Serrano et al., 2020; Ortega-Paredes et al., 2020; Vinueza et al., 2021), the characterization of *Salmonella* was unclear, and potentially underestimated. Genomic analysis suggested potential downstream transmission of highly related isolates; however, further studies are needed to evaluate both multiserovar populations and transmission of *Salmonella* in surface water, as these features likely play a critical role food safety.

The body of work presented in this dissertation addresses significant knowledge gaps in human health and food safety. These studies identified environmental sources of complex *Salmonella* populations in wild birds and surface water containing clinically relevant serovars. Additionally, factors contributing to food safety risks in these reservoirs were identified, which can be utilized in farm-specific risk management strategies, including species-specific mitigation of wild birds and the identification of weather conditions that increase the likelihood of irrigation water contamination. Finally, this research establishes a framework for a One Health approach to *Salmonella* surveillance, which can be utilized to reduce the incidence of *Salmonella* in food safety.

5.2 References

- Berge, A. C. B., Dueger, E. L., & Sischo, W. M. (2006). Comparison of *Salmonella enterica* serovar distribution and antibiotic resistance patterns in wastewater at municipal water treatment plants in two California cities. *Journal of Applied Microbiology*, 101(6), 1309–1316. <https://doi.org/10.1111/j.1365-2672.2006.03031.x>
- Borja-Serrano, P., Ochoa-Herrera, V., Maurice, L., Morales, G., Quilumbaqui, C., Tejera, E., & Machado, A. (2020). Determination of the Microbial and Chemical Loads in Rivers from the Quito Capital Province of Ecuador (Pichincha)—A Preliminary Analysis of Microbial and Chemical Quality of the Main Rivers. *International Journal of Environmental Research and Public Health*, 17(14), 5048. <https://doi.org/10.3390/ijerph17145048>
- Burden of disease attributable to unsafe drinking-water, sanitation and hygiene. (2023). (Accessed January 8, 2025). <https://www.who.int/publications/i/item/9789240075610>
- Casanova, L. M., Hill, V. R., & Sobsey, M. D. (2020). Antibiotic-resistant *Salmonella* in swine wastes and farm surface waters. *Letters in Applied Microbiology*, 71(1), 117–123. <https://doi.org/10.1111/lam.13242>
- Cason, E. E., Carlson, A. V., Siemens, A. L., & Shariat, N. W. (2024). High-resolution Serotyping Reveals *Salmonella* Surveillance Challenges in the Turkey Industry. *Journal of Food Protection*, 87(10), 100319. <https://doi.org/10.1016/j.jfp.2024.100319>

CDC. (2024). CDC Yellow Book: Health Information for International Travel. Oxford University Press. (Accessed January 9, 2025).

<https://wwwnc.cdc.gov/travel/page/yellowbook-home>

Chen, Z., Moreno-Switt, A. I., Reyes-Jara, A., Delgado Suarez, E., Adell, A. D., Oliveira, C. J. B., Bonelli, R. R., Huang, X., Brown, E., Allard, M., Grim, C., Bell, R., Meng, J., & Toro, M. (2024). A multicenter genomic epidemiological investigation in Brazil, Chile, and Mexico reveals the diversity and persistence of *Salmonella* populations in surface waters. *MBio*, 15(7), e0077724.

<https://doi.org/10.1128/mbio.00777-24>

Cho, S., Hiott, L. M., House, S. L., Woodley, T. A., McMillan, E. A., Sharma, P., Barrett, J. B., Adams, E. S., Brandenburg, J. M., Hise, K. B., Bateman McDonald, J. M., Ottesen, E. A., Lipp, E. K., Jackson, C. R., & Frye, J. G. (2022). Analysis of *Salmonella enterica* Isolated from a Mixed-Use Watershed in Georgia, USA: Antimicrobial Resistance, Serotype Diversity, and Genetic Relatedness to Human Isolates. *Applied and Environmental Microbiology*, 88(10).

<https://doi.org/10.1128/aem.00393-22>

Cho, S., Jackson, C. R., & Frye, J. G. (2020). The prevalence and antimicrobial resistance phenotypes of *Salmonella*, *Escherichia coli* and *Enterococcus* sp. in surface water. *Letters in Applied Microbiology*, 71(1), 3–25.

<https://doi.org/10.1111/lam.13301>

Deaven, A. M., Ferreira, C. M., Reed, E. A., Chen See, J. R., Lee, N. A., Almaraz, E., Rios, P. C., Marogi, J. G., Lamendella, R., Zheng, J., Bell, R. L., & Shariat, N. W. (2021). *Salmonella* Genomics and Population Analyses Reveal High Inter- and

- Intraseroval Diversity in Freshwater. *Applied and Environmental Microbiology*, 87(6). <https://doi.org/10.1128/AEM.02594-20>
- Devarajan, N., Weller, D. L., Jones, M., Adell, A. D., Adhikari, A., Allende, A., Arnold, N. L., Baur, P., Beno, S. M., Clements, D., Olimpi, E. M., Critzer, F., Green, H., Gorski, L., Ferelli Gruber, A., Kovac, J., McGarvey, J., Murphy, C. M., Murphy, S. I., ... Karp, D. S. (2023). Evidence for the efficacy of pre-harvest agricultural practices in mitigating food-safety risks to fresh produce in North America. *Frontiers in Sustainable Food Systems*, 7, 1101435. <https://doi.org/10.3389/fsufs.2023.1101435>
- Flohre, A., Fischer, C., Aavik, T., Bengtsson, J., Berendse, F., Bommarco, R., Ceryngier, P., Clement, L. W., Dennis, C., Eggers, S., Emmerson, M., Geiger, F., Guerrero, I., Hawro, V., Inchausti, P., Liira, J., Morales, M. B., Oñate, J. J., Pärt, T., ... Tschamtkke, T. (2011). Agricultural intensification and biodiversity partitioning in European landscapes comparing plants, carabids, and birds. *Ecological Applications*, 21(5), 1772–1781. <https://doi.org/10.1890/10-0645.1>
- Gonthier, D. J., Sciligo, A. R., Karp, D. S., Lu, A., Garcia, K., Juarez, G., Chiba, T., Gennet, S., & Kremen, C. (2019). Bird services and disservices to strawberry farming in Californian agricultural landscapes. *Journal of Applied Ecology*, 56(8), 1948–1959. <https://doi.org/10.1111/1365-2664.13422>
- Gordus, A. G. , R. E. M. and E. R. A. (2011). Wildlife Survey for *E. coli* O157:H7 and *Salmonella* in the Central Coastal Counties of California.
- Gorski, L., Liang, A. S., Walker, S., Carychao, D., Aviles Noriega, A., Mandrell, R. E., & Cooley, M. B. (2022). *Salmonella enterica* Serovar Diversity, Distribution, and

- Prevalence in Public-Access Waters from a Central California Coastal Leafy Green-Growing Region from 2011 to 2016. *Applied and Environmental Microbiology*, 88(3). <https://doi.org/10.1128/aem.01834-21>
- Gorski, L., Parker, C. T., Liang, A., Cooley, M. B., Jay-Russell, M. T., Gordus, A. G., Atwill, E. R., & Mandrell, R. E. (2011). Prevalence, Distribution, and Diversity of *Salmonella enterica* in a Major Produce Region of California. *Applied and Environmental Microbiology*, 77(8), 2734–2748. <https://doi.org/10.1128/AEM.02321-10>
- Haley, B. J., Cole, D. J., & Lipp, E. K. (2009). Distribution, Diversity, and Seasonality of Waterborne Salmonellae in a Rural Watershed. *Applied and Environmental Microbiology*, 75(5), 1248–1255. <https://doi.org/10.1128/AEM.01648-08>
- Hamilton, K. E., Umber, J., Hultberg, A., Tong, C., Schermann, M., Diez-Gonzalez, F., & Bender, J. B. (2015). Validation of Good Agricultural Practices (GAP) on Minnesota Vegetable Farms. *Foodborne Pathogens and Disease*, 12(2), 145–150. <https://doi.org/10.1089/fpd.2014.1817>
- Hebert, P. D. N., Cywinska, A., Ball, S. L., & deWaard, J. R. (2003). Biological identifications through DNA barcodes. *Proceedings of the Royal Society of London. Series B: Biological Sciences*, 270(1512), 313–321. <https://doi.org/10.1098/rspb.2002.2218>
- Hernandez, S. M., Welch, C. N., Peters, V. E., Lipp, E. K., Curry, S., Yabsley, M. J., Sanchez, S., Presotto, A., Gerner-Smidt, P., Hise, K. B., Hammond, E., Kistler, W. M., Madden, M., Conway, A. L., Kwan, T., & Maurer, J. J. (2016). Urbanized White Ibises (*Eudocimus albus*) as Carriers of *Salmonella enterica* of

- Significance to Public Health and Wildlife. PLOS ONE, 11(10), e0164402.
<https://doi.org/10.1371/journal.pone.0164402>
- Hwengwere, K., Paramel Nair, H., Hughes, K. A., Peck, L. S., Clark, M. S., & Walker, C. A. (2022). Antimicrobial resistance in Antarctica: is it still a pristine environment? *Microbiome*, 10(1), 71. <https://doi.org/10.1186/s40168-022-01250-x>
- IFSAC. (2024). Foodborne Illness Source Attribution Estimates – United States, 2022. Accessed January 8, 2025). <https://www.cdc.gov/ifsac/php/data-research/annual-report-2022.html>.
- Ivanova, N. V., Zemlak, T. S., Hanner, R. H., & Hebert, P. D. N. (2007). Universal primer cocktails for fish DNA barcoding. *Molecular Ecology Notes*, 7(4), 544–548. <https://doi.org/10.1111/j.1471-8286.2007.01748.x>
- Jokinen, C. C., Koot, J., Cole, L., Desruisseau, A., Edge, T. A., Khan, I. U. H., Koning, W., Lapen, D. R., Pintar, K. D. M., Reid-Smith, R., Thomas, J. L., Topp, E., Wang, L. Y., Wilkes, G., Ziebell, K., van Bochove, E., & Gannon, V. P. J. (2015). The distribution of *Salmonella enterica* serovars and subtypes in surface water from five agricultural regions across Canada. *Water Research*, 76, 120–131. <https://doi.org/10.1016/j.watres.2015.02.038>
- Joo, S., & Park, S. (2012). Identification of bird species and their prey using DNA barcode on feces from Korean traditional village groves and forests (maeulsoop). *Animal Cells and Systems*, 16(6), 488–497. <https://doi.org/10.1080/19768354.2012.720939>

- Kerr, K. C. R., Stoeckle, M. Y., Dove, C. J., Weigt, L. A., Francis, C. M., & Hebert, P. D. N. (2007). Comprehensive DNA barcode coverage of North American birds. *Molecular Ecology Notes*, 7(4), 535–543. <https://doi.org/10.1111/j.1471-8286.2007.01670.x>
- Kirk, J. H., Holmberg, C. A., & Jeffrey, J. S. (2002). Prevalence of *Salmonella* spp in selected birds captured on California dairies. *Journal of the American Veterinary Medical Association*, 220(3), 359–362. <https://doi.org/10.2460/javma.2002.220.359>
- Langholz, J., & Jay-Russell, M. (2013). Potential role of wildlife in pathogenic contamination of fresh produce. *Human-Wildlife Interactions*, 7(1), 140–157.
- Li, B., Vellidis, G., Liu, H., Jay-Russell, M., Zhao, S., Hu, Z., Wright, A., & Elkins, C. A. (2014). Diversity and Antimicrobial Resistance of *Salmonella enterica* Isolates from Surface Water in Southeastern United States. *Applied and Environmental Microbiology*, 80(20), 6355–6365. <https://doi.org/10.1128/AEM.02063-14>
- Mahagamage, M. G. Y. L., Pathirage, M. V. S. C., & Manage, P. M. (2020). Contamination Status of *Salmonella* spp., *Shigella* spp. and *Campylobacter* spp. in Surface and Groundwater of the Kelani River Basin, Sri Lanka. *Water*, 12(8), 2187. <https://doi.org/10.3390/w12082187>
- Maurer, J. J., Martin, G., Hernandez, S., Cheng, Y., Gerner-Smidt, P., Hise, K. B., Tobin D'Angelo, M., Cole, D., Sanchez, S., Madden, M., Valeika, S., Presotto, A., & Lipp, E. K. (2015). Diversity and Persistence of *Salmonella enterica* Strains in Rural Landscapes in the Southeastern United States. *PLOS ONE*, 10(7), e0128937. <https://doi.org/10.1371/journal.pone.0128937>

- McEgan, R., Chandler, J. C., Goodridge, L. D., & Danyluk, M. D. (2014). Diversity of *Salmonella* Isolates from Central Florida Surface Waters. *Applied and Environmental Microbiology*, 80(21), 6819–6827.
<https://doi.org/10.1128/AEM.02191-14>
- Micallef, S. A., Rosenberg Goldstein, R. E., George, A., Kleinfelter, L., Boyer, M. S., McLaughlin, C. R., Estrin, A., Ewing, L., Jean-Gilles Beaubrun, J., Hanes, D. E., Kothary, M. H., Tall, B. D., Razeq, J. H., Joseph, S. W., & Sapkota, A. R. (2012). Occurrence and antibiotic resistance of multiple *Salmonella* serotypes recovered from water, sediment and soil on mid-Atlantic tomato farms. *Environmental Research*, 114, 31–39. <https://doi.org/10.1016/j.envres.2012.02.005>
- Murphy, C. M., Strawn, L. K., Chapin, T. K., McEgan, R., Gopidi, S., Friedrich, L., Goodridge, L. D., Weller, D. L., Schneider, K. R., & Danyluk, M. D. (2022). Factors Associated With *E. coli* Levels in and *Salmonella* Contamination of Agricultural Water Differed Between North and South Florida Waterways. *Frontiers in Water*, 3. <https://doi.org/10.3389/frwa.2021.750673>
- Navarro-Gonzalez, N., Wright, S., Aminabadi, P., Gwinn, A., Suslow, T. V., & Jay-Russell, M. T. (2020). Carriage and Subtypes of Foodborne Pathogens Identified in Wild Birds Residing near Agricultural Lands in California: a Repeated Cross-Sectional Study. *Applied and Environmental Microbiology*, 86(3).
<https://doi.org/10.1128/AEM.01678-19>
- Obe, T., Siceloff, A. T., Crowe, M. G., Scott, H. M., & Shariat, N. W. (2023). Combined Quantification and Deep Serotyping for *Salmonella* Risk Profiling in Broiler

- Flocks. *Applied and Environmental Microbiology*, 89(4).
<https://doi.org/10.1128/aem.02035-22>
- Olimpi, E. M., Ke, A., Baur, P., Carlisle, L., Esquivel, K. E., Glaser, T., Snyder, W. E., Waterhouse, H., Bowles, T. M., Kremen, C., & Karp, D. S. (2024). Ungrazed seminatural habitats around farms benefit bird conservation without enhancing foodborne pathogen risks. *Landscape Ecology*, 39(7), 128.
<https://doi.org/10.1007/s10980-024-01907-y>
- Oni, R. A., Sharma, M., & Buchanan, R. L. (2015). Survival of *Salmonella enterica* in Dried Turkey Manure and Persistence on Spinach Leaves. *Journal of Food Protection*, 78(10), 1791–1799. <https://doi.org/10.4315/0362-028X.JFP-15-047>
- Ortega-Paredes, D., Barba, P., Mena-López, S., Espinel, N., Crespo, V., & Zurita, J. (2020). High quantities of multidrug-resistant *Escherichia coli* are present in the Machángara urban river in Quito, Ecuador. *Journal of Water and Health*, 18(1), 67–76. <https://doi.org/10.2166/wh.2019.195>
- Rivadeneira, P., Hilson, C., Justice-Allen, A., & Jay-Russell, M. (2016). Pathogen Risks Related to the Movement of Birds Frequenting Livestock and Fresh Produce Growing Areas in the Southwestern U.S. *Proceedings of the Vertebrate Pest Conference*, 27, 258–263. <https://doi.org/10.5070/V427110583>
- Rivadeneira, P., Kross, S., Navarro-Gonzalez, N., & Jay-Russell, M. (2018). A Review of Bird Deterrents Used in Agriculture. *Proceedings of the Vertebrate Pest Conference*, 28, 218–223. <https://doi.org/10.5070/V42811040>
- Ruiz, B. G.-V., Espinar, A. C., & Carmona, M. J. B. (1987). A comparative study of strains of *Salmonella* isolated from irrigation waters, vegetables and human

- infections. *Epidemiology and Infection*, 98(3), 271–276.
<https://doi.org/10.1017/S0950268800062026>
- Siceloff, A. T., Ohta, N., Norman, K. N., Loneragan, G. H., Norby, B., Scott, H. M., & Shariat, N. W. (2021). Antimicrobial Resistance Hidden within Multiserovar *Salmonella* Populations. *Antimicrobial Agents and Chemotherapy*, 65(6), e00048-21. <https://doi.org/10.1128/AAC.00048-21>
- Siceloff, A. T., Waltman, D., & Shariat, N. W. (2022). Regional *Salmonella* Differences in United States Broiler Production from 2016 to 2020 and the Contribution of Multiserovar Populations to *Salmonella* Surveillance. *Applied and Environmental Microbiology*, 88(8). <https://doi.org/10.1128/aem.00204-22>
- Smith, J. C., Varriano, S., Roach, K., Snipes, Z., Dawson, J. L., Shealy, J., Dunn, L. L., Snyder, W. E., & Shariat, N. W. (2023). Prevalence and molecular characterization of *Salmonella* isolated from wild birds in fresh produce environments. *Frontiers in Microbiology*, 14.
<https://doi.org/10.3389/fmicb.2023.1272916>
- Smith, O. M., Edworthy, A., Taylor, J. M., Jones, M. S., Tormanen, A., Kennedy, C. M., Fu, Z., Latimer, C. E., Cornell, K. A., Michelotti, L. A., Sato, C., Northfield, T., Snyder, W. E., & Owen, J. P. (2020). Agricultural intensification heightens food safety risks posed by wild birds. *Journal of Applied Ecology*, 57(11).
<https://doi.org/10.1111/1365-2664.13723>
- Smith, O. M., Olimpi, E. M., Navarro-Gonzalez, N., Cornell, K. A., Frishkoff, L. O., Northfield, T. D., Bowles, T. M., Edworthy, M., Eilers, J., Fu, Z., Garcia, K., Gonthier, D. J., Jones, M. S., Kennedy, C. M., Latimer, C. E., Owen, J. P., Sato,

- C., Taylor, J. M., Wilson-Rankin, E. E., ... Karp, D. S. (2022). A trait-based framework for predicting foodborne pathogen risk from wild birds. *Ecological Applications*, 32(2). <https://doi.org/10.1002/eap.2523>
- Smith, O. M., Snyder, W. E., & Owen, J. P. (2020). Are we overestimating risk of enteric pathogen spillover from wild birds to humans? *Biological Reviews*, 95(3), 652–679. <https://doi.org/10.1111/brv.12581>
- Somda, N. S., Bonkougou, I. J. O., Sambe-Ba, B., Drabo, M. S., Wane, A. A., Sawadogo-Lingani, H., & Savadogo, A. (2021). Diversity and antimicrobial drug resistance of non-typhoid *Salmonella* serotypes isolated in lettuce, irrigation water and clinical samples in Burkina Faso. *Journal of Agriculture and Food Research*, 5, 100167. <https://doi.org/10.1016/j.jafr.2021.100167>
- Strawn, L. K., Gröhn, Y. T., Warchocki, S., Worobo, R. W., Bihn, E. A., & Wiedmann, M. (2013). Risk Factors Associated with *Salmonella* and *Listeria monocytogenes* Contamination of Produce Fields. *Applied and Environmental Microbiology*, 79(24), 7618–7627. <https://doi.org/10.1128/AEM.02831-13>
- Topalcengiz, Z., Spanninger, P. M., Jeamsripong, S., Persad, A. K., Buchanan, R. L., Saha, J., LeJeune, J., Jay-Russell, M. T., Kniel, K. E., & Danyluk, M. D. (2020). Survival of *Salmonella* in Various Wild Animal Feces That May Contaminate Produce. *Journal of Food Protection*, 83(4), 651–660. <https://doi.org/10.4315/0362-028X.JFP-19-302>
- Vadde, K. K., McCarthy, A. J., Rong, R., & Sekar, R. (2019). Quantification of Microbial Source Tracking and Pathogenic Bacterial Markers in Water and Sediments of

Tiaoxi River (Taihu Watershed). *Frontiers in Microbiology*, 10.

<https://doi.org/10.3389/fmicb.2019.00699>

Varriano, S., Snyder, W., Smith, J. C., Shariat, N. W., & Dunn, L. L. (2025). Detering Wild Birds During Fruit And Vegetable Production. *Food Protection Trends*, 45(1), 27–35.

Vinueza, D., Ochoa-Herrera, V., Maurice, L., Tamayo, E., Mejía, L., Tejera, E., & Machado, A. (2021). Determining the microbial and chemical contamination in Ecuador's main rivers. *Scientific Reports*, 11(1), 17640.

<https://doi.org/10.1038/s41598-021-96926-z>

Weller, D., Brassill, N., Rock, C., Ivanek, R., Mudrak, E., Roof, S., Ganda, E., & Wiedmann, M. (2020). Complex Interactions Between Weather, and Microbial and Physicochemical Water Quality Impact the Likelihood of Detecting Foodborne Pathogens in Agricultural Water. *Frontiers in Microbiology*, 11.

<https://doi.org/10.3389/fmicb.2020.00134>

APPENDIX A

Supplemental Material for Chapter 2

Supplemental Table 2.1. ΔAICc and Akaike weight (ω) for detection of *Salmonella* and weather effect models.

Model	β Value (SE, p)			ΔAICc	ω
	Precipitation	Humidity	Wind		
Humidity		-0.2898 (0.179, 0.105)		0	0.379
Precipitation + Humidity	0.2348 (0.190, 0.215)	-0.3375 (0.183, 0.065)		0.45	0.289
Precipitation + Humidity + Wind	0.2788 (0.197, 0.158)	-0.2833 (0.190, 0.129)	-0.2006 (0.208, 0.335)	1.54	0.169
Precipitation	0.165 (0.189, 0.381)			1.88	0.139

Supplemental Table 2.2. Genome assembly statistics

	Genome Size (Mb)	Number of contigs	N50 (Kb) Value
JSBird1	4.95	132	103.59
JSBird2	4.99	193	66.87
JSBird3	4.74	129	114.65
JSBird4	4.74	122	115.51
JSBird5	4.75	40	735.89
JSBird6	4.77	101	140.49
JSBird7	4.77	102	159.79
JSBird8	4.77	137	98.89
JSBird9	4.77	122	127.85
JSBird10	4.74	121	94.07
JSBird11	4.67	37	613.89
JSBird12	5.03	129	101.89
JSBird13	4.84	184	86.61
JSBird14	4.91	177	71.96
JSBird15	4.84	149	102.96
JSBird16	4.61	268	36.06
JSBird19	4.80	55	405.57
JSBird20	4.85	57	430.46
JSBird21	4.92	59	417.46

North Georgia	
May	June
Temperature (°F)	3.88
Precipitation (in)	-1.41
Humidity (%)	4.61
Wind (mph)	-0.09
July	August
Temperature (°F)	2.05
Precipitation (in)	0.12
Humidity (%)	-1.18
Wind (mph)	0.45
September	October
Temperature (°F)	-0.83
Precipitation (in)	0.56
Humidity (%)	-0.96
Wind (mph)	-8.58
	-13.71
	0.93

*Sampling in 2021 took place between June 13th-October 14th

*Sampling in 2022 took place between May 31st-August 24th

South Georgia/Florida	
May	June
Temperature (°F)	2.12
Precipitation (in)	4.82
Humidity (%)	9.48
Wind (mph)	-0.87
July	August
Temperature (°F)	2.38
Precipitation (in)	-0.89
Humidity (%)	-2.43
Wind (mph)	0.11
September	October
Temperature (°F)	1.00
Precipitation (in)	-3.08
Humidity (%)	-2.79
Wind (mph)	-1.27
	1.86
	0.59

*Sampling in 2021 took place between May 26th-September 18th

*Sampling in 2022 took place between May 23rd-June 29th

Tennessee	
May	June
Temperature (°F)	4.13
Precipitation (in)	0.04
Humidity (%)	2.91
Wind (mph)	0.64
July	August
Temperature (°F)	0.97
Precipitation (in)	0.20
Humidity (%)	4.02
Wind (mph)	0.63
September	October
Temperature (°F)	-1.06
Precipitation (in)	-0.24
Humidity (%)	0.16
Wind (mph)	0.23
	1.12
	0.81

*Sampling in 2021 took place between June 23rd-October 1st

*Sampling in 2022 took place between June 22nd-August 30th

South Carolina	
May	June
Temperature (°F)	2.69
Precipitation (in)	0.30
Humidity (%)	4.40
Wind (mph)	0.62
July	August
Temperature (°F)	2.55
Precipitation (in)	0.94
Humidity (%)	2.41
Wind (mph)	0.86
September	October
Temperature (°F)	-1.05
Precipitation (in)	-0.28
Humidity (%)	0.04
Wind (mph)	0.16
	1.32
	-0.01

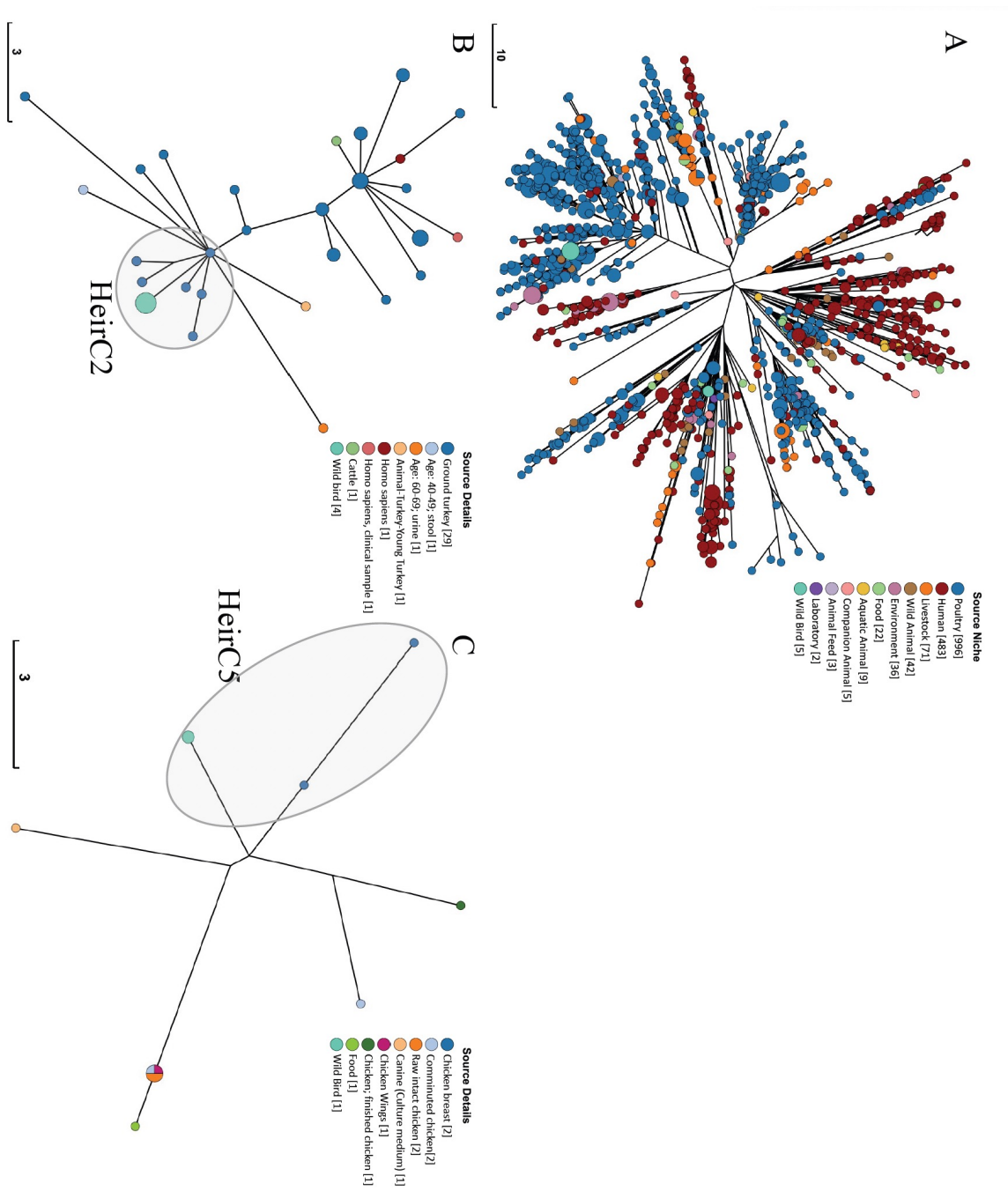
*Sampling in 2021 took place between August 3rd-October 21st

*Sampling in 2022 took place between June 9th-August 10th

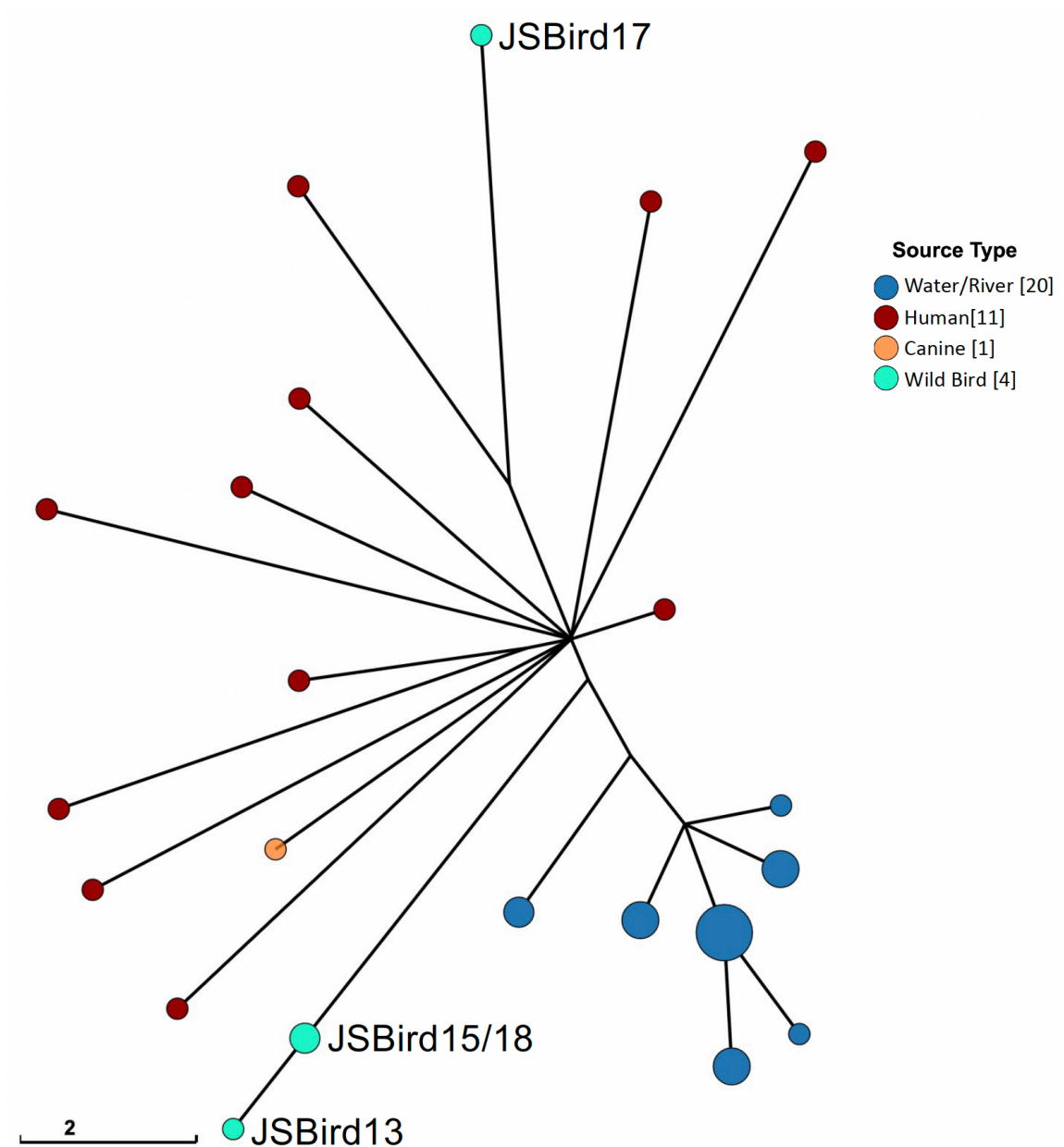
Supplemental Figure 2.1. Comparison of weather conditions for sampling regions.

Temperature, precipitation, humidity, and wind were compared between 2021 and 2022 collection seasons. Averages for each variable (except precipitation where cumulative rainfall was calculated) are shown in under the respective month for each region.

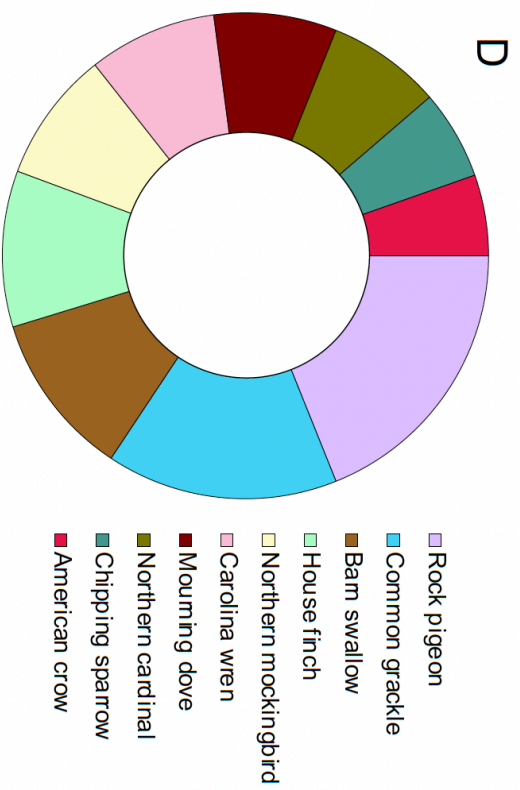
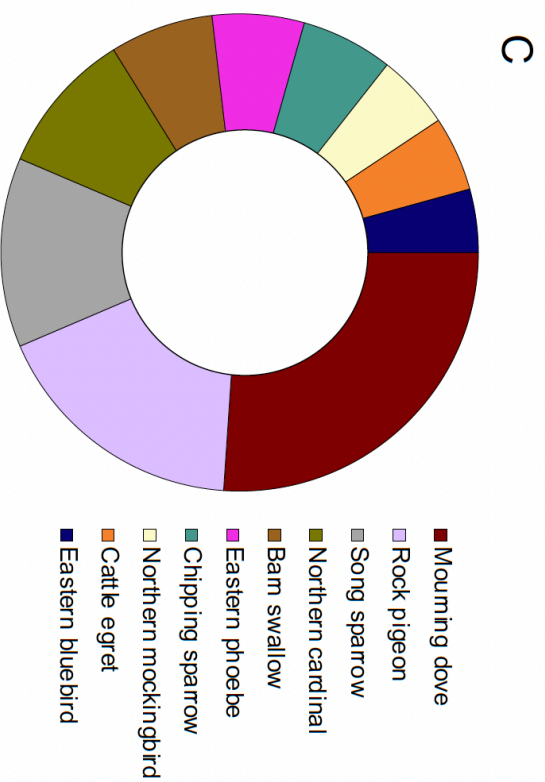
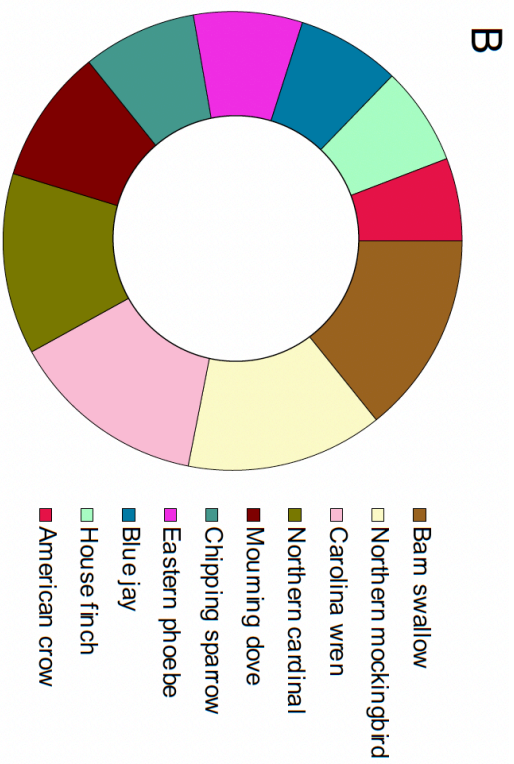
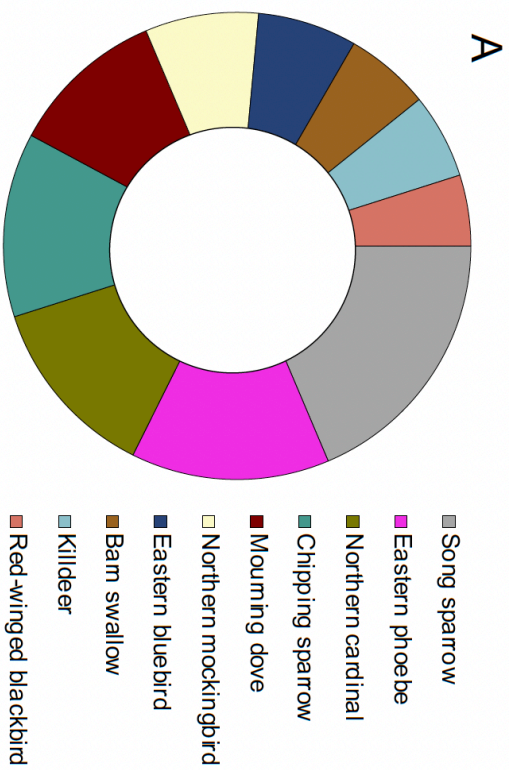
Negative numbers (red boxes) indicate a lower value in 2022 and positive numbers (blue boxes) indicate a higher value in 2022. Dates of the first and last sampling in each region for both years are listed to the right of the weather values.



Supplemental Figure 2.2. Serovar Hadar isolates include variable poultry relationships between turkey and chicken isolates. (A) Phylogenetic relationships between serovar Hadar isolates from wild bird feces in this study (turquoise circles) to isolates on Enterobase shown on a GrapeTree plot. (B) JSBird3, 4, 5, and 10 isolates and Enterobase isolates with up to 10 cgMLST allelic differences. (C) JSBird11 and Enterobase isolates with up to 10 cgMLST allelic differences. (Scale bar) Number of cgMLST allelic differences.



Supplemental Figure 2.3. Serovar Newport isolates are similar to human and environmental isolates. Serovar Newport isolates from wild bird feces (turquoise circles) with phylogenetic relationships to isolates within Enterobase shown on a GrapeTree plot. (Scale bar) Number of cgMLST allelic differences.



Supplemental Figure 2.4. Wild bird point counts. Top 10 species for the following categories: (A) Visits with species observation in-field, (B) visits with species observation off-field, (C) in-field individual observations, and (D) off-field individual observations.

APPENDIX B

Supplemental Material for Chapter 3

Supplementary Materials and Methods

Sample collection and *Salmonella* isolation

A modified Moore swab and peristaltic pump were used to collect 10 L of creek water from each site (n = 456 samples)(Figure S1A). Modified Moore swabs were placed at approximately midchannel resting on the sediment for the duration of pumping. The swabs were stored on ice until returning to the lab, where 100 ml of buffered peptone water (BPW; Neogen, Lansing, MI) and novobiocin (40 mg/L; Thermo Scientific Chemicals, Waltham, MA) were added. The swabs were hand-massaged for 1 minute, then incubated, shaking at 42°C for 20-24 hours. Following incubation, 1 and 0.1 mL of BPW were transferred into 10 mL of tetrathionate (TT; Hardy Diagnostics, Santa Maria, CA) and Rappaport-Vassiliadis (RV; Hardy Diagnostics, Santa Maria, CA) broth, respectively, and statically incubated at 37°C for 20-24 hours. Starting at month 13, 0.1 mL of enriched BPW was inoculated into a modified semisolid Rappaport-Vassiliadis (MSRV; Hardy Diagnostics, Santa Maria, CA) plate and incubated at 37°C for 20-24 hours. *Salmonella* presence was tested by streaking the TT and RV cultures, along with any presumptive *Salmonella* growth from the MSRV plates, onto xylose lysine tergitol-4 (XLT-4; Hardy Diagnostics, Santa Maria, CA) plates, followed by incubation at 37°C for 24-48 hours. During the first 12 months, if the TT and RV enrichments did not result in presumptive *Salmonella* colonies, then the enriched BPW culture was directly plated onto XLT-4 (Figure S1B). Three colonies were selected and restreaked for isolation onto new XLT-4 plates. After incubation at 37°C for 24-48 hours, a single colony was streaked onto Luria-Bertani (LB; Hardy Diagnostics, Santa Maria, CA) agar, then confirmed with serum agglutination (BD Difco, Franklin Lakes, NJ). All confirmed *Salmonella* isolates

were saved via glycerol stocks in a -80°C freezer. From all enrichments, 1 mL was removed and centrifuged at 18,000 rcf for 3 minutes, then the supernatant was removed, and the pellets were stored at -20°C for later use.

Antimicrobial susceptibility testing

One *Salmonella*-positive isolate from each site in each month was selected for antimicrobial susceptibility testing (AST). When selecting isolates, preference was given to the first colony recovered from TT, followed by RV and BPW. To measure the susceptibility of the *Salmonella* isolates, the Kirby-Bauer disk diffusion assay was used to test ten different antibiotics: ampicillin (10 µg), amoxicillin-clavulanic acid (20/10 µg), ceftriaxone (30 µg), gentamicin (10 µg), streptomycin (10 µg), tetracycline (30 µg), ciprofloxacin (5 µg), sulfamethoxazole-trimethoprim (25 µg), nalidixic acid (30 µg), and chloramphenicol (30 µg) (BD Diagnostics, Franklin Lakes, NJ). Bacterial overnight cultures were standardized to 0.5 McFarland (ThermoScientific Remel, Lenexa, KS) and spread plated onto Mueller-Hinton agar (BD BBL, Franklin Lakes, NJ). Antibiotic disks were placed on the plates, followed by incubation for 18-22 hours at 37°C. Zones of inhibition were measured and the results were interpreted with the CLSI standards to determine if isolates were susceptible, intermediate, or resistant (CLSI, 2024). Isolates were categorized as multidrug resistant (MDR) if they were resistant to three or more classes of antibiotics. For subsequent analyses, an isolate with intermediate resistance was considered as “susceptible”.

Whole genome sequencing

To confirm the serovar of each isolate with antimicrobial resistance and compare the relatedness to publicly available *Salmonella* genomes, a single colony was picked from

the LB plate used for *Salmonella* confirmation, inoculated in an LB agar stab, and shipped to the FDA Center for Food Safety and Applied Nutrition (FDA-CFSAN, now Human Food Program) for library preparation and sequencing. Raw Illumina sequence reads were uploaded to NCBI (BioProject PRJNA186035) for further processing through the Prokaryotic Genome Annotation Pipeline and the resulting assemblies were used for further analysis (Serovar Infantis: BioSample ; Serovar Typhimurium: BioSample).

Supplemental Table 3.1. Final generalized linear mixed model for *Salmonella* prevalence.

	Estimate	Std. Error	z value	Pr(> z)	OddsRatio	LowerCI	UpperCI
(Intercept)	0.15033	4.522179	0.033243	0.973481	1.162	0	8210.097
Spring	0.739343	0.259482	2.849307	0.004381	2.094	1.26	3.479
Summer	-2.772454	0.300477	-9.226852	2.79E-20	0.063	0.035	0.113
Winter	0.352735	0.261828	1.347204	0.177915	1.423	0.852	2.379
Temp	-0.05905	0.070748	-0.834652	0.403913	0.943	0.82	1.083
Humidity	-0.132516	0.104371	-1.269668	0.204203	0.875	0.714	1.073
Wind	0.18682	0.046893	3.983976	6.78E-05	1.206	1.1	1.322
Precip	0.188953	0.033516	5.637634	1.72E-08	1.208	1.13	1.291
Hay pasture	0.033582	0.005133	6.542097	6.07E-11	1.035	1.024	1.045
Mixed forest	0.115867	0.020004	5.792277	6.94E-09	1.123	1.08	1.168
I(Temp^2)	0.000896	0.000494	1.81315	0.069809	1.001	1.001	1.001
I(Humidity^2)	0.00143	0.000733	1.949504	0.051235	1.001	0.999	1.003

Supplemental Table 3.2. Final generalized linear mixed model for *Salmonella* population complexity.

	Estimate	Std. Error	df	t value	Pr(> t)	OddsRatio	LowerCI	UpperCI
(Intercept)	-3.5375163	1.27479562	106.278724	-2.7749674	0.00652395	0.029	0.002	0.354
Spring	0.74102359	0.30118276	248.021969	2.46037848	0.01456161	2.098	1.163	3.785
Summer	-0.507279	0.35467614	248.931894	-1.4302598	0.15389642	0.602	0.3	1.208
Winter	-0.7368902	0.3038573	247.200975	-2.4251194	0.01602059	0.479	0.264	0.868
Humidity	0.05412746	0.01270287	248.944623	4.26104181	2.89E-05	1.055	1.029	1.083
Wind	0.13041264	0.05581676	247.079948	2.33644212	0.02026832	1.139	1.02	1.271
Hay pasture	0.02960848	0.00755701	9.32887179	3.91801513	0.00328558	1.03	1.014	1.047
Mixed forest	0.14233459	0.03648874	4.71875035	3.90078175	0.0127588	1.153	1.074	1.237

Supplemental Table 3.3. Full list of serovars identified by USDA-FSIS in food animal samples and clinical cases, along with the serovars identified in samples from this study.

BEAM top 10	Beef	Chicken	Pork	Turkey	Creek
Enteritidis 24780	Montevideo 103	Kentucky 2615	Anatum 335	Hadar 99	Rubislaw 163
Newport 15610	Anatum 62	Infantis 2578	I 4, [5], 12 i:- 273	Schwarzengrund 64	Muenchen I 133
Typhimurium 12661	Muenchen 49	Enteritidis 1015	Infantis 220	Infantis 63	Montevideo II 126
Javiana 7565	Dublin 30	Typhimurium 839	Derby 181	Other 54	Aqua/Inverness 96
I 4, 5 12 i:- 5755	Infantis 30	Schwarzengrund 381	Johannesburg 166	Senftenberg 50	Give I 70
Infantis 5506	Newport 30	Alachua 78	Eko 153	Reading 47	Typhimurium 53
Oranienburg 4386	Cerro 29	Thompson 78	London 151	Typhimurium 36	Gaminara 32
Saintpaul 3634	Muenster 26	I 4, [5], 12 i:- 51	Agona 110	Muenchen 30	Infantis 31
Braenderup 3595	Kentucky 22	Johannesburg 51	Uganda 107	Uganda 26	Hartford 29
Muenchen 3330	Typhimurium 19	Other 37	Typhimurium 78	I 4, [5], 12 i:- 12	Braenderup 26
	Brandenburg 13	Braenderup 34	Muenchen 62	Anatum 8	Mississippi II 23
	Agona 12	Hadar 27	Adelaide 57	Brandenburg 8	Brazil I 22
	Uganda 12	Heidelberg 25	Ohio 56	Montevideo 8	Montevideo I 22
	Altona 9	Mbandaka 20	Worthington 56	Albany 7	Alabama/Bareilly III 13
	Give 9	Blockley 17	Berta 51	Newport 5	Oranienburg 13
	I 4, [5], 12 i:- 9	Senftenberg 17	Rissen 37	Worthington 5	Anatum 12
	Mbandaka 9	Anatum 15	Schwarzengrund 37	Enteritidis 4	Thompson 12
	Meleagridis 7	Montevideo 11	Chailey 33	Kentucky 4	Muenchen II 9
	Schwarzengrund 7	Agona 10	Brandenburg 28	Orion 4	Newport II 9
	Other 6	Muenchen 10	Senftenberg 28	Ouakam 4	Hadar 8
	Enteritidis 6	Newport 8	Mbandaka 21	Cubana 3	Agbeni 7
	Eko 5	Cerro 7	Manhattan 20	Litchfield 3	Kentucky I 7
	London 5	Berta 6	Panama 20	Bovismorbificans 2	Enteritidis 5
	Barranquilla 4	Litchfield 6	Saintpaul 20	Heidelberg 2	Kiambu 5
	Derby 4	Uganda 4	Litchfield 19	Johannesburg 2	Luciana 5
	Lubbock 4	Eko 3	Reading 19	Liverpool 2	Untypeable 5
	Senftenberg 4	Muenster 3	Soerenga 18	London 2	Johannesburg/Urbana 2
	Thompson 4	Reading 3	Bovismorbificans 17	Muenster 2	Rissen 2
	Adelaide 3	Tennessee 3	Alachua 15	Amsterdam 1	Senftenberg I 2
	Braenderup 3	Albany 2	Cerro 14	Barranquilla 1	Uganda 2
	Amsterdam 2	Derby 2	Give 14	Gateshead 1	Agona 1
	Bovismorbificans 2	Elomrane 2	Montevideo 14	Give 1	Cubana 1
	Bredeney 2	Fresno 2	Newport 14	Olten 1	Derby III 1
	Idikan 2	Liverpool 2	Other 13	Rissen 1	Dublin 1
	Johannesburg 2	Oranienburg 2	Heidelberg 9	Ruiru 1	Java 1
	Kiambu 2	Ouakam 2	Kentucky 9		Mbandaka 1
	Litchfield 2	Widmarsh 2	Kiambu 9		Poona 1
	Ohio 2	Barranquilla 1	Braenderup 8		Saintpaul I 1
	Orion 2	Brandenburg 1	Muenster 8		Saintpaul II 1
	Reading 2	Chailey 1	Bredeney 6		Senftenberg II 1
	Saintpaul 2	Dublin 1	Orion 6		
	Albany 1	Gateshead 1	Hadar 5		
	Apapa 1	Give 1	Liverpool 5		
	Bareilly 1	Havana 1	Altona 4		
	Berta 1	Livingstone 1	Barranquilla 4		
	Blockley 1	London 1	Havana 4		
	Cannstatt 1	Minnesota 1	Livingstone 4		
	Corvallis 1	Ohio 1	Kedougou 3		
	Eastbourne 1	Orion 1	Thompson 3		
	Fresno 1	Oslo 1	Cannstatt 2		
	Gaminara 1	Putten 1	Choleraesuis 2		
	Heidelberg 1	Rissen 1	Enteritidis 2		
	Manhattan 1	Roodepoort 1	Indiana 2		
	Minnesota 1	Soerenga 1	Krefeld 2		
	Panama 1	Worthington 1	Meleagridis 2		
	Paratyphi B var. L(+) tartrate+ 1		Minnesota 2		
	Poona 1		Tennessee 2		
	Takoradi 1		Gatuni 1		
			Hartford 1		
			Idikan 1		
			Poona 1		

Supplemental Table 3.4. Select metadata downloaded from NCBI for serovar Infantis isolates included in Figure 3.4A.

BioSample	Collected by	Collection date	Location	Isolation type	Isolation source	SNP cluster	Assembly	Run	Isolate	Strain	Computed types
SAMN09444132	USDA-FSIS	2018	USA:NE	environmental/other	animal-swine-market swine	PDS000097345.3	GCA_008447705.1	SRR7358414	PDT000332042.1	FSIS11810649	antigen_formula=7:r:1,5_serotype=Infantis
SAMN22215935		2021-08	USA	clinical		PDS000097345.3	GCA_020465275.1	SRR16292359	PDT001148831.1	PNUSAS235240	antigen_formula=7:r:1,5_serotype=Infantis
SAMN24114991	Not Collected	2017	USA	clinical	Not Collected	PDS000097345.3	GCA_021719155.1	SRR17247291	PDT001232544.1	18-038875-057	antigen_formula=7:r:1,5_serotype=Infantis
SAMN25608883		2014	Slovenia	clinical		PDS000116014.1	GCA_024995025.1	SRR17880716	PDT001389277.1	S105	antigen_formula=7:r:1,5_serotype=Infantis
SAMN25608903		2015	Slovenia	clinical		PDS000116014.1	GCA_024992195.1	SRR17880860	PDT001389419.1	S125	antigen_formula=7:r:1,5_serotype=Infantis

Supplemental Table 3.5. Select metadata downloaded from NCBI for serovar Infantis isolates included in Figure S3.6A.

Yellow highlight indicates samples chosen for Figure S3.6B.

BiSample	Collected by	Collection date	Location	Isolation type	Isolation source	SNP cluster	Assembly	Run	Isolate	Strain	Computed types
SAM00019585		2000	Japan Kyusyu			PS000116910.1	GCA_Q24958125	DRR022723	PD001391564.1	1637	antigen, formula+? ? ? 1.5, serotype+Infantis
SAM00019584		2000	Japan Kyusyu			PS000116938.2	GCA_Q24957865	DRR022740	PD001391571.1	1642	antigen, formula+? ? ? 1.5, serotype+Infantis
SAM00019510		1998	Japan Kyusyu			PS000118011.2	GCA_Q24957445	DRR022728	PD001391559.1	407	antigen, formula+? ? ? 1.5, serotype+Infantis
SAM00019612		1998	Japan Kyusyu			PS000116172.1	GCA_Q24958065	DRR022730	PD001391561.1	426	antigen, formula+? ? ? 1.5, serotype+Infantis
SAM00019817		1999	Japan Kyusyu			PS000116013.2	GCA_Q24958245	DRR022717	PD001391548.1	800	antigen, formula+? ? ? 1.5, serotype+Infantis
SAM00019823		2000	Japan Kyusyu			PS000116034.2	GCA_Q24957485	DRR022716	PD001391547.1	G117	antigen, formula+? ? ? 1.5, serotype+Infantis
SAM00019828		2000	Japan Kyusyu			PS000116032.1	GCA_Q24957485	DRR022758	PD001391588.1	R60	antigen, formula+? ? ? 1.5, serotype+Infantis
SAM00019843		2000	Japan Kyusyu			PS000116159.1	GCA_Q24958983	DRR022780	PD001391610.1	R116	antigen, formula+? ? ? 1.5, serotype+Infantis
SAM00019846		2000	Japan Kyusyu			PS000116163.1	GCA_Q24958945	DRR022774	PD001391604.1	G83	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA104349014		2017	Ireland	clinical	human	PS000117390.1	GCA_Q25075726	ERR217672	PD0010045699.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA104349051		2017	Germany		food	PS000027773.3	GCA_Q101165115	ERR220348	PD00027773.3		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA104349559		2015	Germany		food	PS000018208.4	GCA_Q10417045	ERR2200357	PD00027786.3		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA104349512		2018	Germany		food	PS000027748.7	GCA_Q10461468	ERR220072	PD00027780.3		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA104455251	DTU	19/18	Denmark		feed	PS000073471.1	GCA_Q10142175	ERR2433502	PD00007330.2		Infantis
SAMEA111550573	Hospital of The Social Welfare Institute	11/18/20	Paraguay	clinical	feces	PS000125084.1	GCA_Q26352875	ERR10432297	PD0014933329.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA112357034		2008	Rugs			PS0000208485.81	GCA_Q331185015	ERR176944	PD000181605.1	VISAVIE_V008_Q29558M	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580037	National Institute for Communicable Disease, South Africa	3/28/04	South Africa	clinical	Human Rectal Swab	PS000109857.5	GCA_Q23492265	ERR756218	PD001010290.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580041	National Institute for Communicable Disease, South Africa	11/29/04	South Africa	clinical	Human Stool	PS000109868.5	GCA_Q23492275	ERR756222	PD0010110294.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580042	National Institute for Communicable Disease, South Africa	11/29/04	South Africa	clinical	Human Rectal Swab	PS000109864.9	GCA_Q23492025	ERR756223	PD0010110295.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580050	National Institute for Communicable Disease, South Africa	12/15/06	South Africa	clinical	Human Stool	PS000109881.5	GCA_Q23491805	ERR756236	PD0010110308.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580163	National Institute for Communicable Disease, South Africa	5/7/12	South Africa	clinical	Human Stool	PS000111747.2	GCA_Q24225515	ERR717193	PD001381509.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580170	National Institute for Communicable Disease, South Africa	12/14/12	South Africa	clinical	Human Blood culture	PS000114466.1	GCA_Q24479326	ERR717403	PD001310691.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580171	National Institute for Communicable Disease, South Africa	11/6/13	South Africa	clinical	Human Stool	PS000077867.3	GCA_Q24267295	ERR761404	PD001358639.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580184	National Institute for Communicable Disease, South Africa	10/15/13	South Africa	clinical	Human Stool	PS000013888.3	GCA_Q24271165	ERR761436	PD001358683.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580200	National Institute for Communicable Disease, South Africa	8/18/14	South Africa	clinical	Human Stool	PS000114825.1	GCA_Q24440378	ERR761471	PD001358855.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580253	National Institute for Communicable Disease, South Africa	11/25/06	South Africa	clinical	Human Stool	PS000109873.5	GCA_Q23498465	ERR756235	PD0010110307.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580260	National Institute for Communicable Disease, South Africa	2/18/07	South Africa	clinical	Human Stool	PS000109881.7	GCA_Q23491725	ERR756245	PD0010110317.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580266	National Institute for Communicable Disease, South Africa	7/18/07	South Africa	clinical	Human Stool	PS000109862.3	GCA_Q23498345	ERR756272	PD0010110322.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580286	National Institute for Communicable Disease, South Africa	12/17/07	South Africa	clinical	Human Stool	PS000109867.2	GCA_Q23498265	ERR756277	PD0010110327.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580301	National Institute for Communicable Disease, South Africa	4/20/09	South Africa	clinical	Human Stool	PS000011736.7	GCA_Q23880115	ERR756300	PD000138760.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580303	National Institute for Communicable Disease, South Africa	10/18/13	South Africa	clinical	Human Blood culture	PS000114212.1	GCA_Q24263185	ERR761439	PD001358610.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580310	National Institute for Communicable Disease, South Africa	11/8/13	South Africa	clinical	Human Stool	PS000014238.2	GCA_Q23962550	ERR761450	PD001358630.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580336	National Institute for Communicable Disease, South Africa	2/7/14	South Africa	clinical	Human Stool	PS000113875.1	GCA_Q24270775	ERR761453	PD001358545.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA11580399	National Institute for Communicable Disease, South Africa	1/15/14	South Africa	clinical	Human Stool	PS000154629.1	GCA_Q24478585	ERR761450	PD001370707.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA1535595	Istituto Zooprofilattico Sperimentale delle Regioni Lazio e Toscana, Via Appia Nuova 1411, 00176, Rome, Italy	2007	Italy			PS000008352.5	GCA_Q10179715	ERR1012862.3	PD00012862.3		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA3535995	Istituto Zooprofilattico Sperimentale delle Regioni Lazio e Toscana, Via Appia Nuova 1411, 00176, Rome, Italy	2009	Italy			PS000058372.1	GCA_Q11033465	ERR1014110	PD000128412.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA4668129		2016	Germany		animal	GCA_Q10181655	ERR258602	PD000343710.1			antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA4932113						PS000005624.71	GCA_Q11230565	ERR431713	PD00014467.1	MA_RM_119	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA5565920						PS000077492.1	GCA_Q13640815	ERR428988	PD000773338.1	MA_N007_023	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA5575925						PS000068608.9	GCA_Q12599675	ERR438231	PD000739717.1	MA_C08_063	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA5978362	ISLT	2016	Italy	clinical		PS000008630.1	GCA_Q144807625	ERR426286	PD000008630.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA6057638		2018	Germany		food	PS000051895.5	GCA_Q10294343	ERR3849973	PD000064642.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA6056020		2013	Germany		animal	PS000051862.1	GCA_Q20618565	ERR381356	PD000064651.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA6056279		2018	Germany		animal	PS000051922.5	GCA_Q10295885	ERR381613	PD000067107.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA6262611		7/5/17	Turkey		chicken wing	PS000098908.1	GCA_Q20994405	ERR389721	PD001186441.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA6263512		10/28/17	Turkey		chicken wing	PS000098965.1	GCA_Q20994425	ERR389722	PD001186441.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA6263625		5/10/17	Turkey		chicken legs	PS000098981.1	GCA_Q20994105	ERR3897235	PD001186455.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA6942336		1999	Germany		boiler	PS000077687.3	GCA_Q18111588	ERR423826	PD000918645.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA7464848		2018	Germany		boiler	PS000016386.2	GCA_Q30141375	ERR432888	PD000113735.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA7828918		2016	Ireland	clinical	human	PS000117335.1	GCA_Q23084235	ERR110627	PD000441550.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMEA8053918		2015	Ireland	clinical	human	PS000114931.1	GCA_Q23298345	ERR182816	PD000441516.1		antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN11902298	NYSDOH	8/16/07	USA,NY	environmental+other	food	PS000108885.140	GCA_Q11455385	SRR549438	PD000000719.3	NY_BAC070006871	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN11902343	NYSDOH	8/22/09	USA,NY	environmental+other	chicken	PS000003388.53	GCA_Q11455905	SRR1158047	PD000003307.5	NY_BAC090003911	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN11902347	NYSDOH	8/17/08	USA,NY	environmental+other	green onion	PS000033022.4	GCA_Q11455085	SRR1106268	PD000002950.3	NY_BAC090004010	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12535058	FL	12/21/10	USA,FL	environmental+other	cilantro	PS000085033.850	GCA_Q11460595	SRR1041446	PD000001635.3	FL_FLDACS-08003	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12521148	New Mexico Public Health Lab	2010	USA,NM	environmental+other	ground turkey	PS000002971.74	GCA_Q38809275	SRR117476	PD000001026.6	AZ_T05844	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN1255314	FDA Contracted Laboratory	10/12/07	USA,SD	environmental+other	bovine carcass	PS000002392.298	GCA_Q38839415	SRR149819	PD000000481.2	CFSA0000987	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12451507	United States Food and Drug Administration/Office of Regulatory Affairs/ New York Human and Animal Food Laboratory	2/17/11	USA	environmental+other	snail	GCA_Q102702915	SRR1164504	PD000003330.3	FNE0723	antigen, formula+? ? ? 1.5, serotype+Infantis	
SAMN12451592	United States Food and Drug Administration/Office of Regulatory Affairs/ Atlanta Human and Animal Food Laboratory	8/13/11	USA	environmental+other	fromen calamari ring	PS0000032466.13	GCA_Q11633805	SRR121230	PD000002179.3	FSE0064	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12452171	United States Food and Drug Administration/Office of Regulatory Affairs/Windward Engineering Analytical Center	5/26/12	USA	environmental+other	peas	PS000002981.9	GCA_Q10268105	SRR1179187	PD000003560.3	FMA0059	antigen, formula+? ? ? 1.5, serotype+Infantis
FDA_SAMN12367727	FDA	2011	USA,CA	clinical	chicken sample	PS000001365.4	GCA_Q10633205	SRR1060648	PD000002227.3	IVB_38572	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12671468	FDA Shaoxun Zhao	2008-11	USA,CA	environmental+other	chicken breast	PS000007076.841	GCA_Q10636055	SRR1123618	PD000002888.3	N2304	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12470574		2008-11	clinical		colt sample	PS000002389.10	GCA_Q00006925		PD000002091.2	115944	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12483567	United States Food and Drug Administration/Office of Regulatory Affairs/ San Francisco Human and Animal Food Laboratory	3/23/11	USA	environmental+other	cat claw powder herbal product	PS000105341.2	GCA_Q11626455	SRR1157845	PD000003302.3	FSP009	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12484524	Minnesota Department Of Health	9/16/03	USA,MN	environmental+other	feces	PS000003248.6	GCA_Q08762025	SRR128478	PD000003027.4	MDH-2014-00050	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12484532	Minnesota Department Of Health	9/16/03	USA,MN	environmental+other		PS000003206.8	GCA_Q11080865	SRR1282397	PD000002978.3	MDH-2014-00059	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12678547	United States Food and Drug Administration	2/10/13	USA	environmental+other	environmental swab	PS000002846.6	GCA_Q08630545	SRR1481663	PD000004807.4	FC02141	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12678643	United States Food and Drug Administration	2/10/13	USA	environmental+other	brewer yeast	PS000002265.2	GCA_Q08669025	SRR1484547	PD000001889.4	FM01022	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12678851	United States Food and Drug Administration	5/01/13	USA	environmental+other	edid parley	PS000002846.17	GCA_Q11080545	SRR1258667	PD000002671.3	FSW0124	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12678853	United States Food and Drug Administration	11/8/12	USA	environmental+other	pumpkin seed powder	PS000002243.3	GCA_Q01741175	SRR1484516	PD000007683.3	FSW0158	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12686448	United States Food and Drug Administration	9/8/10	USA	environmental+other	meat bone meal	PS000001702.5	GCA_Q10205335	SRR1633322	PD000004337.3	FSW0192	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12699512	Minnesota Department of Health	11/15/02	USA,MN	environmental+other	tissue	PS000004301.3	GCA_Q11448675	SRR1461812	PD000003963.3	MDH-2014-00348	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12699501	Minnesota Department of Health	8/1/03	USA,MN	environmental+other		PS000003186.7	GCA_Q08629545	SRR108670	PD000003974.4	MDH-2014-00525	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12742080	FDA Contracted Laboratory	11/8/06	USA,WI	environmental+other	porcine colon	PS000003240.2	GCA_Q11632405	SRR1292224	PD000003050.3	CFSA0012847	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12845479	United States Food and Drug Administration/Office of Regulatory Affairs/Pacific Northwest Laboratory	8/7/04	USA	environmental+other	scorpion	PS000004455.1	GCA_Q27736355	SRR3627072	PD000008203.3	PD268781.2	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12845488	United States Food and Drug Administration	6/17/04	USA	environmental+other	fromen guinea meat	PS000111461.1	GCA_Q24107855	SRR1869738	PD001032471.1	FD438320 B	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12845835	United States Food and Drug Administration/Office of Regulatory Affairs/ Arkansas Human and Animal Food Laboratory	11/25/08	USA	environmental+other	soy lecithin fluid	PS000001930.2	GCA_Q08115525	SRR205814	PD000006692.2	FD439523.1 B	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12846127	United States Food and Drug Administration/Office of Regulatory Affairs/Southeast Food and Feed Laboratory	6/24/08	USA	environmental+other		PS000001894.8	GCA_Q08913563	SRR345701	PD000001759.4	FD447418 sub A val040	antigen, formula+? ? ? 1.5, serotype+Infantis
SAMN12841707	United States Food and Drug Administration/Office of Regulatory Affairs/ Irvine Human and Animal Food Laboratory	10/26/10	USA	environmental+other	fromen cat crab	PS000001805.4					

SAMN02919009	United States Food and Drug Administration/Office of Regulatory Affairs/Pacific Northwest Laboratory	4/20/13	USA	environmental/other	pet treat	PD00001634113	GCA_008986375.1	SR04418762	PD00001634113	FDAB098373.1	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN0318567	United States Food and Drug Administration/Office of Regulatory Affairs New York Human and Animal Food Laboratory	7/22/14	USA	environmental/other	shell stuffed carmelon cheese	PD000031468	GCA_010533069.1	SRK1969993	PD000031474	FDAB07173	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN0318568	FDA Contracted Laboratory	3/9/13	USA-MO	environmental/other	environmental swab sponge	PD000031468.1	GCA_008432951.1	SRK1614991	PD000040973.2	CFSAN025196	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03192331	FDA Contracted Laboratory	5/17/13	USA-CA	environmental/other	produce	PD000032395564	GCA_008512061.1	SRK1615207	PD000040918.2	CFSAN025208	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03192338	PHE	2/12/09	United Kingdom: South of England	clinical	human	PD000037701.61	GCA_006287049.1	SRK1623088	PD000041099.4	H12240441	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03195014	PHE	2012-09	United Kingdom: South of England	clinical	human	PD000032925940	GCA_008311551.1	SRK1654513	PD000041041.4	H123540745	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03196128	PHE	2012-05	United Kingdom: South of England	clinical	human	PD000032463148	GCA_011076891.1	SRK1645344	PD000042952.4	H122540361	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03196129	PHE	2012-10	United Kingdom: Midlands and East of England	clinical	human	PD00003038511.3	GCA_004270043.1	SRK1654493	PD000043051.4	H12404043	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03196129	PHE	2012-11	United Kingdom: South of England	clinical	human	PD00003294248	GCA_008030061.1	SRK1645584	PD000043275.5	H124740809	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03196191	PHE	2012-04	United Kingdom: North of England	clinical	human	PD000030087428	GCA_008020091.1	SRK1643903	PD000043243.5	H12140568	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03196192	PHE	2012-11	United Kingdom: Midlands and East of England	clinical	human	PD00003038511.3	GCA_008020091.1	SRK1645584	PD000043243.5	H124740809	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03196933	PHE	2012-09	United Kingdom: South of England	clinical	human	PD000040443479	GCA_008119071.1	SRK1645987	PD000044715.5	H123805027	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03196413	PHE	2013-01	United Kingdom: South of England	clinical	human	PD000037444.61	GCA_000202471.1	SRK1641672	PD000043052.1	H12020449	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03177794	United States Food and Drug Administration/Office of Regulatory Affairs New York Human and Animal Food Laboratory	11/6/14	USA	environmental/other	individually quick frozen scallop adductor	PD0000343225	GCA_006119251.1	SRK1690919	PD000044193.2	FDAB2253.1	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03198677	FDA Contracted Laboratory	4/16/07	USA-SD	environmental/other	beef	PD000031468.2	GCA_008556401.1	SRK1722878	PD000045990.2	CFSAN021146	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03216239	USDA-FSIS	2012	USA-PA	environmental/other	egg yolk	PD0000354887.1	GCA_008518101.1	SRK1745555	PD000047093.3	CFSAN024812	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03253335	Middle East Technical University Food Engineering Department	7/18/12	Turkey	environmental/other	chicken meat	PD000031430.1	GCA_004192975.1	SRK1745554	PD000047098.3	CFSAN004222	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03253336	Middle East Technical University Food Engineering Department	7/18/12	Turkey	environmental/other	chicken meat	PD000031430.1	GCA_004192975.1	SRK1745554	PD000047098.3	CFSAN004222	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03253337	Middle East Technical University Food Engineering Department	11/20/12	Turkey	environmental/other	chicken meat	PD000031430.1	GCA_004192975.1	SRK1745554	PD000047098.3	CFSAN004222	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03253338	Middle East Technical University Food Engineering Department	12/17/12	Turkey	environmental/other	chicken meat	PD000031430.1	GCA_004192975.1	SRK1745554	PD000047098.3	CFSAN004222	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03253339	Middle East Technical University Food Engineering Department	12/19/12	Turkey	environmental/other	chicken meat	PD000031430.1	GCA_004192975.1	SRK1745554	PD000047098.3	CFSAN004222	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03272225	USDA-FSIS	2008	USA-PA	environmental/other	egg yolk	PD000031430.1	GCA_004192975.1	SRK1745554	PD000047098.3	CFSAN004222	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03291525	United States Food and Drug Administration	2008	USA	environmental/other	chicken breast	PD0000303847	GCA_008756581.1	SRK1822308	PD000050328.3	WHPH_SAL_400956	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03320201	USDA-FSIS	54/03	Colombia	environmental/other	potato and meat bouillon	PD0000744853	GCA_008910301.1	SRK3806562	PD000013434.3	CFSAN006522	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03463818	USDA-FSIS	7/25/11	USA-VA	environmental/other	beef	PD0000338988.1	GCA_002683781.1	PD0000278383.1	PD0000278383.1	CFSAN007350	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03464908	FDA Contracted Laboratory	5/0/14	USA-AR	environmental/other	chicken	PD000031292.1	GCA_008842381.1	SRK2035396	PD000065132.2	CFSAN031263	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465084	PHE	2014-10	United Kingdom: North of England	clinical	human	PD0000324261.3	GCA_011091651.1	SRK1897892	PD000052975.2	87602	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465051	PHE	2014-09	United Kingdom: London	clinical	human	PD000032435.28	GCA_011483841.1	SRK1988151	PD000053132.2	46157	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465051	PHE	2014-08	United Kingdom: London	clinical	human	PD000044228.28	GCA_011627251.1	SRK1988329	PD000053302.2	46087	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465051	PHE	2014-07	United Kingdom: North of England	clinical	human	PD000032447.4	GCA_011487051.1	SRK1984465	PD000053465.2	46621	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465051	PHE	2014-10	United Kingdom: Midlands and East of England	clinical	human	PD000032475.18	GCA_011696151.1	SRK1988508	PD000053475.2	46928	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465051	PHE	2014-10	United Kingdom: North of England	clinical	human	PD000032484.1	GCA_010876281.1	SRK1988622	PD000053553.1	46103	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465053	PHE	2014-09	United Kingdom: South of England	clinical	human	PD0000303553.63	GCA_011095051.1	SRK1988650	PD000053623.2	49717	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465054	PHE	2014-11	United Kingdom: South of England	clinical	human	PD000032475.18	GCA_011696151.1	SRK1988508	PD000053475.2	46928	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465055	PHE	2014-10	United Kingdom: Midlands and East of England	clinical	human	PD000032484.1	GCA_010876281.1	SRK1988622	PD000053553.1	46103	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465053	PHE	2014-12	United Kingdom: South of England	clinical	human	PD000032484.1	GCA_010876281.1	SRK1988622	PD000053553.1	46103	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465056	PHE	2014-12	United Kingdom: Midlands and East of England	clinical	human	PD000032484.1	GCA_010876281.1	SRK1988622	PD000053553.1	46103	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03465056	PHE	2014-01	United Kingdom: Midlands and East of England	clinical	human	PD000032484.1	GCA_010876281.1	SRK1988622	PD000053553.1	46103	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03474046	PHE	2014-08	United Kingdom: North of England	clinical	human	PD000032440.52	GCA_008950991.1	SRK1985152	PD000054067.2	38491	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03474046	PHE	2014-02	United Kingdom: North of England	clinical	human	PD000032440.52	GCA_008950991.1	SRK1985152	PD000054067.2	38491	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03475003	PHE	2014-10	United Kingdom: London	clinical	human	PD0000303116.2	GCA_008534351.1	SRK1985358	PD000055065.2	63715	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03475623	PHE	2014-09	United Kingdom: North of England	clinical	human	PD0000303939.13	GCA_008932581.1	SRK1985502	PD000055067.2	61289	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03483004	PHE	2014-03	United Kingdom: Midlands and East of England	clinical	human	PD0000303939.13	GCA_008932581.1	SRK1985502	PD000055067.2	61289	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478905	PHE	2014-08	United Kingdom: Midlands and East of England	clinical	human	PD000032863.1	GCA_008674641.1	SRK1986827	PD000056181.2	40237	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478985	PHE	2014-05	United Kingdom: South of England	clinical	human	PD000032455.28	GCA_008112901.1	SRK1986907	PD000056947.2	47384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03477234	PHE	2014-07	United Kingdom: North of England	clinical	human	PD000032842.1	GCA_008674741.1	SRK1987444	PD00005722.2	36472	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03477356	PHE	2014-08	United Kingdom: South of England	clinical	human	PD000032825.11	GCA_008982601.1	SRK1987268	PD000057262.2	38933	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478997	PHE	2014-09	United Kingdom: South of England	clinical	human	PD000032897	GCA_008701521.1	SRK1988009	PD000057562.2	38498	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478185	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478213	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478213	PHE	2014-08	United Kingdom: North of England	clinical	human	PD0000303939.13	GCA_008932581.1	SRK1985502	PD000055067.2	61289	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478213	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478213	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-05	United Kingdom: North of England	clinical	human	PD000031399.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	sanatige-inhibits
SAMN03478931	PHE	2014-07	United Kingdom: London	clinical	human	PD000032844.1	GCA_008106251.1	SRK1988009	PD000058113.2	37384	anigen	komauar7.1.1.5	

SAMN0647493		2017-01	USA	clinical	stool	PD5000153317.2	GCA 008179886.1	SRRS396349	PD1000189831.2	PNUSAS005939	anigen, formatum?+1.5, serotype+infantis
SAMN0658989	CDC	2017-02	USA	clinical		PD5000052604.1	GCA 006486955.1	SRRS399993	PD1000189831.2	PNUSAS100533	anigen, formatum?+1.5, serotype+infantis
SAMN0649871	CDC	2017-03	USA	clinical		PD5000020620.15	GCA 00584815.3	SRRS400015	PD1000189831.2	PNUSAS014638	anigen, formatum?+1.5, serotype+infantis
SAMN0659336	CDC	2017-03	USA	clinical	stool	PD5000020150.5	GCA 009194395.1	SRRS434463	PD1000201084.2	PNUSAS011316	anigen, formatum?+1.5, serotype+infantis
SAMN0648177	CDC	2017-03	USA	clinical		PD5000032431.2	GCA 006762405.1	SRRS501151	PD1000209811.2	PNUSAS12037	anigen, formatum?+1.5, serotype+infantis
SAMN0653446		2017-03	USA	clinical		PD5000021746.1	GCA 006486955.1	SRRS43002	PD1000211503.2		anigen, formatum?+1.5, serotype+infantis
SAMN0629009		2017-03	USA	clinical		PD5000029301.8	GCA 006759145.1	SRRS454400	PD1000211495.2	PNUSAS13325	anigen, formatum?+1.5, serotype+infantis
SAMN07169375	United States Food and Drug Administration	2017-04	USA	environmental/bother	pig stool	PD5000021721.1	GCA 010417385.1	SRRS454400	PD1000211495.2	CFASAS02400	anigen, formatum?+1.5, serotype+infantis
SAMN07152380	PHE	2017-05	United Kingdom, United Kingdom	clinical	human	PD5000011732.1	GCA 0059247665.1	SRRS533333	PD1000211783.2	345026	anigen, formatum?+1.5, serotype+infantis
SAMN07145338	PHE	2017-01	United Kingdom, United Kingdom	clinical	human	PD5000028853.1	GCA 009148405.1	SRRS453800	PD1000211864.2	335317	anigen, formatum?+1.5, serotype+infantis
SAMN07145365	PHE	2017-02	United Kingdom, United Kingdom	clinical	human	PD5000030323.2	GCA 007233951.1	SRRS43312	PD1000211893.2	348708	anigen, formatum?+1.5, serotype+infantis
SAMN07155558	PHE	2017-03	United Kingdom, United Kingdom	clinical	human	PD5000016766.28	GCA 009249185.1	SRRS540405	PD1000211904.2	352824	anigen, formatum?+1.5, serotype+infantis
SAMN07155951	PHE	2017-01	United Kingdom, United Kingdom	clinical	human	PD5000032467.1	GCA 009283005.1	SRRS458208	PD1000212704.2	345029	anigen, formatum?+1.5, serotype+infantis
SAMN0716329	USDA-FSIS	2017-04	USA	environmental/bother	Product-Raw intact-Pork	PD5000020078.1	GCA 006090715.1	SRRS480621	PD1000218390.2	PNUSAS014844	anigen, formatum?+1.5, serotype+infantis
SAMN07180133	PHE	2017	USA,CA	environmental/bother	Product-Raw intact-Pork	PD5000033050.3	GCA 008827575.1	SRRS632691	PD1000213903.2	FSB1701443	anigen, formatum?+1.5, serotype+infantis
SAMN07180327	PHE	2017-02	United Kingdom, United Kingdom	clinical	human	PD500002884.5	GCA 009290985.1	SRRS63273	PD1000214373.2	342929	anigen, formatum?+1.5, serotype+infantis
SAMN07180751	PHE	2017-02	United Kingdom, United Kingdom	clinical	human	PD5000023488.1	GCA 009298725.1	SRRS63312	PD1000214818.2	348709	anigen, formatum?+1.5, serotype+infantis
SAMN07180843	PHE	2017-05	United Kingdom, United Kingdom	clinical	human	PD5000029145.2	GCA 008918655.1	SRRS633351	PD1000214742.2	365880	anigen, formatum?+1.5, serotype+infantis
SAMN07180939		2017-04	USA	clinical		PD5000029135.1	GCA 008814465.1	SRRS635509	PD1000214648.2	PNUSAS014844	anigen, formatum?+1.5, serotype+infantis
SAMN07188309	CDC	2017-04	USA	clinical	stool	PD5000026135.21	GCA 006545195.1	SRRS658602	PD100021248.2	PNUSAS015254	anigen, formatum?+1.5, serotype+infantis
SAMN07209958	MOI	2017-05	Thru	clinical	stool	PD5000029106.38	GCA 008485005.1	SRRS680622	PD1000218391.2	MD01_Pw62	anigen, formatum?+1.5, serotype+infantis
SAMN07209958	USDA-FSIS	2017-05	USA,MT	environmental/bother	Product-Raw intact-Pork	PD5000071666.4	GCA 008680475.1	SRRT45627	PD100033871.1	MD01_Pw62	anigen, formatum?+1.5, serotype+infantis
SAMN07219594	USDA-FSIS	2017-04	USA,MT	clinical	animal-swine-market-swine	PD5000029863.1	GCA 008552975.1	SRRS693827	PD1000218870.2	FSB1701915	anigen, formatum?+1.5, serotype+infantis
SAMN07219594	USDA-FSIS	2017-04	USA,NE	environmental/bother	Product-Raw intact-Pork	PD5000029705.8	GCA 008571405.1	SRRT59771	PD1000222670.2	PNUSAS14373	anigen, formatum?+1.5, serotype+infantis
SAMN07248689	CDC	2017-07	USA	clinical	stool	PD5000030241.1	GCA 006761445.1	SRRS729647	PD1000223084.2	PNUSAS020290	anigen, formatum?+1.5, serotype+infantis
SAMN07450553		2017-07	USA	clinical		PD5000028921.1	GCA 008488005.1	SRRS908518	PD1000220263.2	PNUSAS191777	anigen, formatum?+1.5, serotype+infantis
SAMN07492559		2017-06	USA	clinical		PD5000032451.1	GCA 006761445.1	SRRS729647	PD1000223084.2	PNUSAS020290	anigen, formatum?+1.5, serotype+infantis
SAMN07492559	CDC	2017-06	USA	clinical		PD5000010335.1	GCA 007045105.1	SRRS932401	PD100023467.2	PNUSAS020759	anigen, formatum?+1.5, serotype+infantis
SAMN07625944	USDA-FSIS	2017	USA,IA	environmental/bother	animal-swine-market-swine	PD5000071393.2	GCA 009698295.1	SRRS944770	PD1000224587.2	PNUSAS190901	anigen, formatum?+1.5, serotype+infantis
SAMN07662460	CDC	2017-08	USA	clinical		PD5000020803.12	GCA 006241875.1	SRRS918639	PD1000216239.2	PNUSAS016239	anigen, formatum?+1.5, serotype+infantis
SAMN07662460	CDC	2017-08	USA	clinical		PD5000028599.2	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07662596	CDC	2017-08	USA	clinical		PD5000028561.1	GCA 008573125.1	SRRT10782	PD1000234325.2	anigen, formatum?+1.5, serotype+infantis	
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	2017-07	USA	clinical		PD5000028452.6	GCA 008437305.1	SRRS606698	PD1000239107.2	FSB1703746	anigen, formatum?+1.5, serotype+infantis
SAMN07684350	CDC	20									

SAMN0924126	PHE	2015-12	United Kingdom: United Kingdom	clinical	human	PD500010863.3	GCA_00387055.1	SRH748024	PD7000342437.1	195856	antigen	formula?7.1.5.seropositive-infants
SAMN0924126	PHE	2015-08	United Kingdom: United Kingdom	clinical	human	PD5000307029.6	GCA_007281925.1	SRH7480297	PD7000342474.1	563631	antigen	formula?7.1.5.seropositive-infants
SAMN0924152	USDA-FSSIS	2018	United Kingdom: United Kingdom	clinical	animal-cattle-dairy-cow	PD5000010484.0	GCA_005542115.1	SRH7484586	PD7000342584.1	FSSB11811380	antigen	formula?7.1.5.seropositive-infants
SAMN0931356	Lebanese Agriculture Research Institute	2017-07	Lebanon	environmental/other	poultry	PD500003948.0	GCA_006264295.1	SRH7504356	PD7000344445.1	NC_VH0_5031	antigen	formula?7.1.5.seropositive-infants
SAMN0931349	Tropical Governmental Hospital	6/20/12	Lebanon	clinical	stool	PD500004300.2	GCA_008074265.1	SRH7504234	PD7000344418.1	NC_VH0_5031	antigen	formula?7.1.5.seropositive-infants
SAMN0943383	PHE	2018-06	United Kingdom: United Kingdom	environmental/other	food	PD5000025193.5	GCA_004241045.1	SRH7501901	PD7000344162.1	870703	antigen	formula?7.1.5.seropositive-infants
SAMN0943383	PHE	2017-11	United Kingdom: United Kingdom	clinical	food	PD500011603.7	GCA_003895571.1	SRH7511919	PD7000343309.1	492517	antigen	formula?7.1.5.seropositive-infants
SAMN0943383	PHE	2017-11	United Kingdom: United Kingdom	clinical	food	PD5000037180.0	GCA_007272955.1	SRH7516602	PD7000344779.1	425854	antigen	formula?7.1.5.seropositive-infants
SAMN0946004	CDC	2018-04	USA	clinical	human	PD5000047179.8	GCA_006774425.1	SRH7524811	PD7000353222.1	PNJUSAS140517	antigen	formula?7.1.5.seropositive-infants
SAMN0963372	PHE	2015-08	United Kingdom: United Kingdom	clinical	human	PD500010984.7	GCA_007286855.1	SRH7533360	PD7000351179.1	4814	antigen	formula?7.1.5.seropositive-infants
SAMN0963372	CDC	2017-07	USA	clinical	human	PD5000033763.3	GCA_006725851.1	SRH7533360	PD7000351179.1	PNJUSAS16727	antigen	formula?7.1.5.seropositive-infants
SAMN0976134	University for Development Studies	5/23/16	Ghana: Yamale	environmental/other	goat	PD500014466.8	GCA_004310545.1	SRH7765194	PD7000471625.1	NC6	antigen	formula?7.1.5.seropositive-infants
SAMN0992015	UNIFAR	11/10/14	Ecuador	environmental/other	chicken feed raw material	PD500003852.0	GCA_005539355.1	SRH7765194	PD700036821.1	201A-058	antigen	formula?7.1.5.seropositive-infants
SAMN0992015	PHE	2018-07	USA	clinical	human	PD500009303.0	GCA_00673748.1	SRH7738974	PD7000366368.1	PNJUSAS1409	antigen	formula?7.1.5.seropositive-infants
SAMN09981742	CDC	2018-07	USA	clinical	human	PD500006320.1	GCA_006973565.1	SRH7758840	PD7000366570.1	PNJUSAS150848	antigen	formula?7.1.5.seropositive-infants
SAMN10091109	CDC	2018-07	USA	clinical	human	PD500006434.8	GCA_006735251.1	SRH7758840	PD7000371489.1	PNJUSAS150908	antigen	formula?7.1.5.seropositive-infants
SAMN10091287	CDC	2018-08	USA	clinical	human	PD500003610.9	GCA_006554185.1	SRH7755008	PD7000376851.9	PNJUSAS151676	antigen	formula?7.1.5.seropositive-infants
SAMN10097780	PHE	2018-08	United Kingdom: United Kingdom	clinical	human	PD5000037032.2	GCA_007017725.1	SRH7841518	PD7000377580.1	591845	antigen	formula?7.1.5.seropositive-infants
SAMN10097780	PHE	2018-09	United Kingdom: United Kingdom	clinical	human	PD5000038706.8	GCA_007276151.1	SRH7875934	PD7000381485.1	600703	antigen	formula?7.1.5.seropositive-infants
SAMN10093994	PHE	2018-07	United Kingdom: United Kingdom	clinical	human	PD50000447118.1	GCA_00703665.1	SRH7878443	PD7000381515.1	568457	antigen	formula?7.1.5.seropositive-infants
SAMN10097135	USDA-FSSIS	2018	United Kingdom: United Kingdom	environmental/other	animal-serine-sow	PD500001927.66	GCA_003863135.1	SRH7883778	PD7000382185.1	FSSB11814083	antigen	formula?7.1.5.seropositive-infants
SAMN10097135	PHE	2018-09	United Kingdom: United Kingdom	clinical	human	PD5000001594.3	GCA_00724795.1	SRH7884620	PD7000383400.1	945890	antigen	formula?7.1.5.seropositive-infants
SAMN10241329	Adolfo Lutz Institute	2014	Brazil	clinical	urine	PD5000026845.91	GCA_003726545.1	SRH7884620	PD7000407491.1	34514	antigen	formula?7.1.5.seropositive-infants
SAMN10281404	Laboratorio Nacional para la Investigacion en Inocuidad Alimentaria	11/10/10	Mexico	environmental/other	river water	PD5000037789.1	GCA_006984945.1	SRH8100768	PD7000396793.1	277	antigen	formula?7.1.5.seropositive-infants
SAMN10281412	Laboratorio Nacional para la Investigacion en Inocuidad Alimentaria	11/10/10	Mexico	environmental/other	river water	PD5000037784.1	GCA_00553215.1	SRH8100768	PD7000398161.1	286	antigen	formula?7.1.5.seropositive-infants
SAMN10391360	CDC	2018-09	USA	environmental/other	animal feed	PD500007594.2	GCA_005456065.1	SRH8170277	PD7000403737.1	PNJUSAS155557	antigen	formula?7.1.5.seropositive-infants
SAMN10391360	United States Food and Drug Administration/Office of Regulatory Affairs/ New York Human and Animal Food Laboratory	2016	USA	environmental/other	animal feed	PD5000049113.18	GCA_005418315.1	SRH817241	PD7000403762.1	FOA1080114-C001-015	antigen	formula?7.1.5.seropositive-infants
SAMN1432554	CDC	2008-08	USA	clinical	human	PD5000070048.7	GCA_005419015.1	SRH8183038	PD700045791.1	PNJUSAS160838	antigen	formula?7.1.5.seropositive-infants
SAMN10438031	PHE	2018-10	United Kingdom: United Kingdom	clinical	human	PD5000051043.2	GCA_004247035.1	SRH8201444	PD7000408739.1	621510	antigen	formula?7.1.5.seropositive-infants
SAMN10533984	CDC	2018-11	United Kingdom: United Kingdom	clinical	human	PD5000040917.1	GCA_006994245.1	SRH8205360	PD7000410258.1	640009	antigen	formula?7.1.5.seropositive-infants
SAMN1059695	CDC	2018-11	USA	clinical	human	PD5000040218.1	GCA_005822905.1	SRH8318950	PD7000430291.1	PNJUSAS163888	antigen	formula?7.1.5.seropositive-infants
SAMN10719940	CDC	2018-10	United Kingdom: United Kingdom	clinical	human	PD5000042812.1	GCA_007263375.1	SRH8427262	PD7000433960.1	637674	antigen	formula?7.1.5.seropositive-infants
SAMN10719940	CDC	2018-10	United Kingdom: United Kingdom	clinical	human	PD5000040615.1	GCA_007400325.1	SRH8427262	PD7000433960.1	648176	antigen	formula?7.1.5.seropositive-infants
SAMN10719940	CDC	2018-01	USA	clinical	human	PD5000043748.2	GCA_005728355.1	SRH8484814	PD7000447500.1	PNJUSAS166620	antigen	formula?7.1.5.seropositive-infants
SAMN10811717	PHE	2018-01	United Kingdom: United Kingdom	clinical	human	PD5000103824.1	GCA_007262485.1	SRH8500603	PD7000454813.1	165243	antigen	formula?7.1.5.seropositive-infants
SAMN10813854	PHE	2018-01	United Kingdom: United Kingdom	clinical	human	PD5000059917.12	GCA_007584065.1	SRH8504758	PD7000455293.1	665308	antigen	formula?7.1.5.seropositive-infants
SAMN10823872	PHE	2018-01	United Kingdom: United Kingdom	clinical	human	PD5000042727.1	GCA_007404945.1	SRH8498778	PD7000455922.1	674576	antigen	formula?7.1.5.seropositive-infants
SAMN10841415	PHE	2018-03	United Kingdom: United Kingdom	clinical	human	PD5000078347.1	GCA_004545935.1	SRH8504758	PD7000454728.1	648505	antigen	formula?7.1.5.seropositive-infants
SAMN10843961	PHE	2018-03	United Kingdom: United Kingdom	clinical	human	PD5000085625.1	GCA_007683075.1	SRH8503770	PD7000455293.1	616866	antigen	formula?7.1.5.seropositive-infants
SAMN10867996	CDC	2018-01	United Kingdom: United Kingdom	clinical	human	PD5000055140.7	GCA_007011325.1	SRH8628011	PD7000458602.1	643019	antigen	formula?7.1.5.seropositive-infants
SAMN10867996	CDC	2018-01	United Kingdom: United Kingdom	clinical	human	PD5000051455.0	GCA_004542625.1	SRH8675486	PD7000458797.1	PNJUSAS167523	antigen	formula?7.1.5.seropositive-infants
SAMN11116725	USDA-FSSIS	2018	USA,CA	environmental/other	Product Raw Ground, Commuted or Otherwise Noninfectious Pork	PD5000138697.9	GCA_006748245.1	SRH8723164	PD7000471765.2	FSSB11812846	antigen	formula?7.1.5.seropositive-infants
SAMN11116725	USDA-FSSIS	2018-12	USA	environmental/other	food	PD5000059661.1	GCA_006944215.1	SRH8710711	PD7000486947.1	PNJUSAS169647	antigen	formula?7.1.5.seropositive-infants
SAMN11191435	United States Food and Drug Administration	2005	USA	environmental/other	food	PD5000059510.1	GCA_006884755.1	SRH1084303	PD7000500522.1	CFSAN077748	antigen	formula?7.1.5.seropositive-infants
SAMN11191435	United States Food and Drug Administration	2005	USA	environmental/other	Not Provided	PD5000055154.1	GCA_006984785.1	SRH1084541	PD7000508512.1	CFSAN077781	antigen	formula?7.1.5.seropositive-infants
SAMN11588621	USDA-FSSIS	2018	USA,NY	environmental/other	chicken carcasses	PD5000181082.3	GCA_006537685.1	SRH1084541	PD7000508512.1	FSSB11813738	antigen	formula?7.1.5.seropositive-infants
SAMN11588644	CDC	2018-03	USA	clinical	human	PD5000050914.3	GCA_005595215.1	SRH8868214	PD7000481084.1	PNJUSAS167473	antigen	formula?7.1.5.seropositive-infants
SAMN11404274	CDC	2019-01	USA	clinical	human	PD5000074034.2	GCA_005471185.1	SRH8888388	PD7000488479.1	PNJUSAS168143	antigen	formula?7.1.5.seropositive-infants
SAMN11419488	CDC	2019-03	USA	clinical	human	PD5000049423.1	GCA_005717485.1	SRH8939117	PD7000494548.1	PNJUSAS172504	antigen	formula?7.1.5.seropositive-infants
SAMN11455511	USDA-FSSIS	2019-01	USA,MO	environmental/other	Product Raw Ground, Commuted or Otherwise Noninfectious Pork	PD5000047115.2	GCA_006750565.1	SRH8924514	PD7000491520.2	FSSB11919903	antigen	formula?7.1.5.seropositive-infants
SAMN11553499	USDA-FSSIS	2019	United Kingdom: United Kingdom	clinical	human	PD5000054411.2	GCA_007255055.1	SRH9025004	PD7000492458.1	725347	antigen	formula?7.1.5.seropositive-infants
SAMN11553499	USDA-FSSIS	2019	USA,PA	environmental/other	commuted chicken	PD5000070128.18	GCA_005538345.1	SRH9005010	PD7000495157.1	FSSB11920355	antigen	formula?7.1.5.seropositive-infants
SAMN11582745	PHE	2019-04	USA	clinical	human	PD5000183371.1	GCA_005517095.1	SRH9018150	PD7000498900.1	PNJUSAS175018	antigen	formula?7.1.5.seropositive-infants
SAMN11582745	PHE	2019-05	United Kingdom: United Kingdom	clinical	human	PD5000020177.1	GCA_006913225.1	SRH9120008	PD7000500591.1	145806	antigen	formula?7.1.5.seropositive-infants
SAMN11962517	CDC	2019-04	USA	clinical	human	PD5000064242.2	GCA_005216085.1	SRH9184978	PD7000501149.1	PNJUSAS167176	antigen	formula?7.1.5.seropositive-infants
SAMN11962595	PHE	2019-05	USA	clinical	human	PD5000065423.1	GCA_005410245.1	SRH9261746	PD7000511993.1	PNJUSAS173934	antigen	formula?7.1.5.seropositive-infants
SAMN12081173	PHE	2019-05	United Kingdom: United Kingdom	clinical	human	PD5000117428.1	GCA_006403715.1	SRH9261122	PD7000519518.1	721144	antigen	formula?7.1.5.seropositive-infants
SAMN12041584	USDA-FSSIS	2019	USA,TX	environmental/other	Product Raw-Infect-Shellfishes	PD5000045533.8	GCA_006429715.1	SRH9285118	PD7000522016.1	FSSB11921837	antigen	formula?7.1.5.seropositive-infants
SAMN12097929	Hamstead Sodgum, Murdoch University	2018-06	Australia: Perth	environmental/other	Retail table eggs	PD5000069262.2	GCA_006842965.1	SRH9306922	PD7000584671.1	65434	antigen	formula?7.1.5.seropositive-infants
SAMN12137517	United Kingdom: United Kingdom	2018-06	United Kingdom: United Kingdom	clinical	human	PD5000116027.3	GCA_007057151.1	SRH9564252	PD7000595325.1	616117	antigen	formula?7.1.5.seropositive-infants
SAMN12248632	CDC	2018-06	USA	environmental/other	animal-cattle-dairy-cow	PD5000079101.1	GCA_007620775.1	SRH9667689	PD7000540533.1	PNJUSAS162878	antigen	formula?7.1.5.seropositive-infants
SAMN1245548	USDA-FSSIS	2018-06	USA	environmental/other	animal-cattle-dairy-cow	PD5000090623.3	GCA_007076065.1	SRH9667689	PD7000540533.1	PNJUSAS162878	antigen	formula?7.1.5.seropositive-infants
SAMN12458251	CDC	2017-07	USA	clinical	human	PD5000048425.1	GCA_007731460.1	SRH9622750	PD7000540533.1	PNJUSAS165452	antigen	formula?7.1.5.seropositive-infants
SAMN12414930	PHE	2018-06	USA	clinical	human	PD5000115416.2	GCA_007875085.1	SRH9660581	PD7000551293.1	PNJUSAS163448	antigen	formula?7.1.5.seropositive-infants
SAMN12417777	USDA-FSSIS	2018-06	USA	environmental/other	animal-chicken-young-chicken	PD5000078457.1	GCA_007860055.1	SRH9671962	PD7000553389.1	FSSB11922889	antigen	formula?7.1.5.seropositive-infants
SAMN12572511	PHE	2019-07	United Kingdom: United Kingdom	clinical	human	PD5000051578.3	GCA_011635395.1	SRH9668531	PD7000556131.1	786635	antigen	formula?7.1.5.seropositive-infants
SAMN12644096	CDC	2019-07	USA	clinical	human	PD5000040225.2	GCA_008194265.1	SRH1004631	PD7000574488.1	PNJUSAS169368	antigen	formula?7.1.5.seropositive-infants
SAMN12684481	USDA-FSSIS	2019	USA,GA	environmental/other	chicken carcass	PD50000184419.7	GCA_008014481.1	SRH10081004	PD7000576320.1	FSSB11902555	antigen	formula?7.1.5.seropositive-infants
SAMN12684481	USDA-FSSIS	2019-08	USA	clinical	human	PD5000050504.0	GCA_008015685.1	SRH10012564	PD7000581395.1	PNJUSAS169204	antigen	formula?7.1.5.seropositive-infants
SAMN12789400	CDC	2019-08	USA,NC	environmental/other	animal-chicken-young-chicken	PD5000059524.1	GCA_00804699.1	SRH10126564	PD7000587719.1	FSSB11924523	antigen	formula?7.1.5.seropositive-infants
SAMN12772739	CDC	2019-08	USA	clinical	human	PD5000182168.1	GCA_008084225.1	SRH10130274	PD7000587643.1	PNJUSAS169861	antigen	formula?7.1.5.seropositive-infants
SAMN12841543	PHE	2019-09	United Kingdom: United Kingdom	clinical	human	PD5000013900.1	GCA_008765315.1	SRH10198881	PD7000590609.1	614342	antigen	formula?7.1.5.seropositive-infants
SAMN12924292	PHE	2019-09	United Kingdom: United Kingdom	clinical	human	PD5000050478.2	GCA_00884225.1	SRH10236702	PD70006011965.1	812272	antigen	formula?7.1.5.seropositive-infants
SAMN12984342	CDC	2019-09	USA	clinical	human	PD5000025931.1	GCA_00892					

SAMN15739188	2020-07	United Kingdom, United Kingdom	clinical	human	PD50000687042.2	GCA_014095955.1	SRK12396163	PD000080152.2	868771	anigen, formula?7 r1.5, serotype-chiknls
SAMN15809758	2020-07	USA	clinical	human	PD500007940.1	GCA_014171115.1	SRK12435392	PD000097029.1	PNUSAS160109	anigen, formula?7 r1.5, serotype-chiknls
SAMN15812433816	2020-07	USA	clinical	human	PD500007020.1	GCA_014206413916	SRK12435392	PD000097029.1	PNUSAS160458	anigen, formula?7 r1.5, serotype-chiknls
SAMN16055883	2020-08	USA	clinical	human	PD5000128661.2	GCA_014473755.1	SRK12586207	PD000028726.1	PNUSAS165114	anigen, formula?7 r1.5, serotype-chiknls
SAMN16122018	2020	USA/VA	environmental/other	commuted beef	PD5000072087.4	GCA_014534995.1	SRK12630733	PD000083699.2	FBS2229298	anigen, formula?7 r1.5, serotype-chiknls
SAMN16212780.1	2020	USA/MA	environmental/other	commuted chicken	PD5000122790.1	GCA_014707995.1	SRK1277584	PD000081641.1	FBS2229291	anigen, formula?7 r1.5, serotype-chiknls
SAMN16337413	2020-09	United Kingdom, United Kingdom	clinical	human	PD5000108317.3	GCA_015190645.1	SRK12762225	PD000082296.1	936378	anigen, formula?7 r1.5, serotype-chiknls
SAMN16377009	2020-08	United Kingdom, United Kingdom	environmental/other	human	PD5000070142.2	GCA_015123135.1	SRK1254949.1	PD000084949.1	868472	anigen, formula?7 r1.5, serotype-chiknls
SAMN16411109	2020-09	United Kingdom, United Kingdom	clinical	human	PD5000075469.9	GCA_014910595.1	SRK12506519	PD000083487.1	937048	anigen, formula?7 r1.5, serotype-chiknls
SAMN16532588	2020	USA/VA	environmental/other	commuted beef	PD5000074153.3	GCA_014484475.1	SRK12684609	PD000086573.1	FBS12204721	anigen, formula?7 r1.5, serotype-chiknls
SAMN16539536	2014	USA	clinical	chicken feces	PD5000070986.2	GCA_015020515.1	SRK12011903	PD000087146	CFSAN107146	anigen, formula?7 r1.5, serotype-chiknls
SAMN16559750	2013	USA	environmental/other	soy glycine max	PD5000106150.5	GCA_015004955.1	SRK12803412	PD000088401.1	CFSAN107129	anigen, formula?7 r1.5, serotype-chiknls
SAMN16634374	2014	USA	environmental/other	egg snout	PD5000074994.3	GCA_015223115.1	SRK12071246	PD000087319.1	CFSAN107167	anigen, formula?7 r1.5, serotype-chiknls
SAMN17043287	2020-10	United Kingdom, United Kingdom	environmental/other	animal	PD5000079492.1	GCA_015448855.1	SRK13020972	PD000086291.1	1021746	anigen, formula?7 r1.5, serotype-chiknls
SAMN16729110	2020-11	United Kingdom, United Kingdom	environmental/other	animal	PD5000078400.1	GCA_015446285.1	SRK13020987	PD000086297.1	1021706	anigen, formula?7 r1.5, serotype-chiknls
SAMN16929636	10/2/17	Canada	clinical	stool	PD500010880.1	GCA_015898855.1	SRK1320701	PD000086224.1	PNUSAS162145	anigen, formula?7 r1.5, serotype-chiknls
SAMN16959899	2020-10	USA	clinical	stool	PD5000079390.1	GCA_015878545.1	SRK13169142	PD000086995.1	PNUSAS162742	anigen, formula?7 r1.5, serotype-chiknls
SAMN16970709	11/17/17	Canada	clinical	stool	PD5000101222.1	GCA_016021875.1	SRK13243600	PD000091249.1	PNUSAS163748	anigen, formula?7 r1.5, serotype-chiknls
SAMN17043287	2020-10	USA	clinical	stool	PD500007142.2	GCA_015907765.1	SRK13243609	PD00009164.1	PNUSAS164982	anigen, formula?7 r1.5, serotype-chiknls
SAMN17115240	2020-11	USA	clinical	chicken carcass	PD5000078434.6	GCA_016133705.1	SRK13273148	PD0000919117.1	PNUSAS165106	anigen, formula?7 r1.5, serotype-chiknls
SAMN17128054	2017	USA	environmental/other	feces	PD500078491.1	GCA_016220425.1	SRK13277022	PD0000919367.1	CFSAN107239	anigen, formula?7 r1.5, serotype-chiknls
SAMN17133957	2016	USA	clinical	feces	PD5001140572.1	GCA_016220965.1	SRK13277447	PD0000919645.1	CFSAN107229	anigen, formula?7 r1.5, serotype-chiknls
SAMN17133955	2020-12	United Kingdom, United Kingdom	clinical	human	PD5000078466.1	GCA_016438845.1	SRK13306476	PD000092509.1	1052039	anigen, formula?7 r1.5, serotype-chiknls
SAMN1717074	2017	Canada	clinical	fecal swab	PD5000079499.1	GCA_016437405.1	SRK13312499	PD000092593.1	CFSAN107261	anigen, formula?7 r1.5, serotype-chiknls
SAMN17218714	10/31/2	Canada	clinical	fecal swab	PD5000079390.1	GCA_017248855.1	SRK13491654	PD000092593.1	PNUSAS167291	anigen, formula?7 r1.5, serotype-chiknls
SAMN17383779	2020-12	USA	clinical	human	PD5000125069.1	GCA_016728905.1	SRK13483934	PD000093817.1	PNUSAS169349	anigen, formula?7 r1.5, serotype-chiknls
SAMN17505197	2020	USA/GA	environmental/other	commuted chicken	PD5000079450.1	GCA_017236455.1	SRK1310829	PD000093860.1	FBS22104459	anigen, formula?7 r1.5, serotype-chiknls
SAMN17516059	2016	Hungary	environmental/other	broiler, feces	PD5000082140.1	GCA_016843555.1	SRK13514232	PD0000972057.1	1158478	anigen, formula?7 r1.5, serotype-chiknls
SAMN17516060	2016	Hungary	environmental/other	broiler, feces	PD5000082132.1	GCA_016844455.1	SRK13514175	PD0000972005.1	1158478	anigen, formula?7 r1.5, serotype-chiknls
SAMN17516063	2016	Hungary	environmental/other	broiler, feces	PD5000082137.3	GCA_016845475.1	SRK13514142	PD000097190.1	1158478	anigen, formula?7 r1.5, serotype-chiknls
SAMN17516084	2018	Hungary	environmental/other	broiler, caecum	PD5000082124.1	GCA_016845035.1	SRK13514168	PD0000971993.1	1158478	anigen, formula?7 r1.5, serotype-chiknls
SAMN17516087	2018	Hungary	environmental/other	broiler, caecum	PD5000082138.1	GCA_016845055.1	SRK13514168	PD0000971999.1	1158478	anigen, formula?7 r1.5, serotype-chiknls
SAMN17516103	2012	Hungary	clinical	human	PD5000118028.4	GCA_016845035.1	SRK13514149	PD0000971913.1	1158478	anigen, formula?7 r1.5, serotype-chiknls
SAMN17709534	2021	USA/MA	environmental/other	commuted chicken	PD5000113165.1	GCA_017182995.1	SRK13089110	PD000094652.2	FBS22104433	anigen, formula?7 r1.5, serotype-chiknls
SAMN17724461	2021-01	USA/NC	clinical	human	PD5000141967.1	GCA_017147955.1	SRK13212393	PD0000947298.1	1982508	anigen, formula?7 r1.5, serotype-chiknls
SAMN17897949	2021	Canada	environmental/other	commuted chicken	PD5000120913.1	GCA_017032155.1	SRK13697443	PD000094831.1	FBS22137760	anigen, formula?7 r1.5, serotype-chiknls
SAMN18051634	2017	Canada	clinical	stool	PD5000091871.1	GCA_016007945.1	SRK14882919	PD0001075960.1	PNUSAS169379	anigen, formula?7 r1.5, serotype-chiknls
SAMN180216436	2021-02	United Kingdom, United Kingdom	clinical	human	PD5000091397.53	GCA_017102045.1	SRK13776438	PD000104058.1	1158478	anigen, formula?7 r1.5, serotype-chiknls
SAMN18266709	2021-02	USA	clinical	human	PD5000083266.1	GCA_017339625.1	SRK13843446	PD000083688.1	PNUSAS169477	anigen, formula?7 r1.5, serotype-chiknls
SAMN1827798	2021	USA/GA	environmental/other	Product Raw Ground, Commuted or Otherwise Nonmated Pork	PD5000088862.1	GCA_017812195.1	SRK14949409	PD000093055.1	FBS22130871	anigen, formula?7 r1.5, serotype-chiknls
SAMN18318424	2017	Australia	clinical	Product Raw Ground, Commuted or Otherwise Nonmated Pork	PD5000084313.4	GCA_017788655.1	SRK14143708	PD000098891.1	AUSMDU0000635	anigen, formula?7 r1.5, serotype-chiknls
SAMN18419433	2017	Australia	clinical	Product Raw Ground, Commuted or Otherwise Nonmated Pork	PD5000084236.6	GCA_017867965.1	SRK14143698	PD000098900.1	AUSMDU00007173	anigen, formula?7 r1.5, serotype-chiknls
SAMN18617692	2021-04	USA	clinical	human	PD5000085625.1	GCA_016199165.1	SRK14332439	PD0001016271.1	PNUSAS169415	anigen, formula?7 r1.5, serotype-chiknls
SAMN19012497	2018	Pakistan	environmental/other	stool	PD5000085537.1	GCA_016277845.1	SRK1421684	PD000101665.1	FMBL10	anigen, formula?7 r1.5, serotype-chiknls
SAMN19012499	2018	Pakistan	environmental/other	stool	PD5000101617.1	GCA_016277225.1	SRK14421782	PD0001018897.1	FMBL10	anigen, formula?7 r1.5, serotype-chiknls
SAMN1928050	2021-04	USA	clinical	human	PD5000091223.3	GCA_016482725.1	SRK14603837	PD0001013729.1	PNUSAS200647	anigen, formula?7 r1.5, serotype-chiknls
SAMN19414275	2021	USA/GA	environmental/other	raw intact chicken	PD5000098654.1	GCA_0168601755.1	SRK14862907	PD000104846.1	FBS22105230	anigen, formula?7 r1.5, serotype-chiknls
SAMN19592115	2021-05	USA	clinical	human	PD5000091383.3	GCA_016774725.1	SRK14748472	PD0001062963.1	PNUSAS204958	anigen, formula?7 r1.5, serotype-chiknls
SAMN1961298	2007-11	United Kingdom	clinical	NPHE_7	PD5000090359.2	GCA_016843675.1	SRK14774719	PD000106369.1	NPHE_7	anigen, formula?7 r1.5, serotype-chiknls
SAMN1962412	2003-09	United Kingdom	clinical	NPHE_8	PD5000091387.1	GCA_016843635.1	SRK14774721	PD000106401.1	NPHE_8	anigen, formula?7 r1.5, serotype-chiknls
SAMN1962413	2004-05	United Kingdom	clinical	NPHE_48	PD5000091390.1	GCA_016843715.1	SRK14774839	PD000106616.1	NPHE_48	anigen, formula?7 r1.5, serotype-chiknls
SAMN1962416	2004-08	United Kingdom	clinical	NPHE_59	PD5000092019.6	GCA_016841435.1	SRK14774861	PD000106650.1	NPHE_59	anigen, formula?7 r1.5, serotype-chiknls
SAMN1962424	2006-08	United Kingdom	clinical	NPHE_30	PD5000109466.4	GCA_016843175.1	SRK14774746	PD000106426.1	NPHE_30	anigen, formula?7 r1.5, serotype-chiknls
SAMN1962447	2009-04	United Kingdom	clinical	NPHE_210	PD5000091391.1	GCA_016841735.1	SRK14774832	PD000106613.1	NPHE_210	anigen, formula?7 r1.5, serotype-chiknls
SAMN19652531	2013-09	United Kingdom	clinical	NPHE_161	PD5000096922.1	GCA_016842255.1	SRK14774777	PD0001066457.1	NPHE_161	anigen, formula?7 r1.5, serotype-chiknls
SAMN19652550	2001-08	United Kingdom	clinical	NPHE_102	PD5000091402.1	GCA_016841115.1	SRK14774835	PD0001066515.1	NPHE_102	anigen, formula?7 r1.5, serotype-chiknls
SAMN19652557	2002-10	United Kingdom	clinical	NPHE_111	PD5000091398.1	GCA_016844055.1	SRK14774672	PD0001066333.1	NPHE_111	anigen, formula?7 r1.5, serotype-chiknls
SAMN1968227	2021-05	United Kingdom, United Kingdom	clinical	human	PD5000092657.1	GCA_016839195.1	SRK14782968	PD000106655.1	1194862	anigen, formula?7 r1.5, serotype-chiknls
SAMN19891286	2021-06	USA	clinical	human	PD5000120222.1	GCA_016939755.1	SRK14810345	PD000109891.1	PNUSAS206949	anigen, formula?7 r1.5, serotype-chiknls
SAMN19891305	2021-06	USA	clinical	human	PD5000094878.1	GCA_016987905.1	SRK14875141	PD0001074853.1	PNUSAS207172	anigen, formula?7 r1.5, serotype-chiknls
SAMN20304972	2021	USA/NC	environmental/other	raw intact chicken	PD5000137700.3	GCA_016289955.1	SRK15192301	PD0001092503.1	FBS22132003	anigen, formula?7 r1.5, serotype-chiknls
SAMN20305633	2021	United States Food and Drug Administration Office of Regulatory Affairs Injunct Human and Animal Food Laboratory	environmental/other	environmental swab sponge	PD5000098696.2	GCA_016544955.1	SRK15245538	PD0001096703.1	FSAN15050615068-001	anigen, formula?7 r1.5, serotype-chiknls
SAMN20521766	2021-07	USA	clinical	NPHE10	PD5000113504.2	GCA_016562115.1	SRK15324166	PD0001101721.1	PNUSAS216992	anigen, formula?7 r1.5, serotype-chiknls
SAMN20521625	2021-07	USA	clinical	NPHE10	PD5000107118.1	GCA_016599155.1	SRK1526439	PD0001018355.1	PNUSAS217051	anigen, formula?7 r1.5, serotype-chiknls
SAMN20521094	2021-07	USA	clinical	NPHE10	PD5000152465.1	GCA_016742205.1	SRK15145379	PD0001104377.1	PNUSAS217869	anigen, formula?7 r1.5, serotype-chiknls
SAMN20521098	2021-07	USA	clinical	NPHE10	PD5000097083.1	GCA_016847245.1	SRK15145833	PD0001104379.1	PNUSAS217879	anigen, formula?7 r1.5, serotype-chiknls
SAMN20603149	2021-07	USA	clinical	NPHE10	PD5000094085.9	GCA_016841105.1	SRK15301674	PD0001103815.1	PNUSAS218526	anigen, formula?7 r1.5, serotype-chiknls
SAMN20610276	2021-08	USA	clinical	NPHE10	PD5000116037.1	GCA_016892195.1	SRK15445874	PD0001110823.1	PNUSAS220397	anigen, formula?7 r1.5, serotype-chiknls
SAMN20624367	2021-07	USA	clinical	NPHE10	PD5000098415.2	GCA_016738915.1	SRK1567276	PD0001111471.1	PNUSAS222625	anigen, formula?7 r1.5, serotype-chiknls
SAMN20627738	2021	USA/NC	environmental/other	commuted chicken	PD5000091162.1	GCA_016772755.1	SRK15614615	PD0001116938.1	FBS21142011	anigen, formula?7 r1.5, serotype-chiknls
SAMN20629424	2021	USA/TX	environmental/other	commuted beef	PD5000095158.2	GCA_016773935.1	SRK15613642	PD0001116863.1	FBS21142105	anigen, formula?7 r1.5, serotype-chiknls
SAMN2161573	2021-08	USA	clinical	NPHE10	PD5000095287.1	GCA_016834305.1	SRK15680218	PD0001120754.1	PNUSAS224588	anigen, formula?7 r1.5, serotype-chiknls
SAMN2192967	2021	USA/GA	environmental/other	product raw intact beef	PD5000095653.3	GCA_016896115.1	SRK1573818	PD0001120831.1	FBS22105527	anigen, formula?7 r1.5, serotype-chiknls
SAMN21971013	2012	Nigeria	environmental/other	feces	PD5000095875.2	GCA_016931995.1	SRK15833393	PD0001127614.1	L2125_002	anigen, formula?7 r1.5, serotype-chiknls
SAMN2197399	2021-08	USA	clinical	NPHE10	PD5000188625.1	GCA_0169995125.1	SRK15886064	PD0001130188.1	PNUSAS227626	anigen, formula?7 r1.5, serotype-chiknls
SAMN21981933	2021-09	USA	clinical	NPHE10	PD5000101936.1	GCA_020115205.1	SRK16071945	PD0001130411.1	PNUSAS230495	anigen, formula?7 r1.5, serotype-chiknls
SAMN2226231	2021	USA/OK	environmental/other	Animal Chicken Young Chicken (jece)	PD5000178719.1	GCA_020717135.1	SRK16880964	PD0001166872.1	FBS21106541	anigen, formula?7 r1.5, serotype-chiknls
SAMN2244529	2021-10	USA	clinical	NPHE10	PD5000120610.2	GCA_020717165.1	SRK16884209	PD0001167113.1	PNUSAS241192	anigen, formula?7 r1.5, serotype-chiknls
SAMN23059642	2021	USA/NC	environmental/other	Animal Chicken Young Chicken (jece)						

Supplemental Table 3.6. Select metadata downloaded from NCBI for serovar Infantis isolates included in Figure S3.6B. Blue highlight indicates samples chosen for Figure 3.4A.

BioSample	Collected by	Collection date	Location	Isolation type	Isolation source	SNP cluster	Assembly	Run	Isolate	Strain	Computed types
SAMD00019623		2000	Japan/Kyushu			PD5000116034.2	GCA_024957485.1	DRR022716	PD/T001391547.1	G117	antigen, formula=7r;1.5.aerotype=infantis
SAMEA11580171	National Institute for Communicable Disease, South Africa	1/16/13	South Africa	clinical	Human Stool	PD5000077667.3	GCA_024267256.1	ERR7671404	PD/T001358639.1		antigen, formula=7r;1.5.aerotype=infantis
SAMEA6942530		1995	Germany		broiler	PD5000077667.3	GCA_016111585.1	ERR4233026	PD/T000918645.1		antigen, formula=7r;1.5.aerotype=infantis
SAMEA7540849		2018	Germany	clinical	stool	PD5000076389.2	GCA_020141375.1	ERR4832839	PD/T001137335.1		antigen, formula=7r;1.5.aerotype=infantis
SAMN02345152	United States Food and Drug Administration Office of Regulatory Affairs Atlantic Human and Animal Food Laboratory	5/13/11	USA	environmental/other	frozen calamari ring	PD5000032456.133	GCA_011633865.1	SRR1212330	PD/T000025179.3	FSE0064	antigen, formula=7r;1.5.aerotype=infantis
SAMN02698446	United States Food and Drug Administration	5/8/10	USA	environmental/other	meat bone meal	PD5000031702.5	GCA_010525935.1	SRR1503322	PD/T000034347.3	FSE0192	antigen, formula=7r;1.5.aerotype=infantis
SAMN02695112	Minnesota Department of Health	11/15/02	USA/MN	environmental/other	tissue	PD5000043391.3	GCA_011446671.1	SRR1461812	PD/T000033985.3	MDH-2014-00346	antigen, formula=7r;1.5.aerotype=infantis
SAMN02695951	Minnesota Department of Health	8/1/03	USA/MN	environmental/other	liver	PD5000032186.7	GCA_006629545.1	SRR1586570	PD/T000039706.4	MDH-2014-00525	antigen, formula=7r;1.5.aerotype=infantis
SAMN02849840	FDA Contracted Laboratory	1/6/07	USA/CO	environmental/other	feces bovine	PD5000032425.4	GCA_011452475.1	SRR1505456	PD/T000034455.3	CFSAN018572	antigen, formula=7r;1.5.aerotype=infantis
SAMN02902678	FDA Contracted Laboratory	1/6/07	USA/WA	environmental/other	bovine necropsy	PD5000031652.1	GCA_006840651.1	SRR1528522	PD/T000035947.2	CFSAN022438	antigen, formula=7r;1.5.aerotype=infantis
SAMN03169152	PHE	2012-11	United Kingdom: Midlands and East of England	clinical	human	PD5000033938.55	GCA_008020361.1	SRR11645904	PD/T000043329.5	H124700500	antigen, formula=7r;1.5.aerotype=infantis
SAMN03195677	FDA Contracted Laboratory	4/16/07	USA/SD	environmental/other	beef	PD5000032452.2	GCA_006856405.1	SRR1722878	PD/T000045990.2	CFSAN027146	antigen, formula=7r;1.5.aerotype=infantis
SAMN03465894	PHE	2014-10	United Kingdom: North of England	clinical	human	PD5000032426.13	GCA_011091665.1	SRR1957992	PD/T000052975.2	57602	antigen, formula=7r;1.5.aerotype=infantis
SAMN04054235	USDA-FSIS	2015	USA/WA	environmental/other	Animal-Swine-Porcine Swine	PD5000032458.97	GCA_008513731.1	SRR2353813	PD/T000082853.2	FSIS1503893	antigen, formula=7r;1.5.aerotype=infantis
SAMN04337197	USDA-FSIS	2015	USA/PA	environmental/other	animal-swine-sow	PD5000032459.168	GCA_008730515.1	SRR3000039	PD/T000097431.2	FSIS1503917	antigen, formula=7r;1.5.aerotype=infantis
SAMN04600407	PHE	2015-04	United Kingdom: United Kingdom	clinical	human	PD5000028362.2	GCA_008852785.1	SRR3322087	PD/T000122482.2	107287	antigen, formula=7r;1.5.aerotype=infantis
SAMN06346072		2016-11	USA	clinical		PD5000032462.11	GCA_006635965.1	SRR5278637	PD/T000189185.2	PNUSA0008690	antigen, formula=7r;1.5.aerotype=infantis
SAMN07460708		2017-07	USA	clinical		PD5000032451.1	GCA_005751645.1	SRR5292667	PD/T000232884.2	PNUSA0002390	antigen, formula=7r;1.5.aerotype=infantis
SAMN08376490	CDC	2015-03	USA	clinical		PD5000103799.1	GCA_005262915.1	SRR6480645	PD/T000278896.2	2015AM-0283	antigen, formula=7r;1.5.aerotype=infantis
SAMN08581681	USDA-FSIS	2017	USA/WI	environmental/other	Product-Raw-Ground, Comminuted or Otherwise Nonintact Pork	PD5000103802.1	GCA_010986755.1	SRR6764209	PD/T000288648.2	FSIS11706862	antigen, formula=7r;1.5.aerotype=infantis
SAMN08862862	USDA-FSIS	2018	USA/MN	environmental/other	Product-Raw-Ground, Comminuted or Otherwise Nonintact Pork	PD5000032457.9	GCA_008491865.1	SRR6959499	PD/T000304086.2	FSIS11805941	antigen, formula=7r;1.5.aerotype=infantis
SAMN09444132	USDA-FSIS	2018	USA/NE	environmental/other	animal-swine-market swine	PD5000097345.3	GCA_008447705.1	SRR7358414	PD/T000332042.1	FSIS11810649	antigen, formula=7r;1.5.aerotype=infantis
SAMN11132121	CDC	2018-12	USA	clinical		PD5000091389.1	GCA_005664275.1	SRR8731274	PD/T000478341.1	PNUSA0509447	antigen, formula=7r;1.5.aerotype=infantis
SAMN15408326	CDC	2020-06	USA	clinical		PD5000084924.5	GCA_013613795.1	SRR12121576	PD/T000774460.1	PNUSA05150624	antigen, formula=7r;1.5.aerotype=infantis
SAMN16122018	USDA-FSIS	2020	USA/WA	environmental/other	comminuted beef	GCA_014534865.1	SRR12630733	PD/T000835602.1	FSIS20200226		antigen, formula=7r;1.5.aerotype=infantis
SAMN19652416	PHE	2004-08	United Kingdom	clinical	human	PD5000074019.6	GCA_018841435.1	SRR14774821	PD/T001066501.1	NPHE_58	antigen, formula=7r;1.5.aerotype=infantis
SAMN20603143	CDC	2021-07	USA	clinical		PD5000094065.9	GCA_019481105.1	SRR15361674	PD/T001103815.1	PNUSA0218526	antigen, formula=7r;1.5.aerotype=infantis
SAMN25608883		2014	Slovenia	clinical		PD5000116014.1	GCA_024950025.1	SRR17880716	PD/T001385277.1	S105	antigen, formula=7r;1.5.aerotype=infantis
SAMN30355371		2022-08	USA	clinical		PD5000116131.2	GCA_024944465.1	SRR21120049	PD/T001392917.1	PNUSA03292578	antigen, formula=7r;1.5.aerotype=infantis
SAMN39330787				clinical		PD5000148192.2	GCA_035583875.1	SRR27478710	PD/T002050087.1	PNUSA0413508	antigen, formula=7r;1.5.aerotype=infantis

Supplemental Table 3.7. Select metadata downloaded from NCBI for serovar Typhimurium isolates included in Figure 3.5A.

BioSample	Collected by	Collection date	Location	Isolation source	Isolation type	SNP cluster	Run	Assembly	Isolate	Strain	Computed types
SAMN07424695	CDC	2017-06	USA	stool	clinical	PDS0000051476.18	SRR5888672	GCA_008884465.1	PD/T000230419.2	PNUSAS018095	antigen, formula=4:1.2,serotype=Typhimurium
SAMN08407830	CDC	2018-01	USA	stool	clinical	PDS0000076391.10	SRH6729918	GCA_016381345.1	PD/T000282937.4	PNUSAS032925	antigen, formula=4:1.2,serotype=Typhimurium
SAMN08607476	CDC	2017-09	USA		clinical	PDS0000049109.27	SRR8782607	GCA_007561945.1	PD/T000289579.2	PNUSAS033431	antigen, formula=4:1.2,serotype=Typhimurium
SAMN10241133	CDC	2018-09	USA		clinical	PDS0000038355.20	SRR8054302	GCA_016300485.1	PD/T000391315.1	PNUSAS056026	antigen, formula=4:1.2,serotype=Typhimurium
SAMN10359377	CDC	2018-10	USA		clinical	PDS0000038355.20	SRR8144689	GCA_016294885.1	PD/T000400298.1	PNUSAS059811	antigen, formula=4:1.2,serotype=Typhimurium
SAMN10395430	FDA Contracted Laboratory	2017	USA,CA	finished almond	environmental/other	PDS0000038355.20	SRR8176610	GCA_006528135.1	PD/T000405208.1	CFSAN087747	antigen, formula=4:1.2,serotype=Typhimurium
SAMN10689682	CDC	2018-12	USA		clinical	PDS0000038355.20	SRR8389118	GCA_016275845.1	PD/T000430161.1	PNUSAS064747	antigen, formula=4:1.2,serotype=Typhimurium
SAMN11102910	KY-M	7/9/18	USA,KY	Odorcoleus virginianus Feces	environmental/other	PDS0000044303.2	SRR8787784	GCA_006239385.1	PD/T000481252.1	SAL-18-VL-LA-KY-0011	antigen, formula=4:1.2,serotype=Typhimurium
SAMN11582741	CDC	2019-04	USA		clinical	PDS0000049109.27	SRR9019092	GCA_016267365.1	PD/T000498892.1	PNUSAS070512	antigen, formula=4:1.2,serotype=Typhimurium
SAMN12036221	CDC	2019-05	USA		clinical	PDS0000054358.20	SRR9283102	GCA_006425235.1	PD/T000521479.1	PNUSAS077320	antigen, formula=4:1.2,serotype=Typhimurium
SAMN12079230	CDC	2019-06	USA		clinical	PDS0000049109.27	SRR9312597	GCA_006467545.1	PD/T000524007.1	PNUSAS079150	antigen, formula=4:1.2,serotype=Typhimurium
SAMN12147598	CDC	2019-05	USA		clinical	PDS0000051476.18	SRR9610707	GCA_007615915.1	PD/T000533382.1	PNUSAS079837	antigen, formula=4:1.2,serotype=Typhimurium
SAMN12392331	CDC	2019-06	USA		clinical	PDS0000049109.27	SRR9856451	GCA_007868925.1	PD/T000550967.1	PNUSAS085644	antigen, formula=4:1.2,serotype=Typhimurium
SAMN12537850	CDC	2019-06	USA		clinical	PDS0000049109.27	SRR9932329	GCA_008202105.1	PD/T000558638.1	PNUSAS088114	antigen, formula=4:1.2,serotype=Typhimurium
SAMN12842595	CDC	2019-09	USA		clinical	PDS0000038355.20	SRR10179687	GCA_016241875.1	PD/T000593792.1	PNUSAS102443	antigen, formula=4:1.2,serotype=Typhimurium
SAMN12889186	CDC	2019-09	USA		clinical	PDS0000051476.18	SRR10215630	GCA_008825405.1	PD/T000598887.1	PNUSAS104449	antigen, formula=4:1.2,serotype=Typhimurium
SAMN12924378	CDC	2019-10	USA		clinical	PDS0000038355.20	SRR10236752	GCA_016230825.1	PD/T000601905.1	PNUSAS105693	antigen, formula=4:1.2,serotype=Typhimurium
SAMN13107897	CDC	2019-09	USA		clinical	PDS0000049109.27	SRR10340447	GCA_009375435.1	PD/T000617246.1	PNUSAS110904	antigen, formula=4:1.2,serotype=Typhimurium
SAMN13244854	CDC	2019-10	USA		clinical	PDS0000038355.20	SRR10420101	GCA_016225105.1	PD/T000625882.1	PNUSAS115298	antigen, formula=4:1.2,serotype=Typhimurium
SAMN13536205	CDC	2019-11	USA		clinical	PDS0000054358.20	SRR10664442	GCA_011182755.1	PD/T000643427.1	PNUSAS121348	antigen, formula=4:1.2,serotype=Typhimurium
SAMN14083425	CDC	2020-01	USA		clinical	PDS0000054358.20	SRR10967570	GCA_011489485.1	PD/T000682507.1	PNUSAS134090	antigen, formula=4:1.2,serotype=Typhimurium
SAMN14438399	CDC	2020-03	USA		clinical	PDS0000054358.20	SRR11401971	GCA_011742965.1	PD/T000712723.1	PNUSAS139746	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15378831	CDC	2020-06	USA		clinical	PDS0000051476.18	SRR12097834	GCA_013651915.1	PD/T000717988.1	PNUSAS149604	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15398937	CDC	2020-06	USA		clinical	PDS0000051476.18	SRR12109672	GCA_016167695.1	PD/T000773851.1	PNUSAS149967	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15419583	CDC	2020-06	USA		clinical	PDS0000051476.18	SRR12129538	GCA_016167435.1	PD/T000774770.1	PNUSAS150862	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15460852	CDC	2020-06	USA		clinical	PDS0000051476.18	SRR12160529	GCA_013522295.1	PD/T000783830.1	PNUSAS151518	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15493347	CDC	2020-06	USA		clinical	PDS0000051476.18	SRR12185040	GCA_013494565.1	PD/T000784644.1	PNUSAS151488	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15517233	CDC	2020-06	USA		clinical	PDS0000049109.27	SRR12207372	GCA_013526655.1	PD/T000787128.1	PNUSAS152531	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15575936	CDC	2020-07	USA		clinical	PDS0000038355.20	SRR12263105	GCA_013545025.1	PD/T000790069.1	PNUSAS153944	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15775940	CDC	2020-07	USA		clinical	PDS0000049109.27	SRR12423984	GCA_016160075.1	PD/T000806632.1	PNUSAS159350	antigen, formula=4:1.2,serotype=Typhimurium
SAMN15863143	CDC	2020-08	USA		clinical	PDS0000051476.18	SRR12481490	GCA_014257065.1	PD/T000817064.1	PNUSAS161843	antigen, formula=4:1.2,serotype=Typhimurium
SAMN16124742	CDC	2020-08	USA		clinical	PDS0000038355.20	SRR12633952	GCA_016158575.1	PD/T000835764.1	PNUSAS166591	antigen, formula=4:1.2,serotype=Typhimurium
SAMN16134730	CDC	2020-08	USA		clinical	PDS0000054358.20	SRR12643153	GCA_014604815.1	PD/T000836650.1	PNUSAS167241	antigen, formula=4:1.2,serotype=Typhimurium
SAMN16451839	CDC	2020-09	USA		clinical	PDS0000051476.18	SRR12827167	GCA_016154295.1	PD/T000862761.1	PNUSAS175260	antigen, formula=4:1.2,serotype=Typhimurium
SAMN16452222	CDC	2020-08	USA		clinical	PDS0000054358.20	SRR12826943	GCA_014953285.1	PD/T000862571.1	PNUSAS173666	antigen, formula=4:1.2,serotype=Typhimurium
SAMN16522471	CDC	2020-09	USA		clinical	PDS0000049109.27	SRR12883077	GCA_014927845.1	PD/T000865377.1	PNUSAS176648	antigen, formula=4:1.2,serotype=Typhimurium
SAMN17138041	CDC	2020-11	USA		clinical	PDS0000054358.20	SRR13289723	GCA_016308605.1	PD/T000921969.1	PNUSAS186397	antigen, formula=4:1.2,serotype=Typhimurium
SAMN17251969	CDC	2020-12	USA		clinical	PDS0000076391.10	SRR13380180	GCA_016589225.1	PD/T000929858.1	PNUSAS187431	antigen, formula=4:1.2,serotype=Typhimurium
SAMN17394085	CDC	2020-12	USA		clinical	PDS0000049109.27	SRR13495192	GCA_017249195.1	PD/T000939029.1	PNUSAS189168	antigen, formula=4:1.2,serotype=Typhimurium
SAMN18743077	PHAC	2018	Canada	stool	clinical	PDS0000049109.27	SRR14280642	GCA_018116045.1	PD/T001009683.1	PNUSAS200621	antigen, formula=4:1.2,serotype=Typhimurium
SAMN19107277	CDC	2021-04	USA		clinical	PDS0000049109.27	SRR14484720	GCA_018352055.1	PD/T001024284.1	PNUSAS201133	antigen, formula=4:1.2,serotype=Typhimurium
SAMN19974806	CDC	2021-06	USA		clinical	PDS0000192391.1	SRR15006564	GCA_019087325.1	PD/T001080017.1	PNUSAS209577	antigen, formula=4:1.2,serotype=Typhimurium
SAMN20090106	CDC	2021-06	USA		clinical	PDS0000093278.1	SRR15058485	GCA_019162375.1	PD/T001083896.1	1019624001	antigen, formula=4:1.2,serotype=Typhimurium
SAMN20306031	CDC	2021-06	USA		clinical	PDS0000093278.1	SRR15192434	GCA_019290995.1	PD/T001092596.1	1021644001	antigen, formula=4:1.2,serotype=Typhimurium
SAMN20677178	CDC	2021-07	USA		clinical	PDS0000076391.10	SRR15377064	GCA_019498445.1	PD/T001106150.1	PNUSAS219220	antigen, formula=4:1.2,serotype=Typhimurium
SAMN20873945	CDC	2021-07	USA		clinical	PDS0000051476.18	SRR15540741	GCA_019707935.1	PD/T001113791.1	PNUSAS221923	antigen, formula=4:1.2,serotype=Typhimurium
SAMN21442709	CDC	2021-08	USA		clinical	PDS0000051476.18	SRR15904945	GCA_020010025.1	PD/T001130701.1	PNUSAS228169	antigen, formula=4:1.2,serotype=Typhimurium
SAMN21570178	CDC	2021-08	USA		clinical	PDS0000054358.20	SRR16023372	GCA_020121535.1	PD/T001136369.1	PNUSAS230456	antigen, formula=4:1.2,serotype=Typhimurium
SAMN22575512	CDC	2021-10	USA		clinical	PDS0000051476.18	SRR16578854	GCA_020649135.1	PD/T001162192.1	PNUSAS239477	antigen, formula=4:1.2,serotype=Typhimurium
SAMN27562167	CDC	2022-03	USA		clinical	PDS0000054358.20	SRR18741443	GCA_022982875.1	PD/T001286699.1	PNUSAS265900	antigen, formula=4:1.2,serotype=Typhimurium
SAMN29058726	CDC	2022-05	USA		clinical	PDS0000049109.27	SRR19660945	GCA_023735575.1	PD/T001329142.1	PNUSAS277091	antigen, formula=4:1.2,serotype=Typhimurium
SAMN29472493	CDC	2022-06	USA		clinical	PDS0000049109.27	SRR19918648	GCA_024123825.1	PD/T001351043.1	PNUSAS279926	antigen, formula=4:1.2,serotype=Typhimurium
SAMN29554865	CDC	2022-06	USA		clinical	PDS0000054358.20	SRR20015298	GCA_024189215.1	PD/T001354067.1	PNUSAS281387	antigen, formula=4:1.2,serotype=Typhimurium
SAMN30110362	CDC	2022-07	USA		clinical	PDS0000076391.10	SRR20773590	GCA_024568965.1	PD/T001377903.1	PNUSAS288206	antigen, formula=4:1.2,serotype=Typhimurium
SAMN30120894	CDC	2022-07	USA		clinical	PDS0000051476.18	SRR20776032	GCA_024570545.1	PD/T001377943.1	PNUSAS288117	antigen, formula=4:1.2,serotype=Typhimurium
SAMN30282664	CDC	2022-06	USA		clinical	PDS0000049109.27	SRR21025949	GCA_025020505.1	PD/T001386943.1	PNUSAS279148	antigen, formula=4:1.2,serotype=Typhimurium
SAMN30473555	CDC	2022-07	USA		clinical	PDS0000054358.20	SRR21203170	GCA_024863145.1	PD/T001398211.1	PNUSAS294577	antigen, formula=4:1.2,serotype=Typhimurium
SAMN30806977	CDC	2022-08	USA		clinical	PDS0000049109.27	SRR21530791	GCA_025184945.1	PD/T001414556.1	PNUSAS298977	antigen, formula=4:1.2,serotype=Typhimurium
SAMN30925995	CDC	2022-08	USA		clinical	PDS0000049109.27	SRR21627096	GCA_025301155.1	PD/T001424607.1	PNUSAS300557	antigen, formula=4:1.2,serotype=Typhimurium
SAMN31022944	CDC	2022-09	USA		clinical	PDS0000121793.1	SRR21712787	GCA_025418035.1	PD/T001434404.1	PNUSAS303535	antigen, formula=4:1.2,serotype=Typhimurium
SAMN31136732	CDC	2022-09	USA		clinical	PDS0000049109.27	SRR21786390	GCA_025535175.1	PD/T001440408.1	PNUSAS305274	antigen, formula=4:1.2,serotype=Typhimurium
SAMN31164279	CDC	2022-09	USA		clinical	PDS0000051476.18	SRR21814042	GCA_025589815.1	PD/T001443006.1	PNUSAS305912	antigen, formula=4:1.2,serotype=Typhimurium
SAMN31439003	CDC	2022-09	USA		clinical	PDS0000121793.1	SRR22031776	GCA_025858515.1	PD/T001465735.1	PNUSAS312148	antigen, formula=4:1.2,serotype=Typhimurium
SAMN31670458	CDC	2022-10	USA		clinical	PDS0000049109.27	SRR22245970	GCA_026111315.1	PD/T001482596.1	PNUSAS315606	antigen, formula=4:1.2,serotype=Typhimurium
SAMN31670460	CDC	2022-10	USA		clinical	PDS0000049109.27	SRR22245971	GCA_026110595.1	PD/T001482597.1	PNUSAS315614	antigen, formula=4:1.2,serotype=Typhimurium
SAMN32275466	Washington State University Washington Animal Disease Diagnostic Laboratory	1/6/22	USA,OH	feces (Dromaius novaehollandiae)	environmental/other	PDS0000038355.20	SRR22696161	GCA_026962955.1	PD/T001529903.1	SAL-22-VL-OH-WA-0003	antigen, formula=4:1.2,serotype=Typhimurium
SAMN32421168	CDC	2022-12	USA		clinical	PDS0000051476.18	SRR22901383	GCA_027470025.1	PD/T001549179.1	PNUSAS324236	antigen, formula=4:1.2,serotype=Typhimurium

SAMN32795229	CDC	2022-12	USA	clinical	PDS000049109.27	SRR23121872	GCA_028043485.1	PDT001581797.1	PNUSAS326947	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN33554138	CDC	2022	USA	clinical	PDS000054358.20	SRR23684026	GCA_028917265.1	PDT001649752.1	PNUSAS335255	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN34378121		2023	USA	clinical	PDS000049109.27	SRR24309521	GCA_030220445.1	PDT001707227.1	PNUSAS344545	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN34381829		2023	USA	clinical	PDS000154543.2	SRR24313502	GCA_032269965.1	PDT001708295.1	PNUSAS344096	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN35026792		2023	USA	clinical	PDS000051476.18	SRR24495916	GCA_033667145.1	PDT001728862.1	PNUSAS347886	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN35723225		2023	USA	clinical	PDS000149796.1	SRR24907085	GCA_032883215.1	PDT001778228.1	PNUSAS355697	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN35995608		2023	USA	clinical	PDS000049109.27	SRR25033789	GCA_032687805.1	PDT001791117.1	PNUSAS358588	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN35995619		2023	USA	clinical	PDS000150767.2	SRR25034504	GCA_032687305.1	PDT001791146.1	PNUSAS358607	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN36306403		2023	USA	clinical	PDS000154543.2	SRR25146543	GCA_032177825.1	PDT001809638.1	PNUSAS361102	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN36701919		2023	USA	clinical	PDS000051476.18	SRR25411884	GCA_032369635.1	PDT001827659.1	PNUSAS366057	antigen_formula=4:1:1.2,serotype=Typhimurium
SAMN36770391	CDC	2023	USA	clinical	PDS000154543.2	SRR25462192	GCA_032043915.1	PDT001835110.1	PNUSAS367477	antigen_formula=4:1:1.2,serotype=Typhimurium

Supplemental Table 3.8. Select metadata downloaded from NCBI for serovar Typhimurium isolates included in Figure S3.7A.

Yellow highlight indicates samples chosen for Figure S3.7B.

BioSample	Collected by	Collection date	Location	Isolation type	Isolation source	SNP cluster	Run	Assembly
SAMD00097420		1992	Japan			PD5000076400.2	DRR106839	GCA_015650885.1
SAMD00097438		2012	Italy			PD5000098671.1	DRR106884	GCA_015650845.1
SAMD00097443		1999	Japan			PD5000050000.5	DRR106894	GCA_015651185.1
SAMD00097448		2006	Japan			PD5000139929.1	DRR106906	GCA_015651245.1
SAMD00097451		1980	Japan			PD5000076402.2	DRR106912	GCA_015649895.1
SAMD00097475		2013	Japan			PD5000076396.1	DRR106973	GCA_015650325.1
SAMEA1009558						PD5000041675.17	ERR044712	GCA_010121665.1
SAMEA1009580						PD5000049928.2	ERR044703	GCA_010122525.1
SAMEA1009599						PD5000041640.3	ERR044702	GCA_010121785.1
SAMEA1009601						PD5000042445.1	ERR044713	GCA_010121745.1
SAMEA1009630						PD5000049927.1	ERR044714	GCA_010122565.1
SAMEA1009638						PD5000077449.1	ERR044730	GCA_010121805.1
SAMEA1009664						PD5000042265.3	ERR044708	GCA_010121965.1
SAMEA1009684						PD5000065265.3	ERR044707	GCA_010122035.1
SAMEA1009689						PD5000077253.1	ERR044709	GCA_010122005.1
SAMEA1009700						PD5000040809.8	ERR044706	GCA_010122125.1
SAMEA1009744						PD5000201471.8	ERR044711	GCA_010122245.1
SAMEA1009747						PD5000065255.1	ERR044704	GCA_010122235.1
SAMEA1009750						PD5000077457.3	ERR044721	GCA_010122225.1
SAMEA1009766						PD5000077252.1	ERR044732	GCA_010132915.1
SAMEA1009801						PD5000041683.1	ERR044735	GCA_010121525.1
SAMEA1009853						PD5000065275.1	ERR044700	GCA_010122085.1
SAMEA1009867						PD5000041671.1	ERR044736	GCA_010122505.1
SAMEA1009868						PD5000065150.32	ERR044718	GCA_010132935.1
SAMEA1009873						PD5000002727.19	ERR044740	GCA_010122205.1
SAMEA1009891						PD5000042451.2	ERR044720	GCA_010122325.1
SAMEA104151237	DTU	2014	Denmark			PD5000027880.2	ERR2023545	GCA_010125785.1
SAMEA104157528		2017	Ireland	clinical	human	PD5000117337.1	ERR2030254	GCA_025077515.1
SAMEA104286012		2017	Ireland	clinical	human	PD5000117331.1	ERR2114795	GCA_025077375.1
SAMEA104398339		2010	Germany	animal	animal	PD5000074136.5	ERR2202500	GCA_010132905.1
SAMEA104398348		2013	Germany	animal	animal	PD5000077231.2	ERR2202509	GCA_011424785.1
SAMEA104398349		2015	Germany	animal	animal	PD5000027706.1	ERR2202510	GCA_011128535.1
SAMEA104398350		2015	Germany	animal	animal	PD5000027735.1	ERR2202511	GCA_010458305.1
SAMEA104398352		2015	Germany	animal	animal	PD5000027713.2	ERR2202513	GCA_011401635.1
SAMEA104398354		2015	Germany	food	food	PD5000026816.11	ERR2202515	GCA_010614345.1
SAMEA104398364		2016	Germany	animal	animal	PD5000027720.1	ERR2202525	GCA_010154675.1
SAMEA104411002	APHA	2015	United Kingdom			PD5000074192.3	ERR2208744	GCA_010189705.1
SAMEA104413273	NIPH-NH	10/21/15	Poland			PD5000091078.1	ERR2210574	GCA_018337995.1
SAMEA11504959	Quelimane Central Hospital - CISM	3/12/20	Mozambique	clinical	puncture	PD50000125065.2	ERR10438741	GCA_026328495.1
SAMEA114302261		2010	China		2g	PD5000158216.2	ERR12019122	GCA_031330115.1
SAMEA114526782	A. Smith, National Institute for Communicable Diseases	6/8/23	South Africa	clinical	Human, Other	PD5000100611.11	ERR12142651	GCA_033124025.1
SAMEA114526790	A. Smith, National Institute for Communicable Diseases	5/24/23	South Africa	clinical	Human, Blood culture	PD5000109287.4	ERR12142659	GCA_033124005.1
SAMEA14288404		2008	France	clinical	human	PD5000110445.1		GCA_940677165.1
SAMEA1483990		2011	United Kingdom: Scotland	clinical	not known	PD5000026666.20		GCA_001217585.1
SAMEA1484011		1995	United Kingdom: Scotland	clinical	not known	PD5000026653.1		GCA_001216205.1
SAMEA1484114		1997	United Kingdom: Scotland	clinical	not known	PD5000026676.1		GCA_001222005.1
SAMEA1568486		2001	Canada	clinical	not known	PD5000026659.1		GCA_001116925.1
SAMEA1568512		2001	Canada	clinical	not known	PD5000100240.3		GCA_001096565.1
SAMEA1568522		2001	Canada	clinical	not known	PD5000026663.1		GCA_001161925.1
SAMEA1568541		2001	Canada	clinical	not known	PD5000026689.14		GCA_001127885.1
SAMEA1711387						PD5000009152.4	ERR230403	GCA_010122605.1
SAMEA1711417						PD5000013660.9	ERR230398	GCA_010633465.1
SAMEA1711464						PD5000037123.1	ERR230434	GCA_010122585.1
SAMEA20258668						PD5000041635.1	ERR1891220	GCA_010211645.1
SAMEA20322418						PD5000041685.1	ERR1891524	GCA_010228015.1
SAMEA20328418						PD5000041688.4	ERR1891525	GCA_010230865.1
SAMEA20331418						PD5000041660.1	ERR1891557	GCA_010229205.1
SAMEA20332168						PD5000041641.1	ERR1891561	GCA_010227895.1
SAMEA20337418						PD5000041669.1	ERR1891577	GCA_010229565.1
SAMEA2272227						PD5000065258.1		GCA_000493535.2
SAMEA2272777		2004	Malawi:Blantyre	clinical	blood	PD5000041999.125		GCA_000027025.1
SAMEA2493785	Statens Serum Institut (SSI)	8/23/05	Denmark	clinical		PD5000013655.11	ERR493461	GCA_010121845.1
SAMEA2625959						PD5000042042.1	ERR702724	GCA_010229995.1
SAMEA2625965						PD5000042322.2	ERR702798	GCA_010231575.1
SAMEA2626036						PD5000042057.1	ERR702896	GCA_010231755.1
SAMEA2626065						PD5000097742.1	ERR702837	GCA_010232485.1
SAMEA26869168	DTU	2014	Switzerland			PD5000064232.1	ERR1759212	GCA_017288335.1
SAMEA3269008	DANMAP	1/1/11	Denmark			PD5000026635.1	ERR769345	GCA_010123085.1
SAMEA3493114		2013	Denmark	clinical	human	PD5000042111.1	ERR1556162	GCA_010246135.1
SAMEA349328		2013	Denmark	clinical	human	PD5000050005.2	ERR1540273	GCA_010184185.1
SAMEA349363		2013	Denmark	clinical	human	PD5000027599.6	ERR1540287	GCA_010193865.1
SAMEA349405		2013	Denmark	clinical	human	PD5000042397.12	ERR1540322	GCA_010194905.1
SAMEA349416		2013	Denmark	clinical	human	PD5000040888.6	ERR1540364	GCA_010195905.1
SAMEA349465		2014	Denmark	clinical	human	PD5000050042.4	ERR1540375	GCA_010196085.1
SAMEA349577		2014	Denmark	clinical	human	PD5000042254.4	ERR1540424	GCA_010198385.1
SAMEA349592		2014	Denmark	clinical	human	PD5000050016.3	ERR1540536	GCA_010199865.1
		2014	Denmark	clinical	human	PD5000027609.15	ERR1540551	GCA_010208885.1

SAMEA4390586	DTU	2/8/13	Denmark			PD5000042356.3	ERR1592513	GCA_022473075.1
SAMEA4390587	DTU	3/26/13	Denmark			PD5000105942.1	ERR1592526	GCA_022469875.1
SAMEA4390611	DTU	4/3/13	Denmark			PD5000026822.8	ERR1592530	GCA_022471475.1
SAMEA4390615	DTU	8/14/12	Denmark			PD5000105941.1	ERR1592584	GCA_022470115.1
SAMEA4390619	DTU	8/29/12	Denmark			PD5000042268.3	ERR1592389	GCA_02483425.1
SAMEA4390635	DTU	1/30/12	Germany			PD5000105937.1	ERR1592410	GCA_022473615.1
SAMEA4390644	DTU	12/18/12	Germany			PD5000105938.1	ERR1592418	GCA_022473485.1
SAMEA4390697	DTU	10/31/13	Denmark			PD5000045973.3	ERR1592488	GCA_024826695.1
SAMEA4625254	NIPH-NH	9/14/16	Poland			PD5000088362.1	ERR2543061	GCA_010183345.1
SAMEA4715861		2016	Sweden	clinical	human	PD5000059956.3	ERR2617858	GCA_015878415.1
SAMEA5038452						PD5000065781.1	ERR3864679	GCA_011304605.1
SAMEA5038517						PD5000056421.1	ERR3864617	GCA_011242855.1
SAMEA5039443						PD5000056263.7	ERR3843480	GCA_011223795.1
SAMEA5130710						PD5000107522.2	ERR3967698	GCA_022223645.1
SAMEA5130713						PD5000104835.1	ERR3967664	GCA_022224325.1
SAMEA5130716						PD5000104923.1	ERR3967703	GCA_022223565.1
SAMEA53008168		2016	Ireland	clinical	human	PD5000175787.2	ERR1802430	GCA_025086485.1
SAMEA5423339						PD5000065783.1	ERR4303224	GCA_013536495.1
SAMEA5423396						PD5000099889.1	ERR4303403	GCA_013540495.1
SAMEA5556597						PD5000064969.1	ERR4289448	GCA_013642515.1
SAMEA5557066						PD5000065013.1	ERR4289759	GCA_013636475.1
SAMEA5671027						PD5000042410.3	ERR4338430	GCA_013596545.1
SAMEA5671056						PD5000065622.3	ERR4338542	GCA_013595835.1
SAMEA5671060						PD5000065599.1	ERR4338603	GCA_013605045.1
SAMEA5671100						PD5000042352.4	ERR4338355	GCA_013594035.1
SAMEA5671114						PD5000042403.3	ERR4338395	GCA_013530455.1
SAMEA5671316						PD5000042316.3	ERR4338424	GCA_013595625.1
SAMEA5671330						PD5000065788.1	ERR4338487	GCA_013530195.1
SAMEA5671356						PD5000065777.2	ERR4338581	GCA_013529135.1
SAMEA5671371						PD5000065795.1	ERR4338639	GCA_013529615.1
SAMEA5671393						PD5000065620.1	ERR4338728	GCA_013599915.1
SAMEA5671396						PD5000107959.1	ERR4338737	GCA_013487385.1
SAMEA5671408						PD5000042317.2	ERR4338774	GCA_013562855.1
SAMEA5671413						PD5000065995.4	ERR4338789	GCA_013562725.1
SAMEA5671416						PD5000065770.1	ERR4338804	GCA_013486525.1
SAMEA5671430						PD5000077419.3	ERR4338837	GCA_013486165.1
SAMEA5671433						PD5000065780.1	ERR4338846	GCA_013558625.1
SAMEA5671450						PD5000065774.1	ERR4338898	GCA_013562415.1
SAMEA5671454						PD5000065608.1	ERR4338910	GCA_013558415.1
SAMEA5671468						PD5000107923.1	ERR4338948	GCA_013561535.1
SAMEA5671473						PD5000065623.1	ERR4338961	GCA_013561275.1
SAMEA5671480						PD5000107922.1	ERR4338683	GCA_013600865.1
SAMEA5882646		2018	Germany		livestock	PD5000051888.2	ERR3489631	GCA_010264655.1
SAMEA6057492		2018	Germany		environment	PD5000051919.5	ERR3580827	GCA_009634055.1
SAMEA6057531		2019	Germany	animal		PD5000051917.1	ERR3580866	GCA_009634615.1
SAMEA6057847		2016	Germany	animal		PD5000051887.1	ERR3581182	GCA_009636795.1
SAMEA6057852		2017	Germany	animal		PD5000073725.1	ERR3581187	GCA_015044595.1
SAMEA6057960		2018	Germany	animal		PD5000051906.3	ERR3581295	GCA_009637655.1
SAMEA6058385		2018	Germany	animal		PD5000179743.1	ERR3581719	GCA_009641975.1
SAMEA6058528		2019	Germany	animal		PD5000153462.1	ERR3581862	GCA_009643575.1
SAMEA6514112						PD5000042504.4	ERR3901218	GCA_011798825.1
SAMEA6514128						PD5000056938.1	ERR3901229	GCA_011778265.1
SAMEA6514139						PD5000056907.1	ERR3901248	GCA_011541295.1
SAMEA6514145						PD5000027591.4	ERR3901254	GCA_011788745.1
SAMEA6514153						PD5000042298.5	ERR3901233	GCA_011778705.1
SAMEA6514157						PD5000042481.12	ERR3901235	GCA_011798785.1
SAMEA6514181						PD5000045969.2	ERR3901265	GCA_011541315.1
SAMEA6514185						PD5000056844.1	ERR3901344	GCA_011541395.1
SAMEA6514192						PD5000027583.4	ERR3901273	GCA_011585045.1
SAMEA6514199						PD5000057621.4	ERR3901275	GCA_011585305.1
SAMEA6514201						PD5000056871.2	ERR3901353	GCA_011779145.1
SAMEA6514220						PD5000056881.2	ERR3901358	GCA_011587985.1
SAMEA6514236						PD5000192791.1	ERR3901365	GCA_011790165.1
SAMEA6514251						PD5000178867.1	ERR3901378	GCA_011587665.1
SAMEA6514261						PD5000056902.1	ERR3901382	GCA_011588605.1
SAMEA6514264						PD5000045970.2	ERR3901383	GCA_011778625.1
SAMEA6514269						PD5000027429.6	ERR3901387	GCA_011584925.1
SAMEA6514285						PD5000042384.2	ERR3901398	GCA_011587565.1
SAMEA6514324						PD5000065130.9	ERR3901428	GCA_011587765.1
SAMEA6514364						PD5000045612.1	ERR3901459	GCA_011779945.1
SAMEA6514403						PD5000055942.1	ERR3901483	GCA_011798805.1
SAMEA6514404						PD5000056908.1	ERR3901485	GCA_011585325.1
SAMEA6514407						PD5000056868.4	ERR3901487	GCA_011789065.1
SAMEA6514468						PD5000042395.3	ERR3901535	GCA_011779325.1
SAMEA6514477						PD5000042259.2	ERR3901541	GCA_011782025.1
SAMEA6514483						PD5000056876.1	ERR3901547	GCA_011781945.1
SAMEA6514539						PD5000056943.3	ERR3901688	GCA_011791385.1
SAMEA6514553						PD5000056888.1	ERR3901704	GCA_014338825.1
SAMEA6514560						PD5000056941.1	ERR3901711	GCA_011799555.1
SAMEA6514583						PD5000056932.1	ERR3901732	GCA_011781965.1
SAMEA6514657						PD5000042479.3	ERR3901796	GCA_011783125.1
SAMEA6514709						PD5000042415.6	ERR3901857	GCA_011783355.1
SAMEA6514761						PD5000042315.3	ERR3901906	GCA_011786085.1
SAMEA6514802						PD5000042288.2	ERR3901574	GCA_011788805.1
SAMEA6514806						PD5000056857.1	ERR3901578	GCA_011786905.1
SAMEA6514836						PD5000056890.2	ERR3901606	GCA_011783985.1
SAMEA6514869						PD5000085367.1	ERR3902021	GCA_011784375.1
SAMEA6514876						PD5000056869.1	ERR3902029	GCA_011784225.1

SAMN02367912	FDA: Shaohua Zhao	2010	USA:CT	environmental/other	chicken breast	PD5000030628.1	SRR1840608	GCA 011103075.1
SAMN02368015	FDA: Shaohua Zhao	2011	USA:NY	environmental/other	chicken breast	PD5000032378.1	SRR1122696	GCA 010123745.1
SAMN02368163	FDA	1988	Senegal	clinical		PD5000005713.5	SRR1060516	GCA 010123225.1
SAMN02368368	NCTR-FDA	2000	USA	clinical		PD5000026825.464	SRR1122508	GCA 010714915.1
SAMN02368373	NCTR-FDA	1999	USA	environmental/other	Veterinary Diagnostic	PD5000042369.133	SRR1122503	GCA 010631105.1
SAMN02368708	Technical University Of Denmark			clinical		PD5000040405.7	SRR1106385	GCA 011108435.1
SAMN02378010	Minnesota Department Of Health	6/6/07	USAMN	environmental/other	feces	PD5000013868.575	SRR1177569	GCA 010868365.1
SAMN02378013	Minnesota Department Of Health	6/26/07	USAMN	environmental/other	feces	PD5000013878.168	SRR1177549	GCA 009996925.1
SAMN02378047	Minnesota Department Of Health	4/24/08	USAMN	environmental/other	feces	PD5000032661.317	SRR1166604	GCA 010714915.1
SAMN02378110	Minnesota Department Of Health	6/11/10	USAMN	environmental/other	feces	PD50000138564.9	SRR1158059	GCA 009756055.1
SAMN02378118	Minnesota Department Of Health	7/15/10	USAMN	environmental/other	feces	PD5000027817.12	SRR1158084	GCA 011638235.1
SAMN02378119	Minnesota Department Of Health	7/16/10	USAMN	environmental/other	feces	PD5000032076.9	SRR1158026	GCA 011485085.1
SAMN02378120	Minnesota Department Of Health	8/26/10	USAMN	environmental/other	feces	PD5000032334.1	SRR1158202	GCA 011084105.1
SAMN02378122	Minnesota Department Of Health	8/7/10	USAMN	environmental/other	feces	PD50000113872.9	SRR1158206	GCA 006690745.1
SAMN02378123	Minnesota Department Of Health	8/21/10	USAMN	environmental/other	feces	PD5000032374.2	SRR1157597	GCA 006691785.1
SAMN02378125	Minnesota Department Of Health	9/8/10	USAMN	environmental/other	feces	PD5000032373.2	SRR1090983	GCA 006691465.1
SAMN02378126	Minnesota Department Of Health	9/17/10	USAMN	environmental/other	feces	PD50000065117.13	SRR1089590	GCA 011484805.1
SAMN02378143	Minnesota Department Of Health	4/12/11	USAMN	environmental/other	lung	PD5000032654.261	SRR1089565	GCA 010868865.1
SAMN02378154	Minnesota Department Of Health	11/7/11	USAMN	environmental/other	feces	PD5000029035.8	SRR1067639	GCA 009997745.1
SAMN02378155	Minnesota Department Of Health	6/30/11	USAMN	environmental/other	feces	PD500144856.54	SRR1067620	GCA 010868335.1
SAMN02378169	Minnesota Department Of Health	3/21/12	USAMN	environmental/other	feces	PD5000028842.3	SRR1166648	GCA 011084805.1
SAMN02378182	Minnesota Department Of Health	10/17/12	USAMN	environmental/other	chicken Kiev	PD5000026617.605	SRR1048617	GCA 010941895.1
SAMN02378194	Minnesota Department Of Health	2/11/13	USAMN	environmental/other	chicken breast	PD5000180814.9	SRR1030355	GCA 011089645.1
SAMN02378199	Minnesota Department Of Health	4/29/13	USAMN	environmental/other	raw milk	PD5000026974.4	SRR1029570	GCA 011089285.1
SAMN02403222	VA	7/1/09	USA:VA	environmental/other	Not Provided	PD5000032655.382	SRR1047602	GCA 010705445.1
SAMN02403226	VA	5/30/12	USA:VA	environmental/other	Not Provided	PD5000081815.110	SRR3474702	GCA 004185455.1
SAMN02403227	VA	7/24/13	USA:VA	environmental/other	Not Provided	PD5000026746.181	SRR3474703	GCA 006822585.1
SAMN02403293	VA	12/4/06	USA:VA	environmental/other	Not Provided	PD5000032362.2	SRR1177188	GCA 004224185.1
SAMN02403301	VA	12/29/05	USA:VA	environmental/other	Not Provided	PD5000032361.1	SRR1177544	GCA 007548275.1
SAMN02403303	VA	12/22/05	USA:VA	environmental/other	Not Provided	PD5000026762.119	SRR1177591	GCA 011082945.1
SAMN02403304	VA	7/25/06	USA:VA	environmental/other	Not Provided	PD5000111711.3	SRR1177577	GCA 010946135.1
SAMN02403317	VA	1/19/06	USA:VA	environmental/other	Not Provided	PD5000032357.2	SRR1177571	GCA 006690525.1
SAMN02403320	VA	1/11/06	USA:VA	environmental/other	Not Provided	PD5000026810.182	SRR1177194	GCA 010926375.1
SAMN02403321	VA	12/3/06	USA:VA	environmental/other	Not Provided	PD5000042368.5	SRR1177172	GCA 006862325.1
SAMN02403334	VA	2/28/06	USA:VA	environmental/other	Not Provided	PD5000026745.32	SRR1818523	GCA 010424265.1
SAMN02403343	VA	3/18/06	USA:VA	environmental/other	Not Provided	PD5000060556.1	SRR1177641	GCA 011083085.1
SAMN02403392	VA	2/26/07	USA:VA	environmental/other	Not Provided	PD5000013882.327	SRR1238777	GCA 004227565.1
SAMN02403412	VA	9/30/10	USA:VA	environmental/other	Not Provided	PD5000029655.22	SRR1258614	GCA 011446495.1
SAMN02403418	VA	4/2/12	USA:VA	environmental/other	tomato	PD5000013883.13	SRR1272528	GCA 011086665.1
SAMN02403427	VA	6/1/07	USA:VA	environmental/other	hamburger	PD5000022982.2	SRR1299353	GCA 006689945.1
SAMN02419052	FL	7/27/11	USA:GA	environmental/other	Trachemys scripta from irrigation pond	PD5000032344.1	SRR1067863	GCA 007817195.1
SAMN02419096	FL	2011-05	USA:GA	environmental/other	sediment pond	PD5000032336.1	SRR1105692	GCA 007469495.1
SAMN02483572	United States Food and Drug Administration Office of Regulatory Affairs San Francisco Human and Animal Food Laboratory	1/30/12	USA	environmental/other	hot smoked milkfish	PD5000039817.8	SRR1157599	GCA 011084825.1
SAMN02595031	Arizona Department Of Health	7/18/13	USA:VA	environmental/other	boneless skinless chicken breast	PD5000032335.1	SRR1195648	GCA 007472575.1
SAMN02595611	Minnesota Department Of Health	10/30/07	USAMN	environmental/other	cow stool	PD5000056091.9	SRR1177330	GCA 010424465.1
SAMN02595832	Minnesota Department Of Health	4/22/08	USAMN	environmental/other	pig liver	PD5000054612.3	SRR1266087	GCA 010456805.1
SAMN02595835	Minnesota Department Of Health	5/9/08	USAMN	environmental/other	pig liver	PD5000029630.1	SRR1266040	GCA 010456905.1
SAMN02595843	Minnesota Department Of Health	11/20/08	USAMN	environmental/other	cow tissue composite	PD5000041663.5	SRR1266090	GCA 007834975.1
SAMN02595861	Minnesota Department Of Health	8/24/09	USAMN	environmental/other	pig	PD5000015886.15	SRR1266027	GCA 010655725.1
SAMN02595872	Minnesota Department Of Health	4/8/10	USAMN	environmental/other	bird liver	PD5000037478.28	SRR1266058	GCA 011414115.1
SAMN02595882	Minnesota Department Of Health	11/15/10	USAMN	environmental/other	intestinal raccoon	PD5000032333.1	SRR1266032	GCA 007547895.1
SAMN02595890	Minnesota Department Of Health	4/29/11	USAMN	environmental/other	dog stool	PD5000026818.68	SRR1292271	GCA 010389445.1
SAMN02595904	Minnesota Department Of Health	8/3/12	USAMN	environmental/other	cat liver	PD5000026800.437	SRR1346298	GCA 011417435.1
SAMN02602986						PD5000013874.388		GCA 000213635.1
SAMN02602988						PD5000013889.30		GCA 000198735.1
SAMN02640793	United States Food and Drug Administration	2002	USA	environmental/other	ground beef	PD5000010109.1	SRR1185790	GCA 006691245.1
SAMN02640815	United States Food and Drug Administration	2002	USA	environmental/other	chicken breast	PD5000026832.35	SRR1185842	GCA 00630725.1
SAMN02640878	United States Food and Drug Administration	2002	USA	environmental/other	ground beef	PD5000031535.11	SRR1203017	GCA 006690785.1
SAMN02640882	United States Food and Drug Administration	2002	USA	environmental/other	chicken breast	PD5000050108.12	SRR1202988	GCA 007748355.1
SAMN02640897	United States Food and Drug Administration	2002	USA	environmental/other	chicken breast	PD5000028787.49	SRR1207166	GCA 006690835.1
SAMN02640928	United States Food and Drug Administration	2003	USA	environmental/other	ground turkey	PD500003294.51	SRR1257274	GCA 006822105.1
SAMN02645813	Pennsylvania Department Of Health	1/25/10	USA	environmental/other	chicken breast	PD5000026727.1	SRR1548413	GCA 003867855.1
SAMN02646788	Minnesota Department Of Health	8/24/03	USAMN	environmental/other	feces	PD5000032299.5	SRR1299391	GCA 007906785.1
SAMN02646793	Minnesota Department Of Health	8/13/03	USAMN	environmental/other	feces	PD5000032196.3	SRR1299313	GCA 011448855.1
SAMN02646799	Minnesota Department Of Health	8/27/03	USAMN	environmental/other	feces	PD5000026978.189	SRR1299298	GCA 011081725.1
SAMN02646843	Minnesota Department Of Health	10/20/03	USA:AR	environmental/other	feces	PD5000060958.2	SRR1272873	GCA 011079805.1
SAMN02646948	Minnesota Department Of Health	4/10/07	USAMN	environmental/other	spleen	PD5000077807.6	SRR1233935	GCA 011445675.1
SAMN02646952	Minnesota Department Of Health	7/6/07	USAMN	environmental/other	puffed vegetable snack	PD5000032293.3	SRR1188924	GCA 006632025.1
SAMN02664748	University Of Florida	2011-08	USA:FL	environmental/other	surface water	PD5000046671.2	SRR1207482	GCA 011406715.1
SAMN02664824	Food Safety Laboratory, Cornell University	12/12/07	USA:NY	environmental/other	feces (Bos taurus)	PD5000076645.2	SRR1238752	GCA 010424145.1
SAMN02664832	Food Safety Laboratory, Cornell University	11/8/08	USA:NY	environmental/other	feces (Bos taurus)	PD5000013886.14	SRR1213770	GCA 011447555.1
SAMN02678465	United States Food and Drug Administration	8/13/13	USA	environmental/other	cantaloupe	PD5000027050.110	SRR1269327	GCA 011080645.1
SAMN02678618	United States Food and Drug Administration	10/2/08	USA	environmental/other	tomatillo fresh	PD5000032264.2	SRR1613897	GCA 007816915.1
SAMN02678758	United States Food and Drug Administration	9/16/08	USA	environmental/other	field irrigation system	PD5000078032.25	SRR1509584	GCA 011448095.1
SAMN02678793	United States Food and Drug Administration Office of Regulatory Affairs Pacific Northwest Laboratory	10/6/09	USA	environmental/other	crab in shell	PD5000056051.27	SRR3156729	GCA 007762275.1
SAMN02678865	United States Food and Drug Administration	7/10/13	USA	environmental/other	frozen ostrich	PD500001784.13	SRR1264983	GCA 009932485.1
SAMN02678902	United States Food and Drug Administration	4/21/11	USA	environmental/other	raw shrimp	PD5000028689.2	SRR1272529	GCA 011080765.1
SAMN02696258	United States Food and Drug Administration	10/27/11	USA	environmental/other	lentil vegetable seasoning curry madras	PD5000032220.2	SRR1411118	GCA 008090095.1
SAMN02696395	United States Food and Drug Administration	210/11	USA	environmental/other	frozen grouper fillet fish	PD5000004540.14	SRR1427103	GCA 016377225.1
SAMN02698411	United States Food and Drug Administration	4/12/11	USA	environmental/other	fresh coriander	PD5000026817.30	SRR1427716	GCA 011448395.1
SAMN02699303	FDA	2012	USA:CA	environmental/other	chicken breast	PD500003271.6		GCA 001480485.1
SAMN02699357	FDA	2012	USA:NM	environmental/other	pork chop	PD5000026687.2		GCA 001481175.1
SAMN02699437	Minnesota Department Of Health	5/25/01	USAMN	environmental/other	feces	PD5000013890.5	SRR1333878	GCA 016377265.1
SAMN02699450	Minnesota Department Of Health	1/28/02	USAMN	environmental/other	lung	PD5000079933.2	SRR1346284	GCA 008089745.1
SAMN02699453	Minnesota Department Of Health	4/29/02	USAMN	environmental/other	kidney	PD5000032200.1	SRR1346262	GCA 010498335.1
SAMN02699463	Minnesota Department Of Health	9/16/02	USAMN	environmental/other	feces	PD5000032202.1	SRR1385147	GCA 004218485.1
SAMN02699501	Minnesota Department Of Health	11/1/02	USAMN	environmental/other	feces	PD5000072789.45	SRR1461807	GCA 011448735.1
SAMN02699528	Minnesota Department Of Health	12/12/02	USAMN	environmental/other	lung	PD5000071837.31	SRR1511537	GCA 011450035.1

SAMN0269529	Minnesota Department of Health	12/6/02	USA-MN	environmental/other	issue	PDS000032182.3	SR1511508	GCA 008091565.1
SAMN0269574	Minnesota Department of Health	2/21/03	USA-MN	environmental/other	lung	PDS000005468.3	SR15157680	GCA 008067975.1
SAMN0269567	Minnesota Department of Health	7/8/03	USAN-CA	environmental/other	issue	PDS0000032187.1	SR15174252	GCA 008118306.1
SAMN0269567	Minnesota Department of Health	7/11/03	USAA-R	environmental/other	lung	PDS0000060974.1	SR15174292	GCA 008088105.1
SAMN0269575	Minnesota Department of Health	5/2/04	USA-MN	environmental/other	feces	PDS0000053408.29	SR15149547	GCA 008023655.1
SAMN0269580	Minnesota Department of Health	31/6/05	USA-MN	environmental/other	feces	PDS0000051475.1	SR1752731	GCA 007742675.1
SAMN0269908	Minnesota Department of Health	1/28/10	USA-MN	environmental/other	issue	PDS0000049034.5	SR3035274	GCA 00866455.1
SAMN0269912	Minnesota Department of Health	9/17/10	USA-MN	environmental/other	feces	PDS0000027862.5	SR33306272	GCA 00808815.1
SAMN02712198	FDA Contracted Laboratory	2007-05	USA: United States BRFSo Region 1	environmental/other	feces	PDS0000026739.38	SR33306272	GCA 009041015.2
SAMN02742061	FDA Contracted Laboratory	11/8/06	USACOR	environmental/other	bovine colostrum	PDS0000026833.9	SR1292266	GCA 010904285.1
SAMN02742117	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	environmental bovine	PDS0000032155.2	SR1300679	GCA 006363425.1
SAMN02742123	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	equine environment	PDS0000032154.4	SR1302862	GCA 006690125.1
SAMN02777702	FDA	2003	USAMD	environmental/other	chicken breast	PDS0000032147.9	SR1582184	GCA 006629885.1
SAMN02843556	United States Food and Drug Administration Office of Regulatory Affairs New York Human and Animal Food Laboratory	11/7/01	USA	environmental/other	equid	PDS000002635.9	SR1639646	GCA 008032215.1
SAMN02843557	United States Food and Drug Administration Office of Regulatory Affairs Irvine Human and Animal Food Laboratory	11/8/01	USA	environmental/other	salled egg yolk	PDS000042198.18	SR3385123	GCA 008913985.1
SAMN02843783	United States Food and Drug Administration Office of Regulatory Affairs Pacific Northwest Laboratory	3/5/03	USA	environmental/other	dog chew	PDS000002327.4	SR2096599	GCA 006694585.1
SAMN02843985	United States Food and Drug Administration Office of Regulatory Affairs Irvine Human and Animal Food Laboratory	2/24/04	USA	environmental/other	kasubha safflower	PDS000112852.1	SR2133189	GCA 008997815.1
SAMN02843993	United States Food and Drug Administration Office of Regulatory Affairs New York Human and Animal Food Laboratory	3/2/04	USA	environmental/other	pig ear	PDS000032068.16	SR1744009	GCA 007817315.1
SAMN02844025	United States Food and Drug Administration Office of Regulatory Affairs Irvine Human and Animal Food Laboratory	4/21/04	USA	environmental/other	salled duck egg	PDS0000042321.2	SR2133344	GCA 008997885.1
SAMN02844088	United States Food and Drug Administration Office of Regulatory Affairs Pacific Northwest Laboratory	7/30/04	USA	environmental/other	chutney powder	PDS0000032062.1	SR15155705	GCA 006957365.1
SAMN02844355	United States Food and Drug Administration Office of Regulatory Affairs New York Human and Animal Food Laboratory	4/3/06	USA	environmental/other	mackerel fish	PDS0000027910.23	SR1778912	GCA 008005085.1
SAMN02844501	United States Food and Drug Administration Office of Regulatory Affairs New York Human and Animal Food Laboratory	6/11/07	USA	environmental/other	oyster	PDS0000032040.3	SR1946914	GCA 010676165.1
SAMN02844697	United States Food and Drug Administration Office of Regulatory Affairs Southeast Food and Feed Laboratory	31/0/09	USA	environmental/other	red oak leaf lettuce	PDS000046000.13	SR3219070	GCA 007761495.1
SAMN02845327	United States Food and Drug Administration Office of Regulatory Affairs Arkansas Human and Animal Food Laboratory	8/20/03	USA	environmental/other	red snapper	PDS0000031984.1	SR2534080	GCA 003877935.1
SAMN02845598	United States Food and Drug Administration Office of Regulatory Affairs Irvine Human and Animal Food Laboratory	1/7/05	USA	environmental/other	coriander powder	PDS0000036046.3	SR2102446	GCA 008463645.1
SAMN02845857	United States Food and Drug Administration Office of Regulatory Affairs Southeast Food and Feed Laboratory	11/13/06	USA	environmental/other	ground tomato from field	PDS0000042474.3	SR2002711	GCA 010914375.1
SAMN02845893	United States Food and Drug Administration Office of Regulatory Affairs Arkansas Human and Animal Food Laboratory	2/21/07	USA	environmental/other	extra hot chili powder	PDS0000030026.6	SR2014672	GCA 006018415.1
SAMN02846067	United States Food and Drug Administration Office of Regulatory Affairs New York Human and Animal Food Laboratory	2/7/08	USA	environmental/other	garam masala spice mixture	PDS0000055038.2	SR5883567	GCA 009402185.1
SAMN02846135	United States Food and Drug Administration Office of Regulatory Affairs Southeast Food and Feed Laboratory	7/2/08	USA	environmental/other	feces animal	PDS0000078046.5	SR3092104	GCA 008677375.1
SAMN02846769	United States Food and Drug Administration Office of Regulatory Affairs Irvine Human and Animal Food Laboratory	9/15/09	USA	environmental/other	water	PDS0000053291.39	SR3854321	GCA 011409195.1
SAMN02846949	United States Food and Drug Administration Office of Regulatory Affairs Irvine Human and Animal Food Laboratory	21/7/10	USA	environmental/other	environmental swab	PDS0000048257.21	SR2125026	GCA 008117745.1
SAMN02846950	United States Food and Drug Administration Office of Regulatory Affairs Irvine Human and Animal Food Laboratory	7/15/10	USA	environmental/other	frozen crooked duck egg yolk	PDS0000047482.1	SR1646545	GCA 008023485.1
SAMN02847152	United States Food and Drug Administration Office of Regulatory Affairs Irvine Human and Animal Food Laboratory	7/15/10	USA	environmental/other	frozen crooked duck egg yolk	PDS00011103.34	SR1646546	GCA 008023615.1
SAMN02847552	United States Food and Drug Administration Office of Regulatory Affairs Southeast Food and Feed Laboratory	12/8/10	USA	environmental/other	golden cornflour fillet	PDS000003030.89	SR1693289	GCA 010661765.1
SAMN02849754	FDA Contracted Laboratory	3/7/14	USA	environmental/other	chili flakes	PDS0000037726.1	SR3885497	GCA 005997325.1
SAMN02849788	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	bovine feed	PDS0000031708.3	SR1501486	GCA 006629965.1
SAMN02849946	FDA Contracted Laboratory	2/2/07	USA-WA	environmental/other	feces equine	PDS0000031704.1	SR1501658	GCA 011451135.1
SAMN02894087	FDA	2005	USA-MN	environmental/other	avian carcass rinse water	PDS000013824.215	SR1515026	GCA 011453675.1
SAMN02894174	FDA	2005	USAC-T	environmental/other	chicken breast	PDS0000031698.10	SR1534908	GCA 007546935.1
SAMN02894178	FDA	2005	USA-GA	environmental/other	chicken breast	PDS0000031690.10	SR1528508	GCA 006688865.1
SAMN02894179	FDA	2005	USA-GA	environmental/other	chicken breast	PDS0000025969.60	SR1528483	GCA 006629825.1
SAMN02900050	Not Provided	6/13/13	USA	environmental/other	ground turkey	PDS0000054041.1	SR1528526	GCA 007906755.1
SAMN02900080	Not Provided	9/13/13	USA	environmental/other	water stream	PDS0000030145.1	SR1566330	GCA 006853955.1
SAMN02902665	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	sediment stream	PDS0000031885.2	SR1567297	GCA 008119625.1
SAMN02902721	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	feces	PDS0000031654.2	SR1528533	GCA 008463325.1
SAMN02902756	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	toroise feces	PDS0000076499.1	SR1537514	GCA 011074565.1
SAMN02902789	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	equine culture medium	PDS0000062489.10	SR1544175	GCA 016331065.1
SAMN02910163	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	toroise necropsy abdominal fluid liver	PDS0000055091.3	SR1576564	GCA 010424025.1
SAMN02911914	CDC	2012	USA-ID	environmental/other	feces bovine	PDS000004126.1	SR1534843	GCA 007834895.1
SAMN02911926	CDC	2011	USA	clinical	clinical sample	PDS0000026857.7		GCA 001473345.1
SAMN02911962	CDC	2012	USA	clinical	clinical sample	PDS0000026807.80		GCA 001473385.1
SAMN02911971	CDC	2011	USA	clinical	clinical sample	PDS000013853.325		GCA 001475245.1
SAMN02911984	CDC	2012	USA	clinical	clinical sample	PDS000013712.42		GCA 001475335.1
SAMN02918702	United States Food and Drug Administration Office of Regulatory Affairs Arkansas Human and Animal Food Laboratory	7/8/10	USA	environmental/other	frozen animal feed baby mouse	PDS0000020046.2	SR1735343	GCA 010438435.1
SAMN02927354	CDC	2012-07	USA	clinical		PDS000002705.1280	SR1577857	GCA 008554015.1
SAMN02952708	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	feces avian	PDS0000031605.1	SR1554517	GCA 005524235.1
SAMN02952817	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	porcine necropsy intestine	PDS000013879.106	SR1564495	GCA 006847705.1
SAMN02989090	Minnesota Department of Health	4/14/13	USA-MN	clinical	feces	PDS0000032657.39	SR1569728	GCA 008067215.1
SAMN03002011	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	feces equine	PDS0000052609.2	SR1569681	GCA 008010785.1
SAMN03002026	FDA Contracted Laboratory	1/9/07	USA-WA	environmental/other	Not Provided	PDS0000031586.1	SR1569632	GCA 005935125.1
SAMN03024018	Western Carolina University (WCU)	9/13/13	USAN-CA	environmental/other	bird feces	PDS0000031575.1	SR1773589	GCA 005756135.1
SAMN03069683	Not Provided	24-07-07	USA	environmental/other	water gallon	PDS000166581.3	SR1576129	GCA 006010185.1
SAMN03097225	FERA	2006/2008	United Kingdom	environmental/other	retail meat	PDS0000026929.23	SR1604875	GCA 007177635.1
SAMN03097226	FERA	2006/2008	United Kingdom	environmental/other	retail meat	PDS0000026935.4	SR1604876	GCA 010411025.1
SAMN03097227	FERA	2006/2008	United Kingdom	environmental/other	retail meat	PDS000042461.99	SR1604877	GCA 008090465.1
SAMN03112887	VA	6/20/12	USA-VA	environmental/other	Not Provided	PDS0000074567.36	SR1660417	GCA 010697625.1
SAMN03120578	Hoi-Shan Kwan's Lab	2007	Hong Kong	clinical	blood	PDS0000026807.1		GCA 001295425.1
SAMN03121892	Hoi-Shan Kwan's Lab	2005	Hong Kong	clinical	stool	PDS000004586.4		GCA 001295665.1
SAMN03121897	Hoi-Shan Kwan's Lab	2006	Hong Kong	clinical	stool	PDS000004586.4		GCA 001295695.1
SAMN03121900	Hoi-Shan Kwan's Lab	2006	Hong Kong	clinical	stool	PDS000004607.3		GCA 001296365.1
SAMN03121904	Hoi-Shan Kwan's Lab	2006	Hong Kong	clinical	stool	PDS000004607.3		GCA 001295495.1
SAMN03135169	PHE	10/17/12	United Kingdom: South of England	clinical	human	PDS0000002642.19	SR1618673	GCA 006387065.1
SAMN03144979	the NSW Ministry of Health and the NSW Enteric Reference Laboratory	11/9/06	Australia: Sydney	clinical	faecal sample	PDS0000031528.45	SR1645473	GCA 010796815.1
SAMN03152378	PHE	2012-10	United Kingdom: North of England	clinical	human	PDS000002654.21	SR1635617	GCA 006387395.1
SAMN03152500	PHE	2012-04	United Kingdom: North of England	clinical	human	PDS0000042336.2	SR1635099	GCA 006387975.1
SAMN03152398	PHE	2012-11	United Kingdom: Midlands and East of England	clinical	human	PDS0000026708.89	SR1635098	GCA 006387835.1
SAMN03152400	PHE	2012-12	United Kingdom: London	clinical	human	PDS0000026812.319	SR1635102	GCA 006387875.1
SAMN03152422	PHE	2012-10	United Kingdom: South of England	clinical	human	PDS0000042450.95	SR1635124	GCA 006388515.1
SAMN03152426	PHE	2012-01	United Kingdom: South of England	clinical	human	PDS0000042247.14	SR1635121	GCA 006388975.1
SAMN03152428	PHE	2012-04	United Kingdom: Midlands and East of England	clinical	human	PDS0000030204.56	SR1635127	GCA 006389635.1
SAMN03160320	PHE	2013-07	United Kingdom: Midlands and East of England	clinical	human	PDS0000031472.1	SR1638967	GCA 008010815.1
SAMN03160768	VA	5/30/12	USA-VA	environmental/other	Not Provided	PDS0000026558.115	SR1694063	GCA 008025735.1
SAMN03168520	PHE	2012-09	United Kingdom: South of England	clinical	human	PDS0000030104.7	SR1645148	GCA 008031115.1
SAMN03168554	PHE	2012-10	United Kingdom: London	clinical	human	PDS0000032663.161	SR1645174	GCA 008032065.1
SAMN03168564	PHE	2012-04	United Kingdom: North of England	clinical	human	PDS000004419.6	SR1645184	GCA 007743875.1
SAMN03168693	PHE	2012-06	United Kingdom: South of England	clinical	human	PDS0000026729.64	SR1645265	GCA 011595525.1
SAMN03168721	PHE	2012-08	United Kingdom: North of England	clinical	human	PDS000013822.12	SR1645332	GCA 011595065.1

SAMN03168731	PHE	2012-07		clinical	human	PDS000020392.12	SRR1645347	GCA 011078645.1
SAMN03168750	PHE	2012-04	United Kingdom: North of England	clinical	human	PDS0000031487.3	SRR1645367	GCA 004280775.1
SAMN03168771	PHE	2012-11	United Kingdom: London	clinical	human	PDS0000030226.6	SRR1645387	GCA 011624845.1
SAMN03168788	PHE	2012-01	United Kingdom: South of England	clinical	human	PDS000004621.51	SRR1645430	GCA 011641475.1
SAMN03168804	PHE	2012-07	United Kingdom: South of England	clinical	human	PDS000013823.51	SRR1645446	GCA 010878385.1
SAMN03168853	PHE	2012-07	United Kingdom: South of England	clinical	human	PDS000026747.304	SRR1645553	GCA 006629685.1
SAMN03168856	PHE	2012-06	United Kingdom: South of England	clinical	human	PDS000027000.3	SRR1645555	GCA 011624905.1
SAMN03168892	PHE	2012-12	United Kingdom: Midlands and East of England	clinical	human	PDS000028712.6	SRR1645583	GCA 008043455.1
SAMN03168903	PHE	2012-08	United Kingdom: London	clinical	human	PDS000042439.11	SRR1645601	GCA 008043635.1
SAMN03168956	PHE	2012-01	United Kingdom: London	clinical	human	PDS000056861.1	SRR1645606	GCA 008043675.1
SAMN03168959	PHE	2012-10	United Kingdom: North of England	clinical	human	PDS000056878.5	SRR1645611	GCA 008042975.1
SAMN03168960	PHE	2012-09	United Kingdom: South of England	clinical	human	PDS0000206414.8	SRR1645608	GCA 008035085.1
SAMN03168981	PHE	2012-04	United Kingdom: South of England	clinical	human	PDS000031480.6	SRR1645717	GCA 006627165.1
SAMN03169013	PHE	2012-04	United Kingdom: London	clinical	human	PDS000042323.4	SRR1645748	GCA 008043835.1
SAMN03169028	PHE	2012-08	United Kingdom: South of England	environmental/other	animal	PDS000013784.14	SRR1645761	GCA 008043915.1
SAMN03169052	PHE	2012-08	United Kingdom: North of England	clinical	human	PDS000059907.10	SRR1645787	GCA 011553195.1
SAMN03169074	PHE	2012-03	United Kingdom: North of England	clinical	human	PDS000013762.90	SRR1645810	GCA 008042855.1
SAMN03169122	PHE	2012-09	United Kingdom: South of England	clinical	human	PDS000002738.143	SRR1645855	GCA 008006325.1
SAMN03169164	PHE	2012-04	United Kingdom: Midlands and East of England	clinical	human	PDS000031475.6	SRR1645876	GCA 007742795.1
SAMN03169190	PHE	2012-08	United Kingdom: North of England	clinical	human	PDS00000804.31	SRR1645902	GCA 008002145.1
SAMN03169216	PHE	2012-12	United Kingdom: Midlands and East of England	clinical	human	PDS000040805.2	SRR1645930	GCA 008018355.1
SAMN03169244	PHE	2012-01	United Kingdom: North of England	clinical	human	PDS000044949.12	SRR1645942	GCA 008018675.1
SAMN03169274	PHE	2012-03	United Kingdom: North of England	clinical	human	PDS000047449.1	SRR1645963	GCA 007743455.1
SAMN03169279	PHE	2012-10	United Kingdom: Midlands and East of England	clinical	human	PDS000026707.2	SRR1645986	GCA 008018735.1
SAMN03169325	PHE	2012-06	United Kingdom: North of England	clinical	human	PDS0000085764.1	SRR1646010	GCA 008018855.1
SAMN03169402	PHE	2012-11	United Kingdom: Midlands and East of England	clinical	human	PDS000039585.14	SRR1646070	GCA 008021475.1
SAMN03169443	PHE	2012-11	United Kingdom: Midlands and East of England	clinical	human	PDS000031464.1	SRR1646094	GCA 004192755.1
SAMN03169477	PHE	2012-10	United Kingdom: South of England	clinical	human	PDS000026614.75	SRR1646112	GCA 008020535.1
SAMN03169484	PHE	2012-06	United Kingdom: South of England	clinical	human	PDS000026707.2	SRR1646121	GCA 008020515.1
SAMN03169485	PHE	2012-09	United Kingdom: North of England	clinical	human	PDS000002953.2	SRR1646119	GCA 008020735.1
SAMN03169511	PHE	2012-08	United Kingdom: Midlands and East of England	clinical	human	PDS000030182.5	SRR1646146	GCA 008020015.1
SAMN03169527	PHE	2012-12	United Kingdom: London	clinical	human	PDS000070409.3	SRR1646162	GCA 008020915.1
SAMN03169551	PHE	2012-07	United Kingdom: North of England	clinical	human	PDS000066060.1	SRR1646198	GCA 008021295.1
SAMN03169568	PHE	2012-10	United Kingdom: Midlands and East of England	clinical	human	PDS000013849.88	SRR1646230	GCA 008021395.1
SAMN03169574	PHE	2012-11	United Kingdom: Midlands and East of England	clinical	human	PDS000026643.11	SRR1646238	GCA 008022175.1
SAMN03169627	PHE	2012-07	United Kingdom: South of England	clinical	human	PDS000056921.4	SRR1646285	GCA 008021845.1
SAMN03169682	PHE	2012-08	United Kingdom: London	clinical	human	PDS000031144.2	SRR1646378	GCA 008022995.1
SAMN032255380	Middle East Technical University Food Engineering Department	12/7/12	Turkey	environmental/other	offal	PDS000013796.19	SRR1849305	GCA 006631245.1
SAMN03275998	FDA	2007	USA:TN	environmental/other	ground beef	PDS000031397.8	SRR1967103	GCA 006632545.1
SAMN03276090	FDA	2007	USA:CA	environmental/other	chicken breast	PDS0000038752.8	SRR2070990	GCA 006634385.1
SAMN03276111	FDA	2007	USA:NY	environmental/other	chicken breast	PDS000031391.2	SRR2102423	GCA 006663865.1
SAMN03276121	FDA	2007	USA:CO	environmental/other	ground beef	PDS0000092567.2	SRR2533574	GCA 007816355.1
SAMN03282125	FDA Contracted Laboratory	3/26/14	USA:GA	environmental/other	chicken	PDS000028417.21	SRR1812842	GCA 006016175.1
SAMN03285122	United States Food and Drug Administration	12/12/14	USA	environmental/other	hazelnut	PDS000028975.1	SRR1774093	GCA 006859685.1
SAMN03359716					culture	PDS000026652.51	SRR1814963	GCA 009255405.1
SAMN03464090	United States Food and Drug Administration	8/13/14	USA	environmental/other	water	PDS000011260.2	SRR2054219	GCA 006878805.1
SAMN03465619	PHE	2014-09	United Kingdom: Midlands and East of England	clinical	human	PDS000026809.31	SRR1957724	GCA 010536655.1
SAMN03465697	PHE	2014-09	United Kingdom: South of England	clinical	human	PDS000004546.13	SRR1957801	GCA 011624665.1
SAMN03465728	PHE	2014-11	United Kingdom: London	clinical	human	PDS0000085353.2	SRR1967832	GCA 011085325.1
SAMN03465747	PHE	2014-07	United Kingdom: North of England	clinical	human	PDS0000175001.7	SRR1957845	GCA 010907965.1
SAMN03465757	PHE	2014-08	United Kingdom: Midlands and East of England	clinical	human	PDS000022605.1	SRR1967855	GCA 011407815.1
SAMN03465786	PHE	2014-10	United Kingdom: Midlands and East of England	clinical	human	PDS000030975.3	SRR1957884	GCA 011086175.1
SAMN03465810	PHE	2014-08	United Kingdom: South of England	clinical	human	PDS0000119524.2	SRR1957908	GCA 011093325.1
SAMN03465896	PHE	2014-09	United Kingdom: Midlands and East of England	clinical	human	PDS000013784.23	SRR1957994	GCA 011087425.1
SAMN03465914	PHE	2014-09	United Kingdom: Midlands and East of England	clinical	human	PDS000013717.8	SRR1958012	GCA 011638155.1
SAMN03465943	PHE	2014-10	United Kingdom: North of England	clinical	human	PDS000036017.5	SRR1958042	GCA 011087145.1
SAMN03466136	PHE	2014-08	United Kingdom: North of England	clinical	human	PDS000013747.38	SRR1958243	GCA 011087445.1
SAMN03466160	PHE	2014-08	United Kingdom: North of England	clinical	human	PDS000013840.10	SRR1958267	GCA 011094505.1
SAMN03466213	PHE	2014-10	United Kingdom: North of England	clinical	human	PDS000031258.5	SRR1958320	GCA 006632425.1
SAMN03466246	PHE	2014-10	United Kingdom: Midlands and East of England	clinical	human	PDS000013790.4	SRR1958353	GCA 010964125.1
SAMN03466269	PHE	2014-09	United Kingdom: London	clinical	human	PDS000031255.1	SRR1958376	GCA 004270105.1
SAMN03466279	PHE	2014-10	United Kingdom: London	clinical	human	PDS000030985.3	SRR1958386	GCA 011546975.1
SAMN03466287	PHE	2014-10	United Kingdom: North of England	clinical	human	PDS000065754.24	SRR1958394	GCA 010949275.1
SAMN03466288	PHE	2014-09	United Kingdom: London	clinical	human	PDS0000194632.1	SRR1958395	GCA 011483165.1
SAMN03466298	PHE	2014-10	United Kingdom: London	clinical	human	PDS000028757.393	SRR1958406	GCA 011484025.1
SAMN03466300	PHE	2014-10	United Kingdom: North of England	clinical	human	PDS000026672.5	SRR1958407	GCA 011094805.1
SAMN03466369	PHE	2014-10	United Kingdom: South of England	clinical	human	PDS000050023.2	SRR1958476	GCA 011094925.1
SAMN03466374	PHE	2014-09	United Kingdom: South of England	clinical	human	PDS000031105.2	SRR1958481	GCA 011094865.1
SAMN03466456	PHE	2014-09	United Kingdom: South of England	clinical	human	PDS000030327.1	SRR1958563	GCA 011633225.1
SAMN03466568	PHE	2014-09	United Kingdom: South of England	clinical	human	PDS000051729.2	SRR1958675	GCA 010907205.1
SAMN03468422	PHE	2012-11	United Kingdom: None	clinical	human	PDS000030932.2	SRR1959219	GCA 011101575.1
SAMN03468479	PHE	2014-12	United Kingdom: South of England	clinical	human	PDS000047100.3	SRR1959274	GCA 011102175.1
SAMN03468483	PHE	2014-12	United Kingdom: London	clinical	human	PDS000013771.3	SRR1959278	GCA 010960215.1
SAMN03468530	PHE	2014-09	United Kingdom: North of England	clinical	human	PDS000031228.4	SRR1959284	GCA 006756325.1
SAMN03468568	PHE	2015-01	United Kingdom: Midlands and East of England	clinical	human	PDS0000138659.1	SRR1959363	GCA 011534585.1
SAMN03468568	PHE	2015-01	United Kingdom: North of England	clinical	human	PDS000013788.10	SRR1959400	GCA 011097465.1
SAMN03468607	PHE	2014-12	United Kingdom: South of England	clinical	human	PDS000074030.6	SRR1959425	GCA 011096485.1
SAMN03468611	PHE	2012-11	United Kingdom: None	clinical	human	PDS000030391.4	SRR1959429	GCA 011107635.1
SAMN03468645	PHE	2014-04	United Kingdom: North of England	clinical	human	PDS000031220.12	SRR1959465	GCA 007748295.1
SAMN03468703	PHE	2014-12	United Kingdom: Wales	clinical	human	PDS000028519.2	SRR1959499	GCA 011640355.1
SAMN03468716	PHE	2014-12	United Kingdom: None	clinical	human	PDS0000046543.1	SRR1959502	GCA 01086865.1
SAMN03468768	PHE	2014-12	United Kingdom: South of England	clinical	human	PDS0000093289.4	SRR1960036	GCA 011109995.1

Supplemental Table 3.9. Select metadata downloaded from NCBI for serovar Typhimurium isolates included in Figure S3.7B.

Blue highlight indicates samples chosen for Figure 3.5A.

BioSample	Collected by	Collection date	Location	Isolation type	Isolation source	SNP cluster	Assembly	Run	Isolate	Strain	Computed types
SAMN07424695	CDC	2017-06	USA	clinical	stool	PDS000051476.18	GCA_008884465.1	SRR5888672	PDT000230419.2	PNUSAS018095	antigen formula=4:1.2,serotype=Typhimurium
SAMN07455300		2017-03	USA	clinical	stool	PDS000064404.12	GCA_008435845.1	SRR5921488	PDT000232449.2	PNUSAS017228	antigen formula=4:1.2,serotype=Typhimurium
SAMN07694680		2017-06	USA	clinical	stool	PDS000181137.1	GCA_009553555.1	SRR6107687	PDT000244544.2	PNUSAS024036	antigen formula=4:1.2,serotype=Typhimurium
SAMN08407830	CDC	2018-01	USA	clinical	stool	PDS000076391.11	GCA_016381345.1	SRR6729918	PDT000282937.4	PNUSAS032925	antigen formula=4:1.2,serotype=Typhimurium
SAMN08607476	CDC	2017-09	USA	clinical		PDS000049109.27	GCA_007561945.1	SRR6782607	PDT000289579.2	PNUSAS033431	antigen formula=4:1.2,serotype=Typhimurium
SAMN08773726	CDC	2017-12	USA	clinical	stool	PDS000043305.4	GCA_007202195.1	SRR6880186	PDT000297011.3	PNUSAS033206	antigen formula=4:1.2,serotype=Typhimurium
SAMN09437554	CDC	2017-11	USA	clinical		PDS000126891.2	GCA_006647405.1	SRR7352850	PDT000331872.1	PNUSAS042448	antigen formula=4:1.2,serotype=Typhimurium
SAMN10241133	CDC	2018-09	USA	clinical		PDS000038355.20	GCA_016300485.1	SRR8054302	PDT000391315.1	PNUSAS056026	antigen formula=4:1.2,serotype=Typhimurium
SAMN11102910	KY-M	7/9/18	USA:KY	environmental/other	Odocoileus virginianus Feces	PDS000044303.2	GCA_006239385.1	SRR8767784	PDT000481252.1	SAL-18-VL-LA-KY-0011	antigen formula=4:1.2,serotype=Typhimurium
SAMN11526594	CDC	2019-03	USA	clinical		PDS000101797.2	GCA_005602395.1	SRR8984755	PDT000496339.1	PNUSAS070747	antigen formula=4:1.2,serotype=Typhimurium
SAMN12036221	CDC	2019-05	USA	clinical		PDS000054358.20	GCA_006425235.1	SRR9283102	PDT000521479.1	PNUSAS077320	antigen formula=4:1.2,serotype=Typhimurium
SAMN14083425	CDC	2020-01	USA	clinical		PDS000149796.1	GCA_011489485.1	SRR11067570	PDT000682507.1	PNUSAS134090	antigen formula=4:1.2,serotype=Typhimurium
SAMN15399837	CDC	2020-06	USA	clinical		PDS000150767.2	GCA_016167695.1	SRR12109672	PDT000773851.1	PNUSAS149967	antigen formula=4:1.2,serotype=Typhimurium
SAMN19974806	CDC	2021-06	USA	clinical		PDS000192391.1	GCA_019087325.1	SRR15006564	PDT001080017.1	PNUSAS209577	antigen formula=4:1.2,serotype=Typhimurium
SAMN20090106		2021-06	USA	clinical		PDS000093278.1	GCA_019162375.1	SRR15058485	PDT001083896.1	1019624001	antigen formula=4:1.2,serotype=Typhimurium
SAMN29376357	CDC	2022-05	USA	clinical		PDS000113229.3	GCA_023972455.1	SRR19866112	PDT001347919.1	PNUSAS277622	antigen formula=4:1.2,serotype=Typhimurium
SAMN31022944	CDC	2022-09	USA	clinical		PDS000121793.1	GCA_025418035.1	SRR21712787	PDT001434404.1	PNUSAS303535	antigen formula=4:1.2,serotype=Typhimurium
SAMN34381829		2023	USA	clinical		PDS000154543.2	GCA_032269965.1	SRR24313502	PDT001708295.1	PNUSAS344096	antigen formula=4:1.2,serotype=Typhimurium
SAMN42161866				clinical		PDS000187829.1	GCA_040425175.1	SRR29644432	PDT002234490.1	PNUSAS447913	antigen formula=4:1.2,serotype=Typhimurium
SAMN43494260				clinical		PDS000197424.1	GCA_041764295.1	SRR30541472	PDT002362326.1	PNUSAS470360	antigen formula=4:1.2,serotype=Typhimurium

Supplemental Table S3.10 Weather metadata for day of sampling and two days prior from weather station 1 for Creek A.

Year	Julian Day	Max Air Temperature(F)	Min Air Temperature(F)	Avg Relative Humidity(%)	Avg Wind Speed(mph)	Max Wind Speed (m/s)	Total Solar Radiation(MJ/m ²)	Total Rain(in)
2021	18-Nov	70.9	44.3	74.0	3.4	9.1	9.1	0.0
2021	19-Nov	65.9	35.9	48.5	2.8	5.1	13.6	0.0
2021	20-Nov	53.5	39.0	49.3	4.7	9.1	13.3	0.0
2021	16-Dec	65.8	50.1	68.7	1.5	3.5	5.7	0.0
2021	17-Dec	68.6	55.0	91.5	0.9	2.7	5.1	0.0
2021	18-Dec	63.8	56.8	98.0	1.9	5.1	1.7	0.6
2022	13-Jan	53.3	32.8	68.5	3.8	3.5	7.9	0.0
2022	14-Jan	56.6	35.6	69.6	2.6	2.7	11.8	0.0
2022	15-Jan	43.3	36.6	67.2	6.6	5.1	3.8	0.0
2022	17-Feb	65.8	52.4	88.7	4.9	14.7	2.2	1.4
2022	18-Feb	59.9	32.9	68.3	7.2	9.1	17.0	0.1
2022	19-Feb	57.4	28.9	51.4	4.8	9.9	17.7	0.0
2022	17-Mar	71.4	51.8	75.7	2.8	7.5	20.2	0.0
2022	18-Mar	56.1	47.4	94.3	3.3	12.3	1.8	0.7
2022	19-Mar	62.2	48.2	70.7	6.4	10.7	20.6	0.1
2022	14-Apr	73.8	52.6	65.2	3.6	8.3	16.2	0.0
2022	15-Apr	72.4	42.4	51.3	4.0	9.1	25.3	0.0
2022	16-Apr	65.1	53.1	85.6	2.7	5.1	8.0	0.8
2022	19-May	88.9	65.6	67.9	2.9	7.5	23.8	0.0
2022	20-May	86.8	68.9	69.0	3.0	9.1	24.8	0.0
2022	21-May	86.2	68.0	73.7	3.0	11.5	19.1	0.0
2022	16-Jun	98.2	75.8	68.6	2.2	5.9	25.3	0.0
2022	17-Jun	92.2	73.7	75.7	3.7	11.5	18.8	0.1
2022	18-Jun	90.4	70.0	58.9	3.2	8.3	26.5	0.0
2022	14-Jul	89.8	71.9	70.7	2.3	5.1	22.6	0.0
2022	15-Jul	92.6	67.0	71.5	1.8	5.9	23.9	0.1
2022	16-Jul	90.1	69.4	76.3	2.1	5.1	19.3	0.0
2022	18-Aug	81.8	66.6	82.6	2.3	5.1	17.1	0.2
2022	19-Aug	78.3	69.6	88.7	3.3	6.7	9.1	0.0
2022	20-Aug	80.1	70.1	91.5	2.1	5.1	8.7	0.2
2022	22-Sep	90.4	65.0	69.4	4.3	14.7	17.0	0.0
2022	23-Sep	78.2	53.9	61.6	2.6	5.9	21.6	0.0
2022	24-Sep	84.0	50.9	63.5	2.1	5.1	20.2	0.0
2022	20-Oct	62.3	34.0	52.4	3.2	9.1	17.3	0.0
2022	21-Oct	70.7	36.4	56.8	2.1	5.1	16.9	0.0
2022	22-Oct	76.7	38.1	60.4	2.0	5.1	16.4	0.0
2022	17-Nov	47.9	34.1	46.6	5.3	6.7	13.7	0.0
2022	18-Nov	54.6	28.8	47.3	3.5	7.5	13.2	0.0
2022	19-Nov	51.4	35.3	54.9	5.6	8.3	10.6	0.0
2022	12-Dec	58.0	48.3	88.1	2.9	5.9	8.8	0.0
2022	13-Dec	49.8	42.5	76.2	5.8	9.9	8.4	0.0
2022	14-Dec	45.8	40.3	91.5	7.2	9.1	0.8	0.9
2023	19-Jan	69.1	55.7	67.2	5.6	11.5	8.4	0.1
2023	20-Jan	56.6	41.5	48.6	7.5	9.9	13.1	0.0
2023	21-Jan	49.2	35.1	59.2	4.5	6.7	8.3	0.0
2023	16-Feb	66.7	49.7	89.2	2.2	6.7	6.0	0.0
2023	17-Feb	64.2	33.6	74.8	8.2	12.3	12.3	0.8
2023	18-Feb	58.9	28.6	54.8	2.6	5.1	17.6	0.0
2023	16-Mar	67.9	31.1	49.2	3.2	6.7	20.2	0.0
2023	17-Mar	60.6	43.7	84.6	3.7	8.3	2.1	0.5
2023	18-Mar	53.2	37.3	50.2	6.7	9.1	20.1	0.0
2023	13-Apr	69.4	46.7	76.5	3.2	9.9	12.3	0.3
2023	14-Apr	69.6	57.8	91.3	2.4	5.9	10.2	0.8
2023	15-Apr	77.5	58.0	75.8	2.7	8.3	22.3	0.0
2023	18-May	69.2	60.8	86.3	5.3	8.3	5.9	0.0
2023	19-May	73.4	59.1	77.4	4.3	6.7	16.6	0.0
2023	20-May	79.1	64.1	83.2	3.0	5.1	12.6	0.4
2023	15-Jun	81.5	65.6	81.2	4.1	7.5	20.7	0.4
2023	16-Jun	84.7	65.5	74.4	4.5	9.9	23.8	0.7
2023	17-Jun	87.1	66.6	66.2	3.0	5.9	23.8	0.0
2023	13-Jul	91.5	71.5	72.5	2.9	8.3	24.5	0.0
2023	14-Jul	91.1	74.0	77.4	3.1	8.3	21.0	0.0
2023	15-Jul	89.3	72.8	76.5	3.2	8.3	19.3	0.1
2023	17-Aug	84.5	64.4	70.2	3.5	7.5	23.1	0.0
2023	18-Aug	89.3	68.6	64.4	3.6	6.7	22.6	0.0
2023	19-Aug	84.3	63.1	70.6	3.0	5.9	23.3	0.0
2023	14-Sep	76.8	68.3	79.4	3.1	5.9	7.0	0.0
2023	15-Sep	75.3	65.5	78.2	3.8	5.9	10.3	0.0
2023	16-Sep	72.4	62.7	87.4	2.5	5.1	6.6	0.1
2023	20-Oct	70.0	55.4	79.1	4.0	9.9	11.5	0.3
2023	21-Oct	72.2	50.8	56.0	4.2	8.3	17.0	0.0
2023	22-Oct	75.2	50.8	48.4	4.2	7.5	16.5	0.0

Supplemental Table S3.11 Weather metadata for day of sampling and two days prior from weather station 2 for Creek B.

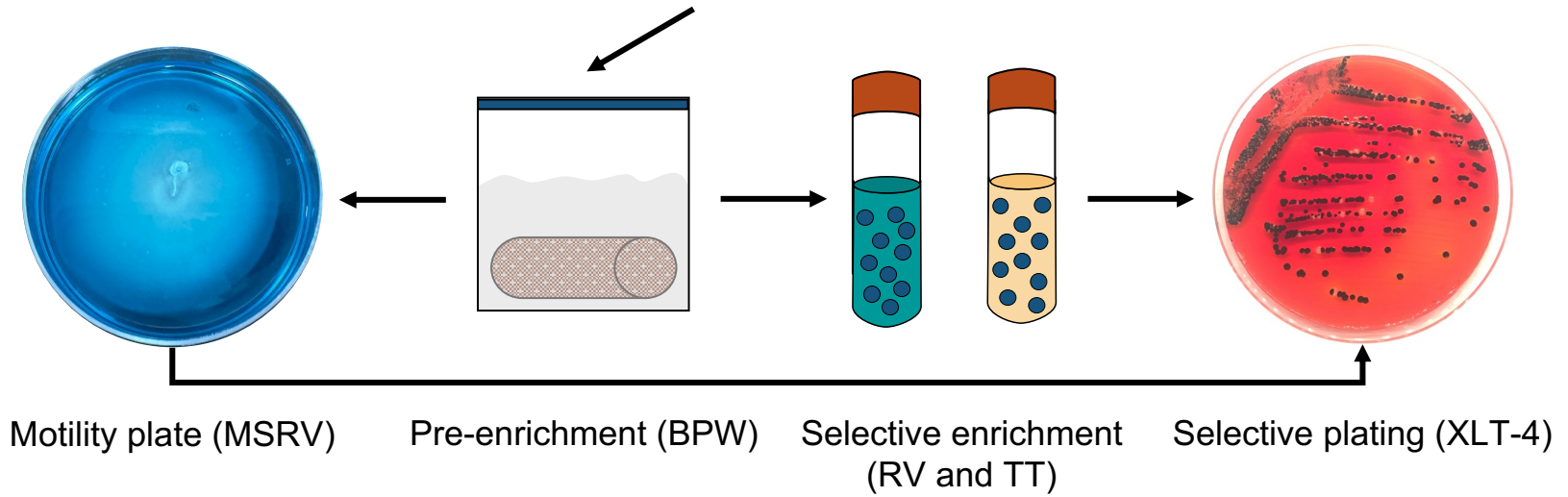
Year	Julian Day	Max Air Temperature(F)	Min Air Temperature(F)	Avg Relative Humidity(%)	Avg Wind Speed(mph)	Max Wind Speed (m/s)	Total Solar Radiation(MJ/m ²)	Total Rain(in)
2021	18-Nov	75.2	42.7	78.4	3.0	7.5	10.7	0.0
2021	19-Nov	61.3	33.9	43.5	2.4	4.3	14.8	0.0
2021	20-Nov	55.4	34.7	52.5	4.5	7.5	14.1	0.0
2021	16-Dec	63.1	44.0	80.7	2.1	4.3	5.3	0.0
2021	17-Dec	67.3	53.9	94.1	1.9	5.1	6.0	0.0
2021	18-Dec	65.6	57.3	99.3	3.2	6.7	2.0	0.8
2022	13-Jan	57.7	29.4	60.3	3.7	8.3	9.6	0.0
2022	14-Jan	56.4	36.1	64.7	2.6	5.9	12.9	0.0
2022	15-Jan	44.1	35.4	69.3	4.2	7.5	3.9	0.0
2022	17-Feb	72.1	52.2	78.9	6.8	13.9	5.9	0.0
2022	18-Feb	67.7	32.3	74.8	5.6	9.9	16.7	1.0
2022	19-Feb	58.4	27.6	52.7	4.2	6.7	18.6	0.0
2022	17-Mar	73.5	52.6	76.5	2.4	5.1	20.5	0.0
2022	18-Mar	58.1	50.3	91.7	3.0	11.5	2.5	0.7
2022	19-Mar	67.9	48.5	69.3	7.1	9.9	20.6	0.0
2022	14-Apr	77.6	56.0	64.3	4.9	9.9	17.1	0.0
2022	15-Apr	75.6	41.8	46.9	3.2	7.5	26.0	0.0
2022	16-Apr	64.0	53.0	89.2	2.8	4.3	7.2	0.6
2022	19-May	93.0	63.4	60.8	3.3	7.5	24.2	0.0
2022	20-May	89.1	67.5	62.4	4.1	9.1	26.6	0.0
2022	21-May	89.2	66.2	71.3	3.7	7.5	19.8	0.0
2022	16-Jun	97.3	73.4	64.2	2.0	5.9	22.3	0.0
2022	17-Jun	95.3	73.4	65.9	2.8	7.5	24.8	0.0
2022	18-Jun	92.4	68.0	57.7	2.2	5.1	28.1	0.0
2022	14-Jul	86.8	71.2	80.7	1.6	5.1	16.6	0.0
2022	15-Jul	90.7	68.8	71.3	1.8	5.1	25.4	0.0
2022	16-Jul	90.2	68.5	75.0	2.2	4.3	24.9	0.0
2022	18-Aug	82.7	69.1	81.1	1.5	5.1	14.3	0.0
2022	19-Aug	76.6	70.1	93.8	1.5	4.3	6.5	0.1
2022	20-Aug	85.4	71.5	84.6	0.9	5.9	14.6	0.5
2022	22-Sep	93.0	64.0	71.4	1.9	11.5	17.9	0.1
2022	23-Sep	78.5	53.0	57.8	2.3	6.7	21.5	0.0
2022	24-Sep	80.7	47.9	64.4	1.4	5.1	20.5	0.0
2022	20-Oct	63.5	31.1	53.6	1.7	5.9	17.4	0.0
2022	21-Oct	69.7	33.6	56.8	1.2	4.3	17.6	0.0
2022	22-Oct	74.3	35.3	62.2	1.3	4.3	17.3	0.0
2022	17-Nov	50.1	30.5	53.7	2.2	5.1	14.6	0.0
2022	18-Nov	55.3	27.2	50.0	1.7	6.7	14.3	0.0
2022	19-Nov	55.0	32.1	55.7	2.5	6.7	11.1	0.0
2022	12-Dec	55.9	43.5	94.9	1.9	3.5	4.0	0.0
2022	13-Dec	50.7	40.4	83.1	3.5	5.9	7.6	0.0
2022	14-Dec	44.9	41.0	89.4	3.8	5.9	0.9	0.5
2023	19-Jan	71.3	56.2	68.8	7.0	10.7	7.5	0.0
2023	20-Jan	60.3	43.1	43.8	5.6	9.1	13.5	0.0
2023	21-Jan	47.7	37.0	60.5	3.4	5.9	5.2	0.0
2023	16-Feb	73.0	43.8	82.6	3.7	7.5	8.7	0.0
2023	17-Feb	66.0	33.9	70.9	8.4	12.3	13.0	0.5
2023	18-Feb	56.7	26.7	51.9	3.0	5.1	18.3	0.0
2023	16-Mar	67.5	29.2	46.1	3.8	6.7	21.9	0.0
2023	17-Mar	60.1	47.9	82.0	4.8	9.9	3.8	0.6
2023	18-Mar	54.6	36.4	52.8	6.0	9.9	19.7	0.0
2023	13-Apr	70.9	44.2	85.1	2.7	10.7	10.5	0.6
2023	14-Apr	68.3	59.2	93.8	2.0	7.5	11.9	0.3
2023	15-Apr	78.7	57.1	74.2	2.8	6.7	23.8	0.0
2023	18-May	68.6	60.4	93.5	4.2	8.3	4.9	0.8
2023	19-May	71.2	58.6	83.8	3.6	7.5	12.0	0.0
2023	20-May	78.4	63.9	81.9	2.4	5.1	15.1	0.1
2023	15-Jun	80.8	66.0	83.0	4.0	5.9	16.3	0.0
2023	16-Jun	86.7	65.6	71.1	2.6	7.5	24.9	0.3
2023	17-Jun	88.9	66.7	64.0	1.7	5.1	26.9	0.0
2023	13-Jul	91.4	72.9	75.6	2.5	5.9	25.8	0.0
2023	14-Jul	92.0	73.1	78.6	2.2	4.3	23.4	0.0
2023	15-Jul	92.9	71.9	74.3	2.8	7.5	24.9	0.0
2023	17-Aug	86.0	64.9	72.8	1.5	5.9	25.3	0.0
2023	18-Aug	88.3	67.1	70.4	1.1	4.3	23.7	0.0
2023	19-Aug	86.5	64.2	75.3	1.6	5.1	24.5	0.0
2023	14-Sep	82.3	66.7	82.8	2.9	6.7	11.7	0.2
2023	15-Sep	77.5	65.3	82.0	2.6	5.1	10.3	0.1
2023	16-Sep	74.6	63.3	88.7	0.9	5.1	7.5	0.1
2023	20-Oct	74.2	53.5	76.0	2.9	9.1	10.9	0.1
2023	21-Oct	74.9	47.4	52.5	3.1	6.7	17.9	0.0
2023	22-Oct	77.9	50.9	50.5	1.9	5.9	17.8	0.0

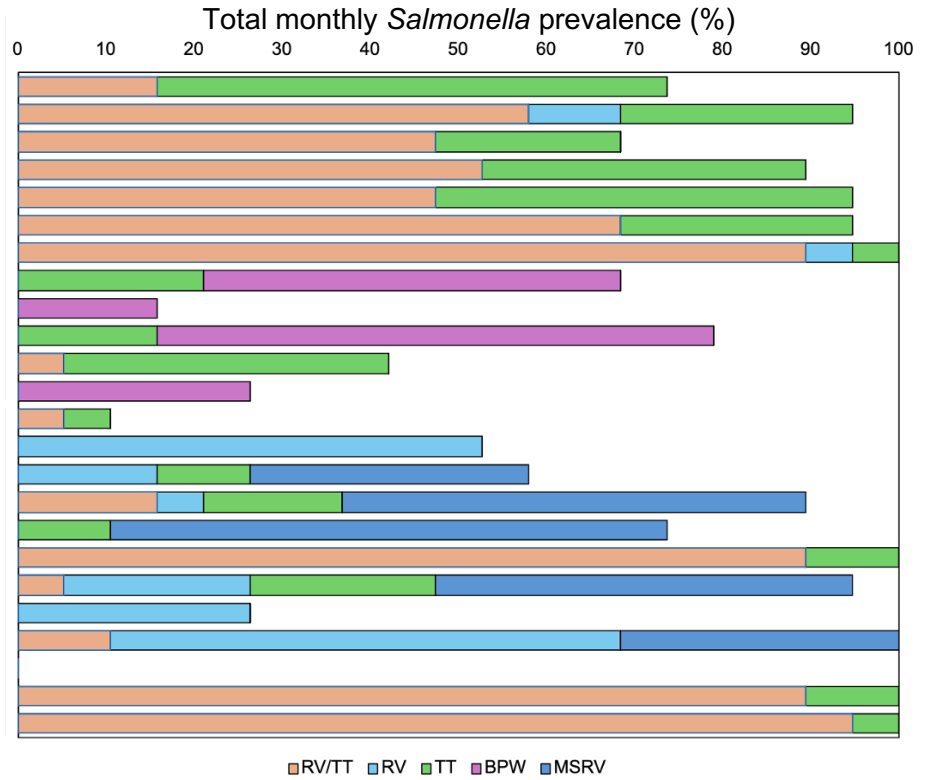
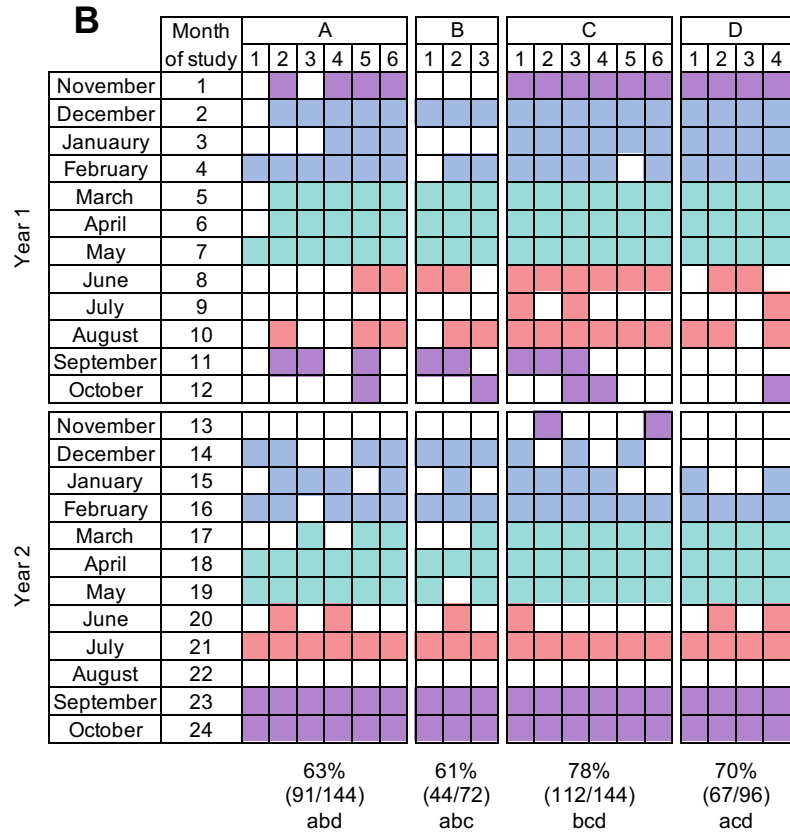
Supplemental Table S3.12 Weather metadata for day of sampling and two days prior from weather station 3 for Creeks C and D.

Year	Julian Day	Max Air Temperature(F)	Min Air Temperature(F)	Avg Relative Humidity(%)	Avg Wind Speed(mph)	Max Wind Speed (m/s)	Total Solar Radiation(MJ/m ²)	Total Rain(in)
2021	18-Nov	74.9	44.2	77.9	5.0	8.3	10.4	0.0
2021	19-Nov	61.9	33.7	48.2	3.0	6.7	13.8	0.0
2021	20-Nov	55.8	34.1	57.6	3.3	7.5	13.3	0.0
2021	16-Dec	63.2	44.8	83.0	2.0	4.3	5.0	0.0
2021	17-Dec	67.2	52.2	94.2	2.5	5.1	5.7	0.0
2021	18-Dec	65.8	56.6	99.3	4.7	6.7	2.0	0.8
2022	13-Jan	57.4	31.2	59.7	6.0	9.9	9.1	0.0
2022	14-Jan	57.1	36.2	65.7	4.1	6.7	12.0	0.0
2022	15-Jan	44.7	35.5	70.9	3.5	6.7	4.0	0.1
2022	17-Feb	72.3	52.2	79.4	8.0	17.1	6.1	0.0
2022	18-Feb	67.3	32.4	75.4	8.6	12.3	15.7	1.0
2022	19-Feb	57.9	27.9	52.6	6.6	9.9	17.8	0.0
2022	17-Mar	74.0	52.4	75.8	2.7	5.9	19.7	0.0
2022	18-Mar	58.2	50.5	91.1	3.3	20.3	2.4	0.7
2022	19-Mar	68.0	48.6	69.3	9.1	10.7	19.8	0.0
2022	14-Apr	77.0	54.8	65.4	6.7	9.1	16.9	0.0
2022	15-Apr	76.8	42.5	45.3	2.9	6.7	25.7	0.0
2022	16-Apr	64.2	53.1	89.6	3.0	4.3	7.3	0.6
2022	19-May	92.3	64.7	62.7	4.8	9.1	24.8	0.0
2022	20-May	89.3	66.9	64.6	4.7	9.1	26.9	0.0
2022	21-May	89.9	66.5	72.1	4.3	7.5	20.7	0.0
2022	16-Jun	98.6	73.3	65.2	1.7	4.3	24.1	0.0
2022	17-Jun	96.6	73.1	68.5	5.5	12.3	25.3	0.0
2022	18-Jun	92.9	68.7	58.8	5.1	9.9	28.2	0.0
2022	14-Jul	88.4	70.6	81.2	2.4	4.3	17.1	0.0
2022	15-Jul	92.0	69.1	72.0	2.3	4.3	24.3	0.0
2022	16-Jul	90.6	68.6	75.7	2.5	5.1	24.8	0.0
2022	18-Aug	83.4	69.5	80.6	0.9	4.3	13.8	0.0
2022	19-Aug	78.0	70.1	94.0	0.6	3.5	6.7	0.1
2022	20-Aug	86.1	71.4	84.9	1.0	7.5	16.0	0.2
2022	22-Sep	94.4	64.9	67.7	5.8	13.1	19.4	0.0
2022	23-Sep	79.9	53.3	55.9	1.4	5.1	22.5	0.0
2022	24-Sep	80.6	47.9	62.6	1.8	5.1	21.5	0.0
2022	20-Oct	63.5	31.8	52.4	2.5	6.7	17.3	0.0
2022	21-Oct	70.1	34.0	54.9	1.2	4.3	17.5	0.0
2022	22-Oct	74.6	36.7	60.3	1.1	3.5	17.0	0.0
2022	17-Nov	49.0	31.0	54.6	4.8	7.5	14.3	0.0
2022	18-Nov	55.1	27.5	49.9	3.2	5.9	13.8	0.0
2022	19-Nov	54.9	30.7	56.6	4.6	9.1	11.0	0.0
2022	12-Dec	56.3	43.0	95.4	1.3	3.5	4.3	0.0
2022	13-Dec	51.0	40.7	82.1	2.2	5.1	7.3	0.0
2022	14-Dec	45.2	41.1	90.5	2.5	5.1	0.9	0.6
2023	19-Jan	71.6	57.2	68.2	8.2	11.5	7.5	0.0
2023	20-Jan	59.4	43.2	45.0	8.9	10.7	13.2	0.0
2023	21-Jan	48.3	38.0	61.2	3.2	7.5	5.5	0.0
2023	16-Feb	73.0	44.8	82.5	3.6	7.5	9.2	0.0
2023	17-Feb	66.0	33.6	72.5	12.0	15.5	12.9	0.4
2023	18-Feb	57.3	27.3	51.5	3.0	5.9	18.3	0.0
2023	16-Mar	68.9	30.3	45.1	4.0	7.5	21.0	0.0
2023	17-Mar	60.2	47.4	84.1	5.9	12.3	3.5	0.7
2023	18-Mar	53.9	36.4	53.9	8.8	10.7	18.7	0.0
2023	13-Apr	72.4	44.7	87.6	1.7	8.3	10.6	0.6
2023	14-Apr	68.8	59.5	94.0	1.7	7.5	11.7	0.3
2023	15-Apr	78.3	56.6	74.6	3.5	8.3	23.4	0.0
2023	18-May	69.0	60.9	93.6	2.0	5.1	5.2	1.5
2023	19-May	72.6	58.8	83.3	1.6	4.3	12.2	0.0
2023	20-May	78.7	64.2	82.2	2.6	5.9	16.3	0.3
2023	15-Jun	80.7	65.2	85.0	4.9	9.1	16.8	0.0
2023	16-Jun	85.9	65.3	72.7	5.4	9.9	25.0	0.1
2023	17-Jun	88.2	65.9	66.0	3.3	7.5	25.5	0.0
2023	13-Jul	91.2	72.9	78.4	2.9	5.9	25.9	0.0
2023	14-Jul	90.9	72.4	80.7	2.6	5.9	22.6	0.0
2023	15-Jul	92.1	71.3	76.1	3.4	7.5	23.9	0.0
2023	17-Aug	85.2	64.6	74.3	2.9	6.7	23.9	0.0
2023	18-Aug	88.2	66.0	72.3	3.2	7.5	25.3	0.0
2023	19-Aug	87.1	65.3	73.7	1.2	4.3	23.7	0.0
2023	14-Sep	83.4	67.0	83.9	1.1	4.3	11.7	0.0
2023	15-Sep	78.6	65.1	83.9	1.4	4.3	11.2	0.1
2023	16-Sep	75.4	62.9	90.3	1.1	5.1	7.9	0.1
2023	20-Oct	73.3	52.7	77.7	5.5	9.9	10.3	0.1
2023	21-Oct	73.8	48.5	55.7	5.6	9.1	17.6	0.0
2023	22-Oct	76.6	53.1	52.4	4.8	9.1	17.4	0.0

A

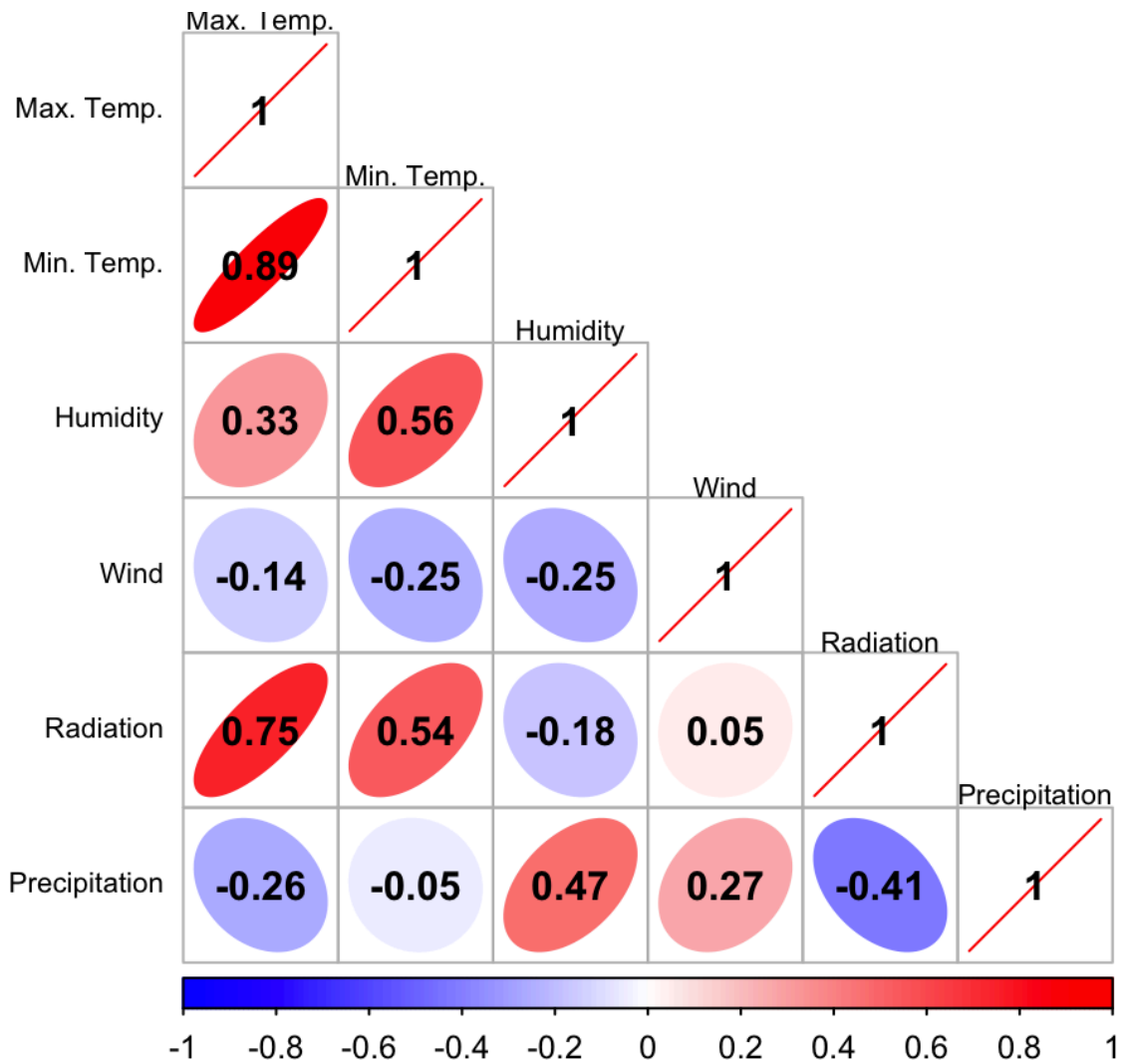
Modified-Moore
swab (10 L)



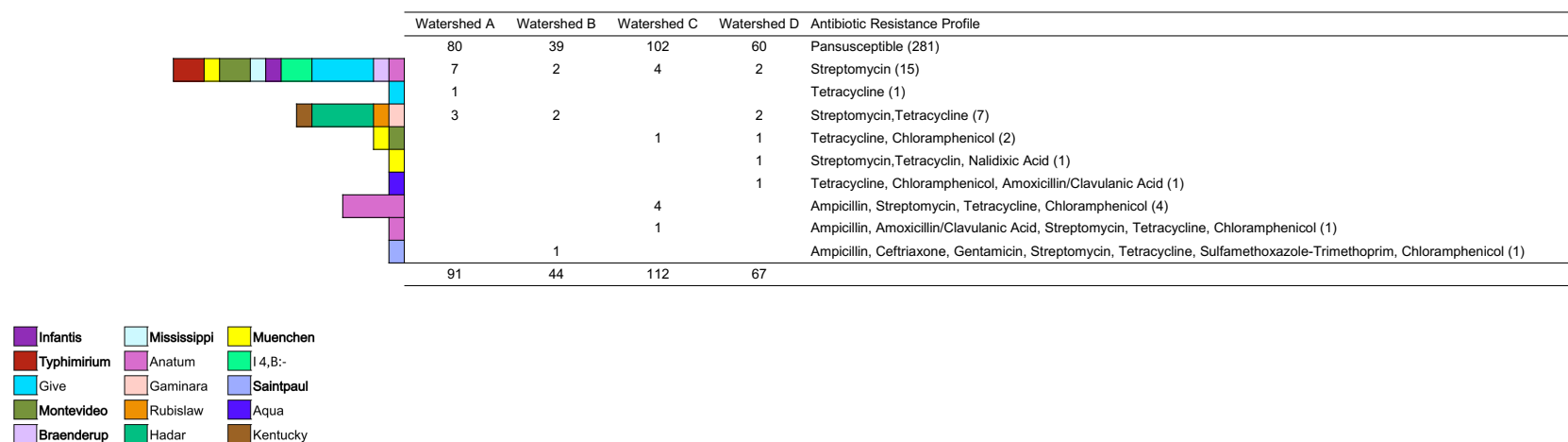


Supplemental Figure 3.1. Sample collection and prevalence of *Salmonella* in creeks.

A) Water samples were collected by pumping 10 liters through a modified Moore swab (MMS). The swab was returned to lab on ice and pre-enriched in buffered peptone water (BPW). Samples were transferred to selective enrichment media in parallel, including Rappaport-Vassiliadis (RV), Tetrathionate (TT), and modified semisolid RV (MSRV). Each enrichment was then plated onto an indicator plate, xylose lysine tergitol-4 (XLT-4) for confirmation. B) Prevalence is shown with colored squares reflecting positive samples and white squares are negative with overall sample number for each creeks shown at the bottom. The color of the square reflects the season of the sample collection, with observed differences in prevalence between them: fall (purple), winter (light blue), spring (light green), summer (light red) (Chi-squared test, $p < 0.001$). Additionally, *Salmonella* recovery varied by creek, with relationships denoted below prevalence values (Chi-squared test, $p < 0.05$). Isolated colonies were saved from XLT-4 plates for all positive enrichments, including from BPW when RV and TT cultures were negative. MSRV was included in the protocol at the beginning of Year 2 due to sustained low prevalence in RV and TT starting in June of Year 1, however, isolates only from this media were collected in five out of 12 months in Year 2.



Supplemental Figure 3.2. Association between different weather variables. Pearson correlation coefficient was calculated for each pairwise combination of weather variables recorded in this study. The resulting coefficients are displayed within each respective box of the matrix, with the size of the circle indicating the strength and direction of association and the color denoting if it was positive (red) or negative (blue).



Supplemental Figure 3.3. Assessment of antibiotic-resistant *Salmonella* isolates. Kirby-Bauer disk diffusion assays were used to assess resistance to ampicillin, amoxicillin-clavulanic acid, ceftriaxone, gentamicin, streptomycin, tetracycline, ciprofloxacin, sulfamethoxazole-trimethoprim, and chloramphenicol. The classification of resistance was determined by CLSI, and an intermediate result was considered as susceptible. Multidrug resistance was defined by resistance to three or more classes of antibiotics. Serovar identification was completed using SeqSero from whole genome sequencing data. Bolded serovars are of highest human clinical importance as denoted by their presence in the top 15 serovars listed on the BEAM Dashboard.

System	Season	Site	Month	Agbeni	Agona	Alabama/Bareilly III	Anatum	Aqua/Inverness	Braenderup	Brazil I	Cubana	Derby III	Dublin	Enteriditis	Gaminara	Give I	Hadar	Hartford	Infantis	Java	Johannesburg/Urbana	Kentucky I	Kiambu	Luciana	Mbandaka	Mississippi II	Montevideo I	Montevideo II	Muenchen I	Muenchen II	Newport II	Oranienburg	Poona	Rissen	Rubislaw	Saintpaul I	Saintpaul II	Senftenberg I	Senftenberg II	Thompson	Typhimurium	Uganda	Untypeable	# of serovars per sample		
A	Fall	2	Nov-21																																								6			
	Fall	2	Sep-22																																									2		
	Fall	3	Sep-22																																										4	
	Fall	4	Nov-21																																										1	
	Fall	5	Nov-21																																										4	
	Fall	5	Sep-22																																										2	
	Fall	5	Oct-22																																										3	
	Fall	6	Nov-21																																										2	
	Winter	1	Feb-22																																											1
	Winter	2	Feb-22																																											1
	Winter	3	Dec-21																																											2
	Winter	3	Feb-22																																											3
	Winter	4	Dec-21																																											3
	Winter	4	Jan-22																																											4
	Winter	4	Feb-22																																											4
	Winter	5	Dec-21																																											3
	Winter	5	Jan-22																																											4
	Winter	5	Feb-22																																											9
	Winter	6	Dec-21																																											4
	Winter	6	Jan-22																																											4
Winter	6	Feb-22																																											6	
A	Spring	1	May-22																																											3
	Spring	2	Mar-22																																										3	
	Spring	2	Apr-22																																											1
	Spring	2	May-22																																											2
	Spring	3	Mar-22																																											5
	Spring	3	Apr-22																																											1
	Spring	3	May-22																																										4	
	Spring	4	Mar-22																																											6
	Spring	4	Apr-22																																											3
	Spring	4	May-22																																											3
	Spring	5	Mar-22																																											

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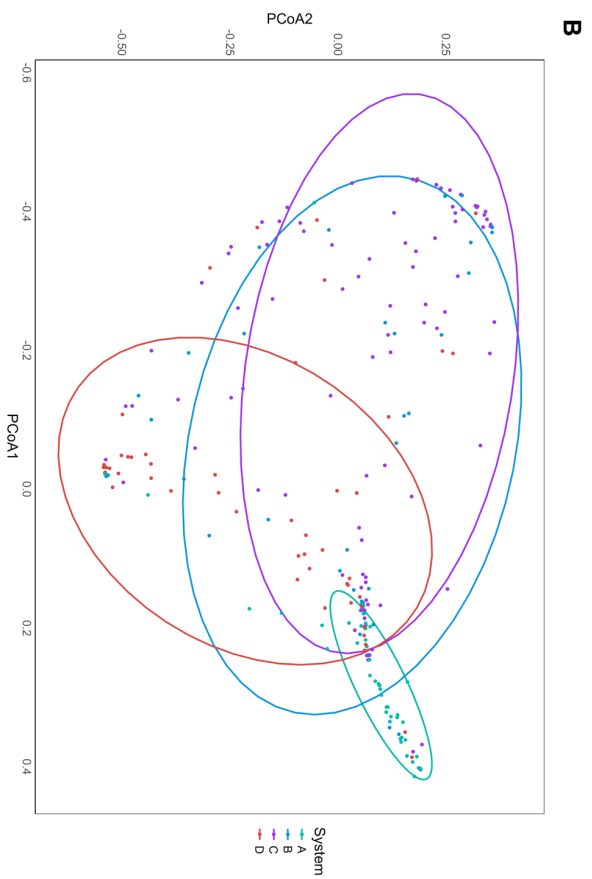
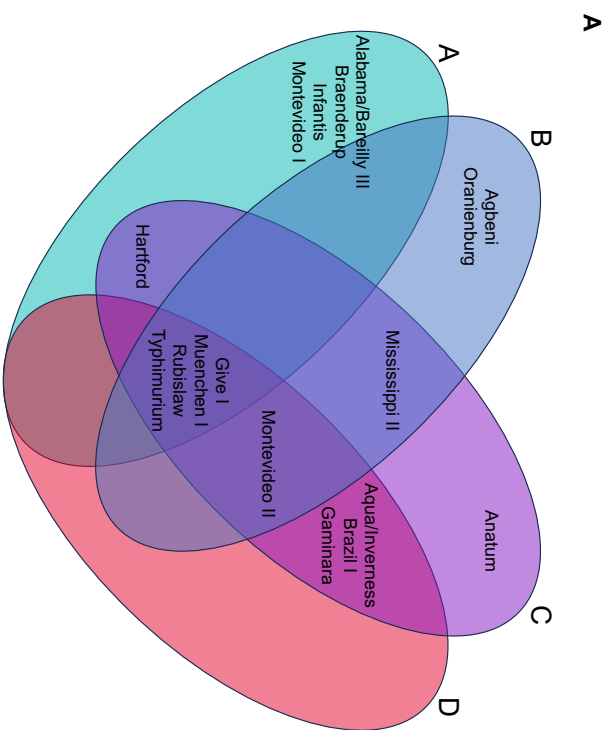
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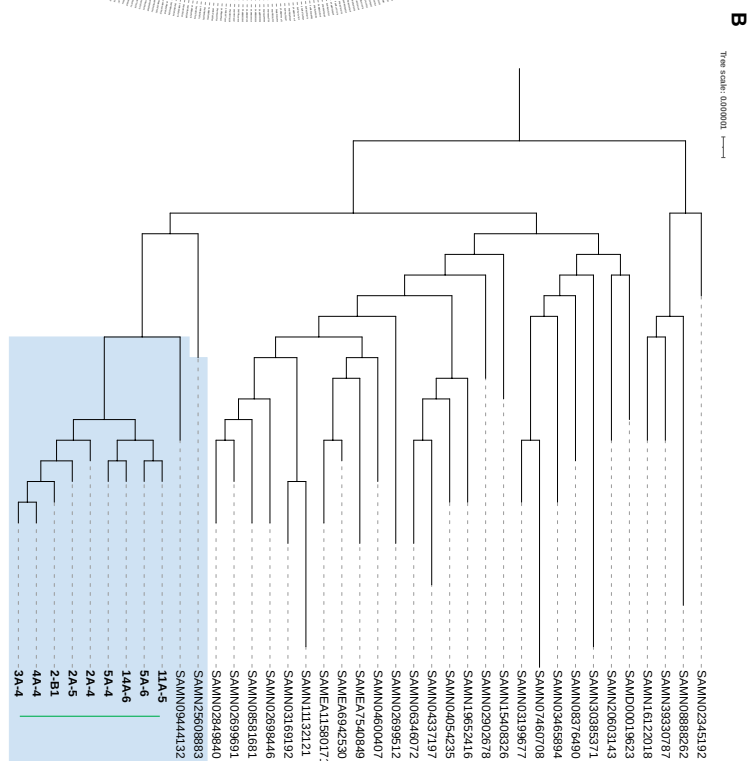
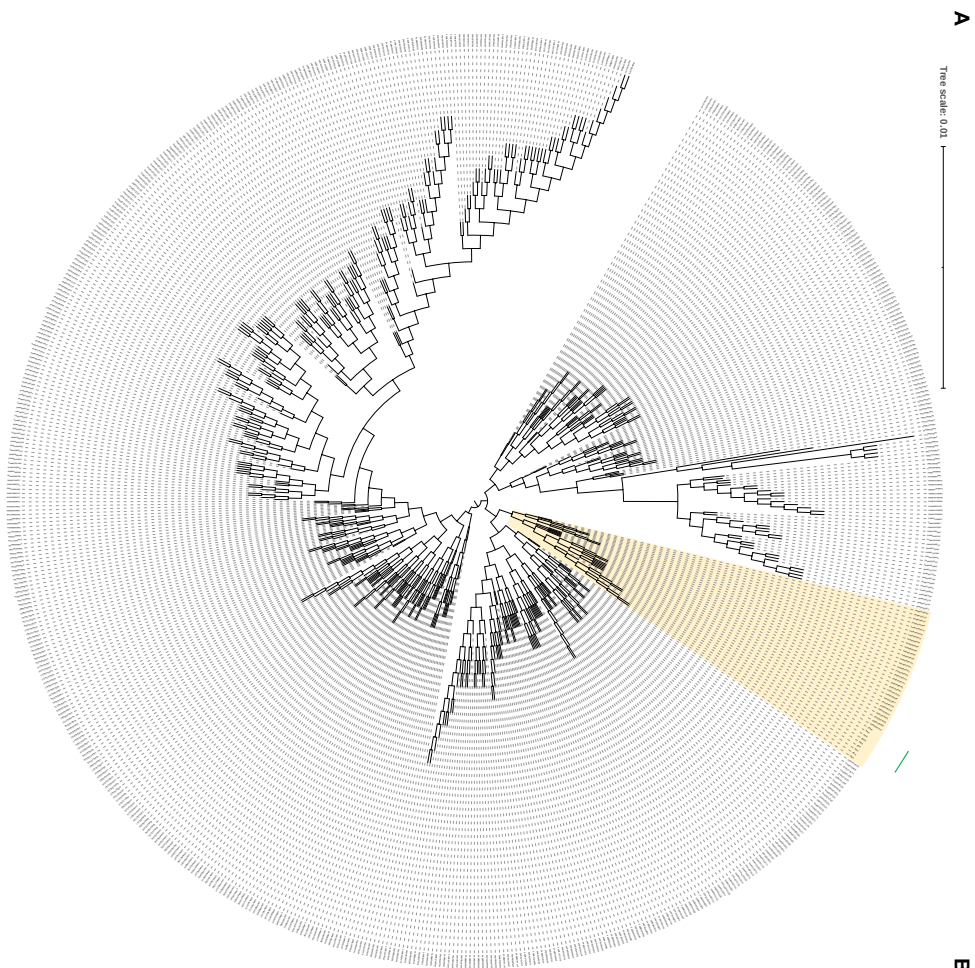
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Relative serovar frequency	
0	
0 - 0.10	
0.10 - 0.25	
0.25 - 0.50	
0.50 - 0.75	
0.75 - 0.90	
0.90 - 1	

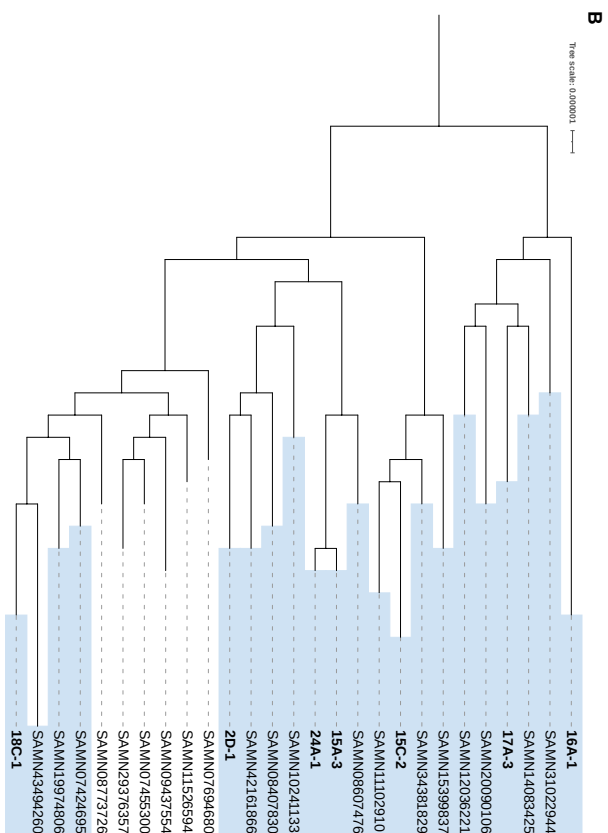
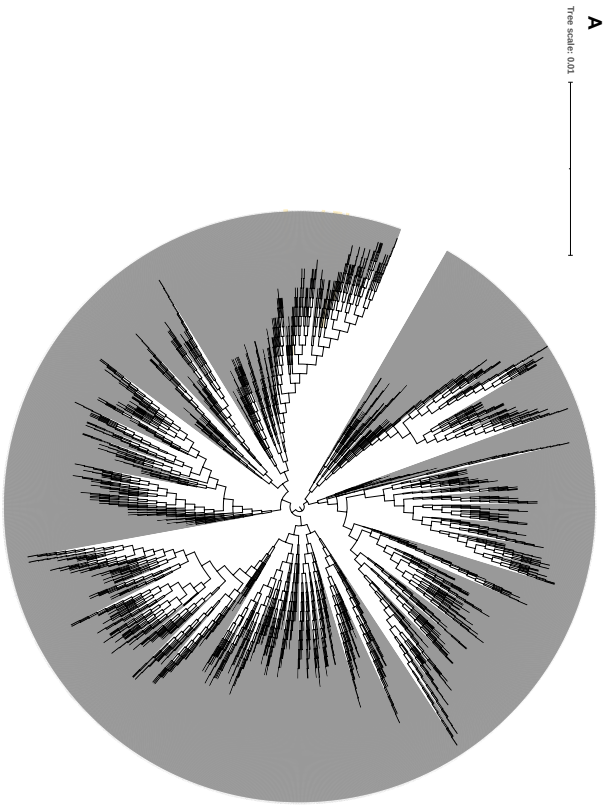
Supplemental Figure 3.4. *Salmonella*-positive water samples contain high serovar diversity. CRISPR-SeroSeq was used to determine the relative abundance of *Salmonella* serovars within each sample, where results from all positive enrichments (BPW, RV, TT) were normalized and combined. Samples are arranged according to the system, season, and site they were collected in, with fall in the first year of the study split between 2021 and 2022. The individual serovars are shown on the top, number of serovars identified per sample included on the right, and the heatmap shows the relative serovar abundance in each sample according to the key. Labels containing two serovars reflects that it is not possible to differentiate based on deep serotyping alone.



Supplemental Figure 3.5. Distribution of serovars and serovar populations across the creeks. A) Venn diagram displays the overlap of serovars found at least five times in each respective creek. Serovars located only within one ellipse were only found five or more times in that creek but could have been identified four or less times in any other creek. B) Principal component analysis of the Jaccard distance of samples within creeks is shown with ellipses containing 75% of each creek's samples. A larger ellipse indicates a more variable population, while a smaller ellipse indicates higher consistency.



Supplemental Figure 3.6. Preliminary serovar *Infantis* phylogenies including nine isolates from this study and publicly available genomes from NCBI. Study isolates are indicated with a green bar, with the label format of month-creek-site, and NCBI isolates are listed by sample ID. The phylogenies are rooted at the midpoint. A) Phylogeny includes one representative isolate from each available SNP cluster belonging to serovar *Infantis* on NCBI (n = 650). Yellow highlight indicates clade selection for subsequent phylogeny. B) Phylogeny includes one representative isolate from the most closely related SNP clusters (n = 30). Blue highlight indicates clade selection for subsequent phylogeny.



Supplemental Figure 3.7. Preliminary serovar Typhimurium phylogenies including seven isolates from this study and publicly available genomes from NCBI. Study isolates are bolded, with the label format of month-creek-site, and NCBI isolates are listed by sample ID. The phylogenies are rooted at the midpoint. A) Phylogeny includes one representative isolate from each available SNP cluster belonging to serovar Typhimurium on NCBI (n = 3,057). Yellow highlight indicates clade selection for subsequent phylogeny. B) Phylogeny includes one representative isolate from the most closely related SNP clusters (n = 20). Blue highlight indicates clade selection for subsequent phylogeny.